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Vegetation succession following the clearance of Rhododendron in the Vale of Ffestiniog.

Hockley, T G.

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VEGETATION SUCCESSION FOLLOWING THE CLEARANCE OF RHODODENDRON IN THE VALE OF FFESTINIOG

A dissertation submitted in partial fulfilment of the requirements for the degree of <u>Magister in Scienta</u> in Environmental Forestry.

School of Agricultural and Forest Sciences
University College of North Wales
Bangor
Gwynedd
North Wales

T.G. HOCKLEY

October 1989

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ABSTRACT

This study analyses vegetation succession following the clearance of <u>Rhododendron</u> in a sessile oakwood in North Wales. The <u>Rhododendron</u> has been cleared since 1984/5 and prior to clearance dominated the understorey.

The succession was analysed in relation to the number of years since clearance; method of clearance; density of the former <u>Rhododendron</u> cover, and vegetation adjoining the site. An experiment was undertaken in the University's nursery to assess the presence and viability of seed within the seedbank, and the ability of the soil to support germination. Over 50% of the samples were dominated by bryophytes (60%), with birch, bramble and <u>Rhododendron</u> seedlings each present at less than 10%.

There was a significant difference between the mean percentage crown cover of each TWINSPAN group; those groups with more light demanding species had a lower mean percentage crown cover than groups dominated by more shade tolerant species. Only bramble and birch germinated in soil collected from the two most recent years of clearance. Their presence is probably due to seed dispersed prior to clearance, and there is therefore no viable seed in the seedbank. Colonisation must therefore rely on immigration of seed.

There was no significant difference between the ability of soil recently infested by <u>Rhododendron</u>, or the ability of peat, to support germination of <u>Betula pendula</u> or <u>Festuca</u> ovina.

There is a significant difference between the frequency and abundance of vegetation and the year of clearance. With the

exception of one plot, frequency and abundance increased with the number of years since clearance.

Rhododendron seedlings are associated with a bryophyte cover. The site contains substantial Rhododendron coppice and seedlings, and continued eradication and containment programmes will be needed.

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TABLE OF CONTENTS

		PAGE	NUMBER
Abstracknown Table List	wledgements of contents of figures of tables of maps	vi	i ii iv v vii .ii .ii
1.0	INTRODUCTION		1
2.0	OBJECTIVES		1
	1.1 Nomenclature 3.2 Introduction to Great Brit 3.3 Seed production and dispersal and germination of the stablishment 3.5 Competitive ability followestablishment 3.5.1 Shade tolerance 3.5.2 Resistance to herbivore 3.5.3 Allelopathy 3.5.4 Mycorrhiza 3.5.5 Freedom from pests and 3.6 Vegetative spread and lor 3.7 Climatic conditions 3.8 Edaphic conditions 3.9 Resistance to herbicide 3.10 Associated fauna	ersal on owing es diseases	2 2 4 5 9 10 10 11 12 13 14 14 15 16
4 4 4 4	SITE DESCRIPTION 4.1 Introduction 4.2 Rhododendron cover 4.3 Tree cover 4.4 Species composition 4.5 Adjacent vegetation 4.6 Geology		18 21 21 23 23 25

5.0	EXPER	IMENTAL DESIGN AND COLLECTION OF DATA	
	5.2	Germination trials Collection of vegetation succession data	26 28
	5.2.1	Tree density and <u>Rhododendron</u> cover	28
	5.2.2 5.2.3	Vegetation survey Adjacent vegetation	29 31
6.0	ANALY	SIS OF DATA	
	6.1 I 6.2 I	Data analysis techniques Data analysis	32 39
		Vegetation associations Crown cover and vegetation associations	39 46
		Vegetation succession and	50
	6.2.4 6.2.5	Species composition within plots Succession of <u>Rhododendron</u> and oak	53 62
	6.3 (Germination trials	64
	6.3.1	Germination of seed within the seedbank	64
	6.3.2	Test sowings	67
7.0	CONCLU	JSIONS AND RECOMMENDATIONS	
	7.2 \ 7.3 \ 7.4 \ 7.5 M	Vegetation associations Vegetation and light intensity Sermination trials Vegetation succession Vanagement implications Verther research	70 70 70 71 73 73
	REFERE	ENCES	74

LIST OF FIGURES

FI	GURES	PAGE	NUMBER
1	Tree species composition	Ž	24
2	Categories of Rhododendron density	3	30
3	TWINSPAN sample classification	3	36
4	TWINSPAN species classification	3	37
5	DECORANA sample ordination coded to accord with TWINSPAN groups	4	17
6	DECORANA species ordination	4	18
7	Mean and standard deviation of percentage crown cover for each TWINSPAN group	4	19
8	DECORANA sample ordination showing year of clearance only	5	51
9	Cumulative frequency and Domin values for each year	5	52
10	Species composition in plot 5	5	4
11	Species composition in plot 4	5	6
12	Species composition in plot 3	5	7
13	Species composition in plot 2	5	8
L4	Species composition in plot 1	5	9
15	Summary of species composition between plots	6	0
16	Rhododendron seedling establishment	6	2

LIST OF TABLES

T	ABLE	PAGE NUMBER
1	Mean yield of $\underline{\text{Festuca ovina}}$ seedlings with and without the presence of $\underline{\text{R.ponticum}}$	11 -
2	Site stand composition (dbh classes)	23
3	The Domin scale	29
4	TWINSPAN two way table	38
5	Distribution of quadrats between TWINSPAN groups	43
6	Summary of vegetation associations	44
7	One-way analysis of variance for the relationship between percentage mean crown cover and TWINSPAN groups	50
8	Analysis of variance for the regression between vegetation succession and year of clearance	52
9	Summary of species composition for each plo	ot 61
10	Species germinating in seedtrays 1 - 10	64
11	Germination rates in seedtrays la - 10a & controls 1 - 5	67
12	One-way analysis of variance for test sowings	68

LIST OF MAPS

MA	APS	PAGE NUMBER
1	Site location	19
2	Location and year of clearance of each plot	20
3	Adjacent vegetation and areas of each	22

APPENDICES

AP	PENDIX PAG	E NUMBER
I	Parasitic flora and fauna associated with <u>R.ponticum</u>	79
II	Bird species abundance following the removal of Rhododendron at Coombes Valley	80
III	Assessment of the quantity of seed per seedtray	81
IV	Species recorded at plot 1	82
V	Species recorded at plot 2	83
VI	Species recorded at plot 3	84
VII	Species recorded at site 4	85
VIII	Species recorded at site 5	86
IX	Species list and abbreviated names	87
Х	Scheffey's test	90

1.0 INTRODUCTION

Rhododendron has colonised many areas to the exclusion of native vegetation. The freely draining, acid soils of Western Britain are particularly favourable to the plant, and in the Snowdonia National Park Rhododendron is found in an area of approximately 34km² (Snowdonia National Park, 1987). No statutory powers of control exist, and the plant has joined the grey squirrel, Japanese knotweed and New Zealand willowherb as undesirable aliens (Simons, 1988). The complete eradication of Rhododendron is dependent on substantial and sustained expenditure; containment is therefore seen as the only viable objective, with clearance undertaken in areas of prime importance only. Attempts to assess current annual expenditure on clearance and containment are hampered by the predominant reliance on voluntary labour, the availability of which is limited. An indication of expenditure can however be derived from the Forestry Commission budget of £6500 for clearance in 1989/90 in two forest districts of North Wales only, and the N.C.C budget for 1989/90 of approximately £3700, most of which will be spent in the Maentwrog area.

2.0 OBJECTIVES

The primary objective of the study was to analyse succession following the clearance of <u>Rhododendron</u>, thereby assessing the vegetation associations present in a given year of clearance. In addition the succession was related to the light intensity; method of clearance; density of the former <u>Rhododendron</u> cover and of the sessile oakwood; and the vegetation adjoining the site. The vegetation developing is

dependent on the ability of the soil to support vegetation and available seed sources. These seed sources may either be seed alighting in the area for the first time, or seed that is stored in the seedbank. Germination trials were undertaken to assess the presence of viable seed in the seedbank, and the ability of the soil to support germination and growth.

3.0 LITERATURE REVIEW

3.1 Nomenclature

The full Latin name <u>Rhododendron ponticum</u>, with a subspecies qualification, is used where appropriate. The British feral form is referred to simply as <u>Rhododendron</u>.

3.2 <u>Introduction to Great Britain</u>

The genus <u>Rhododendron</u>, along with <u>Erica</u>, <u>Calluna</u> and <u>Vaccinium</u>, belong to the family <u>Ericaceae</u> (Clapham et al, 1981). The genus consists of more than 600 species concentrated in South East Asia, with a few representatives in America, Asia Minor and Iberia (Cross, 1975). <u>Rhododendron</u> was first introduced to Britain in 1763 with the first plant established at Kew in 1765 (Bean, 1976). These early introductions were dispatched from Gibraltar, and are likely to have been of the Iberian form from either Spain or Portugal (Shaw, 1984).

According to Flora Europea (Tutin et al, 1972), the Iberian form is of the sub specific status <u>baeticum</u>, the other sub species, <u>ponticum</u>, being naturally distributed in the Balkan Peninsular (Bulgaria, Turkey and the USSR) and Lebanon, a geographical separation of nearly 3000Km (Shaw, 1984). The

distribution of <u>R.ponticum</u> is thought to be the result of Quaternary glaciation. Shaw (1984) quotes Jesson et al (1959), in referring to the fossil remains that have been recovered in various places in Europe, and that indicate a once continuous distribution. Whilst <u>R.ponticum</u> has been found in two sites of interglacial age in Ireland, no such finds have yet been made in Britain (Shaw, 1984).

Since its introduction <u>Rhododendron</u> has been widely planted for game cover and ornamental purposes. Before modern propagation techniques were developed, <u>R.ponticum</u> was commonly used as a rootstock for grafting ornamental varieties (Brown, 1953a); when planted out the rootstock often developed shoots, which in time overwhelmed the less robust scion.

Cross (1975) considers plants in Britain to predominantly resemble R.ponticum, whilst Shaw (1984) believes feral Rhododendrons in North Wales most closely resemble R.baeticum. A number of separate introductions have probably led to both sub species and sub specific hybrids being present (Shaw 1984).

Rhododendron is now widely naturalised in suitable areas, such as the free draining acid soils of Western Britain, and aerial photography and population studies in Snowdonia show that the species has become increasingly widely distributed in the last 15 to 20 years (Shaw, 1984). Rhododendron is estimated to be present in approximately 3,407ha or 1.6% of the National Park, with over 49.8% of conifer woodland, and 14.7% of broadleaved woodland infested (Snowdonia National Park, 1987).

3.3 Seed production and dispersal

In common with other ericaceous species Rhododendron seeds are among the smallest in the plant kingdom, the average weight of seed being assessed by Cross (1981) as micrograms, and by Shaw (1984) as 50 micrograms, or million /kg. Under British conditions Rhododendron flowers annually, with a single flower head (a raceme) containing 3000 to 7000 seeds. Given that Rhododendron is monoecious, and that there are several hundred flowers on a bush, the reproductive potential is high. Flowers on the raceme also open sequentially from top to bottom, reducing loss to high winds or frosts (Shaw, 1984). Bushes originating directly from seed do not usually flower until 10 or 12 years old, although coppice shoots may reproduce earlier. The flower buds are terminal, continued growth being through the activation of three or four lateral buds, resulting in an actively flowering bush quickly becoming bushy (Shaw, 1984). Nearly one year elapses between flowering and the release of seeds, the capsules opening in dry weather in the late winter and spring following flowering (Shaw, 1984). The seeds resemble flattened disks with frills of hairs at each end (Cross, 1975) which are estimated to have only a minor aerodynamic function (Brown, 1953a). Dispersal is mainly by wind, and adopting a sinking velocity of the seed in still air of approximately 1m/sec (Brown, 1953a), the theoretical dispersal distance in strong winds is up to 100 metres (Shaw, 1984). Given the turbulent air conditions found in mountainous regions, Shaw (1984) believes seed in Snowdonia can be carried far greater distances. Dispersal within woodlands is less effective; Brown (1953b) recorded seed

dispersal of only 16.6 metres in open woodland.

3.4 Early life and germination

The ultimate success of dispersal is dependent on the seed alighting on a site suitable for germination, and receiving the conditions needed for subsequent growth. The probability of these criteria being met are extremely low, a problem countered by prolific seed production and tolerance of shade.

Rhododendron seeds will germinate in 5 - 6 days under favourable conditions, although unlike many other weed species the seeds deteriorate quickly. Under field conditions all seeds became non viable after 160 days (Cross, 1973). Even under ideal (artificial) conditions of low temperature and desiccation the seed does not store well, suffering approximately a 50% loss of viability in one year (Shaw, 1984).

Germination will occur on many substrates (Cross, 1975). A germination rate of 80% was recorded on a semi natural medium (horticultural peat, pH 4.0), of 60% on a mor humus of pH 4.4 from an oak / holly wood, and approximately 40% on a brown earth of pH 5.5.

Brown (1953a) recorded germination in less than 1% daylight, although fungal attack prevented accurate estimates of germination rates. Cross (1973) found that the shade cast by Rhododendron thickets (2% total light) reduced the rate and quantity of germination when compared with shade cast by an oak / holly canopy (9% total daylight). In one experiment the rate of germination after 49 days under a Rhododendron canopy was 17%, compared to 75% under an oak / holly canopy

(Cross, 1973). Germination only started to decline in levels of less than 5% full daylight (Cross, 1975). Therefore providing the seeds are not buried, the light in most woodlands is unlikely to restrict germination unless the canopy is very dense.

The effect of temperature on germination is not known (Cross, 1975), although there is no evidence that this is a limiting factor in Britain (Shaw, 1984). Moisture is more critical, short periods of desiccation (perhaps even measured in hours) can be lethal (Shaw, 1984).

The seedling stage is generally considered to be the most hazardous stage in the life of a plant (Cross, 1981), and this transition from dependence on seed reserves to independent assimilates is particularly critical for Rhododendron, which has little competitive ability in the seedling stage, and is readily smothered by low growing vegetation and drifting litter (Cross, 1981). The term "safe site", originally used by Harper et al (1965) is used to describe the conditions in which Rhododendron seeds can germinate and the seedlings develop. A safe site is essential as the cotyledon stage seedling of Rhododendron is only 2 - 5 mm across with a rudimentary root system, and even at the end of one growing season the seedlings may only consist of the cotyledons and one true leaf (Shaw, 1984). During this stage the plant is therefore very vulnerable to adverse conditions such as desiccation, frost heave, trampling, or smothering by the dead parts of other plants. It is only when the plants are 3 - 5 years old, and 5 - 10 cm in height, that their vulnerability declines.

The most favourable safe sites are thin bryophyte carpets,

such as those formed by the pleurocarpus mosses (having lateral fructifications), such as <u>Plagiothecium undulatum</u> (Shaw, 1984). These maintain a relatively humid environment, and act as a protective cushion which lessens the physical impact of heavy drips from the canopy that can uproot exposed seedlings, or bury them with soil thrown up from the drip (Cross, 1981). Taller communities of acrocarpus mosses (having terminal fructifications), such as <u>Polytrichum formosum</u> do not provide such favourable sites as they tend to provide conditions similar to that of grass or a grass / herb sward (Shaw, 1984).

Inherently moist patches of bare mineral soil and humus can also provide safe sites (Cross, 1981). If not moist the dark surfaces of these substrates are subject to extremes of temperature and desiccation; they are often unstable and receive heavy drips from the canopy which can physically destroy the young seedlings. The tightly packed shoots of most Sphagnum spp provide a suitable habitat, although seedlings experience difficulty if they cannot quickly get their roots into the soil (Shaw, 1984).

A ground cover of litter or vegetation, like a thick bryophyte cover, also restricts establishment. This may be due to a combination of burial, drought, and the instability of the substrate. Even if germination does occur, the roots may fail to reach the soil and the seedlings subsequently die during a dry spell (Cross, 1981).

The importance of this bryophyte layer is demonstrated by a Rhododendron clearance in the Killarney oakwoods of Ireland, where for 5 years no Rhododendron seedlings appeared on the peaty humus exposed to the full sun until a 2 mm layer of

mosses became established (Cross, 1981). Cross (1981) also studied the relationship between the depth of bryophyte cover and the number of R.ponticum seedlings and found that one year old plants occurred only in bryophyte carpets of 1cm depth or less, and that 2 - 3 year old plants were concentrated in carpets of a depth of less than 2.5 cm.

Factors favouring the development of bryophytes will consequently favour Rhododendron. Shaw (1984) divides these factors into two categories: static and active. Static factors such as high rainfall and humidity, steep slopes, rocks and shelter reflect local climate and topography. Active factors such as grazing, trampling, burning and other factors of disturbance are - in contrast to static factors controllable. The role of disturbance is emphasised by the observations of Cross (1981) that the establishment of seedlings in the Killarney oakwoods was related to the distribution of loose boulders, rock outcrops and breaks in slope. The distribution also showed a concentration on slopes of between 10° and 50°, with a decrease in numbers of seedlings with increase in angle of slope. This is partly attributed to the tendency of seedlings to root on humus formed from bryophytes. The humus tends to be thicker at the base of slopes and therefore less subject to drought than soil at the top of slopes.

Cross (1981) describes an experiment in the Killarney oakwoods where areas were enclosed to reduce disturbance. It was observed that under a relatively thin canopy a thick herbaceous ground cover developed that had a smoothing effect, burying many of the surface irregularities and reducing the movement of air close to the ground. This in

turn led to litter accumulating either in the vegetation itself or in the pockets formed for example between vegetation and the base of rock outcrops. Species such as Luzula sylvatica which spread vegetatively into thick bryophyte mats or rock outcrops and loose boulders contribute still further to this smoothing effect. The elimination of grazing therefore reduces the number of safe sites available.

3.5 Competitive ability following establishment

3.5.1 Shade tolerance

Prolific seed production and shade tolerance are the methods by which Rhododendron overcomes the difficulty of finding safe sites for germination and early growth. established, Rhododendron is remarkably shade tolerant, with a physiological compensation point of less than 2% full daylight, and with only 5 - 10% of full daylight required for satisfactory growth (Shaw, 1984). The seedlings have a low metabolic rate; they have the ability to modify specific leaf area in response to prevailing light levels (Shaw, 1984). Cross (1975) found an 89% increase in dry weight matter over the winter months for 5 year old Rhododendron seedlings growing in an oakwood. Although there was no change in size this reflects the advantage of being evergreen and possessing the ability to make use of improved light conditions when deciduous woodland is not in leaf. Dormancy of buds is broken by increased day length with shoot extension usually beginning in early May (Cross, 1975).

3.5.2 Resistance to herbivory

Cross (1975) quotes Forsyth (1954) who found that R.ponticum contains andromedo toxin, which is highly poisonous if ingested by mammals. Grazing animals therefore avoid browsing the plant, giving it a significant advantage over other woody plants, including spiny ones such as Crataegus monogyna and Ilex aquifolium (Shaw, 1984). Cross (1981) found no evidence of seedlings in the Killarney oakwoods being damaged or killed by animals, although some seedlings may have been eaten by slugs which are known to eat newly germinated seedlings (Cross, 1973).

Rotherham (1988) investigated the free phenol content of a range of R.ponticum tissues. The concentrations were generally higher in leaves than stems or roots. The highest concentrations were in new leaf (57.0% dry weight), new stem (22.0% dry weight), and adventitious roots (17.7% dry weight). High concentrations of these free phenols may offer resistance to attacks from both vertebrate and invertebrate herbivores (Rotherham, 1988).

3.5.3 Allelopathy

Rotherham (1988) conducted experiments into the allelopathic properties of <u>Rhododendron</u>. Under controlled laboratory conditions <u>R.ponticum</u> was able to suppress the growth of <u>Festuca ovina</u> seedlings, resulting in a decrease in dry weight and almost total inhibition of root growth. This suppression was independent of competition for water, nutrients, space, light or mycorrhizal infection, and applied to <u>Festuca ovina</u> grown either in soil with living <u>Rhododendron</u> plants, or in soil which has recently supported

Rhododendron (Table 1). This effect may be a result of toxic organic acids released from Rhododendron roots into the soil, these organic acids being either simple phenols or short chain aliphatics. Cross (1975) states that, in common with other ericaceous plants, Rhododendron has a deleterious effect on the soil mobilising cations either directly or indirectly by the production of polyphenols.

TABLE 1: MEAN YIELDS OF FESTUCA OVINA SEEDLINGS (MG DRY WEIGHT) AFTER 6 WEEKS GROWTH WITH AND WITHOUT THE PRESENCE OF RHODODENDRON (ROTHERHAM 1988)

Experimental conditions	<u>M</u>	<u>ean yie</u>	lds (mg)_
<pre>(m) = Mycorrhiza present (nm) = Mycorrhiza absent</pre>	Added <u>nu</u> soluti			tilled zed water
(IIII) II GOLLIII Za abbeile	Root	<u>Shoot</u>	Root	Shoot
(i) R.ponticum present (m) (ii) R.ponticum present (nm) (iii) R.ponticum harvested	0.00	0.01	0.00 0.0001	0.01 0.03
prior to treatment (m) (iv) R.ponticum harvested	0.00	0.03	0.00	0.01
prior to treatment (nm) (v) Control (endophyte only (vi) Control (no plant		0.0 4.68	0.00 0.52	0.02 6.15
or fungus)	0.66	5.93	0.47	4.59

3.5.4 Mycorrhiza

Dramatic increases (184% dry matter production) in Rhododendron growth were achieved by mycorrhizal infection in soils of low nutrient status (Rotherham, 1988). Like other ericaceous plants R.ponticum has endotrophic mycorrhiza (Cross, 1975), forming such an extensive and efficient network that neighbouring plants are starved by the soaking up of nutrients around their roots (Simons, 1988). This ability to compete more effectively than competitors for restricted nutrient resources is

particularly important in the acidic soils that are typical of the areas invaded by <u>Rhododendron</u> (Rotherham, 1988).

The view of a competitive advantage given to Rhododendron by mycorrhizal infections is not universally held. Tester (1988) points out that mycorrhizas are found in most plants, including those commonly competing with Rhododendron such as ivy, oak and heather. Indeed the advantage afforded to mycorrhizal plants over plants not forming such associations is itself not absolute, as many weedy species invading disturbed land such as stinging nettles, goosefoot and dock are rarely affected by mycorrhizal fungi. Tester (1988) concludes that the ability of Rhododendron to compete so effectively with native flora is not even partly due to mycorrhizal associations.

3.5.5 Freedom from pests and diseases

Elton (1958) cites <u>Rhododendron</u> as an example of an invasive plant that may be in the process of being brought under control by the later introduction of animals and plants that feed on it. There is little evidence that this has occurred (Shaw, 1984), and due to its presence in Britain between the ice ages, Rotherham (1988) states that the species is already adapted to the pests and diseases found in Britain. Cross (1975) lists the parasitic flora and fauna associated with <u>Rhododendron</u> (Appendix I), although none of these have yet had any serious debilitating or control effect. Given the distribution and size of the genus it is likely that there are predators that could have a great impact on the species (Shaw, 1984). Any natural control mechanism is however likely to conflict with horticultural interests.

3.6 <u>Vegetative spread and longevity</u>

It is not known how long individual bushes of Rhododendron can live. Brown (1953a) reported stems of 60 years, whilst Cross (1975) states the greatest age to be at least 100 years. In mature bushes the age of the oldest stems is limited by mechanical weakness; they become top heavy and are blown over. The centres of the old bushes tend to die out, and the original root system may be replaced by younger, adventitious roots. With age, individual bushes become difficult to distinguish, and maximum age is not easily determined (Shaw, 1984).

Rhododendron is able to spread vegetatively; this is most common on the edge of thickets where unsupported branches become prostrate. Cross (1975) and Brown (1953a) state that adventitious buds are produced from the roots, whereas Shaw (1984) considers that it is often difficult to distinguish between root and stem, and it is from the stem, not the root, that new shoots develop. Many Rhododendron bushes have a length of procumbent stem or layered shoots at their base, whilst others have an accumulation of mosses around the base from which adventitious roots on the stem develop. In respect of the uprooting of bushes as a method of control, aside from ground disturbance creating safe sites, it seems not essential that all roots are removed so long as no piece of stem is left in the ground.

The ability to produce new shoots and coppice means that cutting as a means of control is ineffective, although repeated cutting may eventually lead to death by starvation. Fire has a similar effect to cutting. Normal grass or heather fires only kill young seedlings. Fire may even

favour <u>Rhododendron</u> by destroying other vegetation cover and creating suitable habitats for seedlings (Shaw, 1984)

3.7 Climatic conditions

Despite its southern origin, the distribution of Rhododendron in Britain is unlikely to be limited by temperature, as it endures lower temperatures in Turkey and continental Russia than would be found in the British Isles (Tabbush, 1987). A combination of wind and low temperature probably limit its range in the mountains; in Snowdonia it has been found growing at just over 400m in fairly sheltered conditions, although it seems to produce little seed above approximately 300m (Shaw, 1984). In Ireland Rhododendron has been found growing up to 530m (Cross, 1975).

The ability to tolerate drought, a substantial problem especially during the early years of establishment, depends on the habitat to which it is adapted. Cross (1975) found plants growing in the normally very damp Killarney woods were badly affected by drought in 1972, whereas plants in open habitats in the same area were unaffected. Although able to grow on both north and south facing slopes, growth on south facing slopes tends to be less vigorous (Snowdonia National Park, 1987).

3.8 Edaphic conditions

Rhododendron is an extremely versatile species found growing on a wide range of soils with varying degrees of vigour, flowering in most situations (Shaw, 1984). Most soils supporting R.ponticum have a low cation exchange capacity and low levels of potassium (Cross, 1975). Optimum growth occurs on freely drained, acid soils that do not readily dry

out in summer. Soils containing a high proportion of coarse particles are also favoured (Cross, 1975). Whilst growth is normally poor at a pH of greater than 5.0, Cross (1975) found Rhododendron growing in a range of soils between pH 3.3 and 6.4. At a pH of greater than 5.0, the plants show signs of mineral imbalance, with low levels of manganese in the foliage (Shaw, 1984). Rhododendron is tolerant of high levels of magnesium and as many alkaline soils contain high levels of this cation this may explain the presence of Rhododendron in such soils.

The root system forms a dense, compact mat, ramifying throughout the upper soil horizon. Although preferring deeper soils, a few centimetres of humus over rock is sufficient to support small, fertile bushes. In these situations the plant is highly vulnerable to drought, with death occurring in extreme cases (Shaw, 1984). Cross (1975) states that Rhododendron will not tolerate complete waterlogging of its roots, although it may grow in bogs where the ground is raised enough to avoid direct contact with the water table. Rhododendron has however been found growing directly in bogs, although the plants show premature loss of leaves and nutrient deficiency (Shaw, 1984). In such a situation the plants may form a thick layer of raw humus at the base to enable themselves to be raised above the water table.

3.9 Resistance to herbicides

The resistance of <u>Rhododendron</u> to herbicides may stem from the thick waxy cuticle on the upper surface of mature leaves, the under-side of the leaf being less well protected but less accessible. Immature leaves have a thinner cuticle and bushes are therefore often cut and allowed to regrow before being sprayed. There is also evidence that translocation of herbicide is poor compared with other weedy plants (Shaw, 1984). Chemicals designed to specifically control or eradicate <u>Rhododendron</u> are unlikely to be manufactured, as industrial companies tend to look for 50 million dollar markets (Snowdonia National Park, 1987).

3.10 Associated fauna

Rhododendron provides valuable cover for birds and animals, including blackcaps, chiffchaffs, finches, nightingales, thrushes, otters and badgers (Simons 1988). Despite the shelter afforded to these species, the diversity of bird species tends to be lower in mature oakwoods invaded by Rhododendron than in uninvaded woodlands. Hope Jones (1972) assessed the number of pairs of birds in Welsh woodlands as follows.

	ird density
<u>Pa:</u>	<u>irs per sq km</u>
Open ground	13
Low scrub	130
Medium scrub	178
High scrub	89
Mature oak with sheep	388
Mature oak with gaps and scrub	684
Mature oak with <u>Rhododendron</u>	161
Degenerate oak	67

Becker (1988) states that whilst <u>Rhododendron</u> may be used by mammals and birds for breeding sites, cover and roosts, this may be due to a lack of alternative vegetation. For most species the plant is of little value as a food source; it lacks a high population of insects and has toxic nectar and poisonous leaves.

Following Rhododendron clearance in 1977/8 at the Coombes Valley R.S.P.B reserve in Staffordshire, the diversity of bird species increased (Appendix II), although the overall number of birds using the area was not significantly different. The species that declined either had a notable association with Rhododendron, such as the blackcap and chiffchaff, or were species associated with areas of dense ground cover. Conversely, species showing an increase, or recorded for the first time tended to have a preference for open woodland (Becker, 1988).

4.0 SITE DESCRIPTION

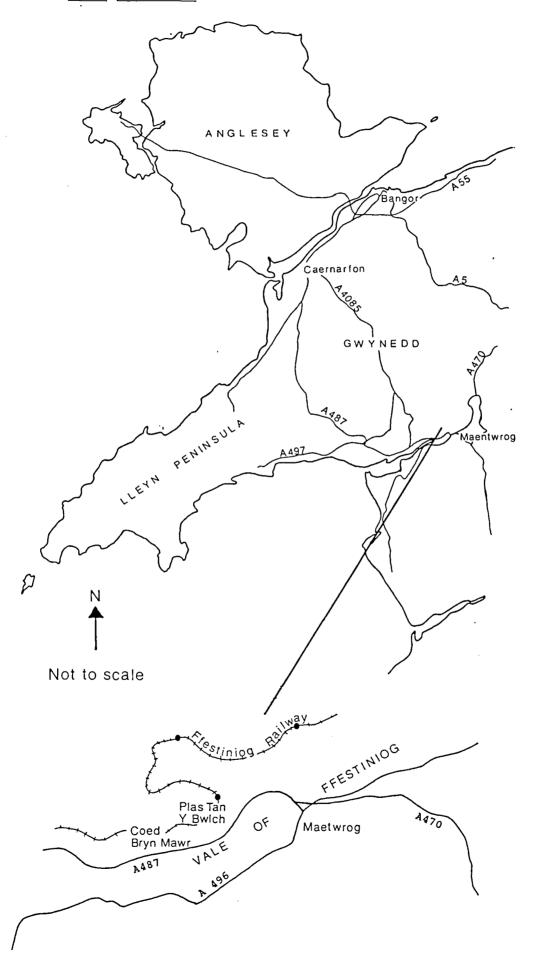
4.1 Introduction

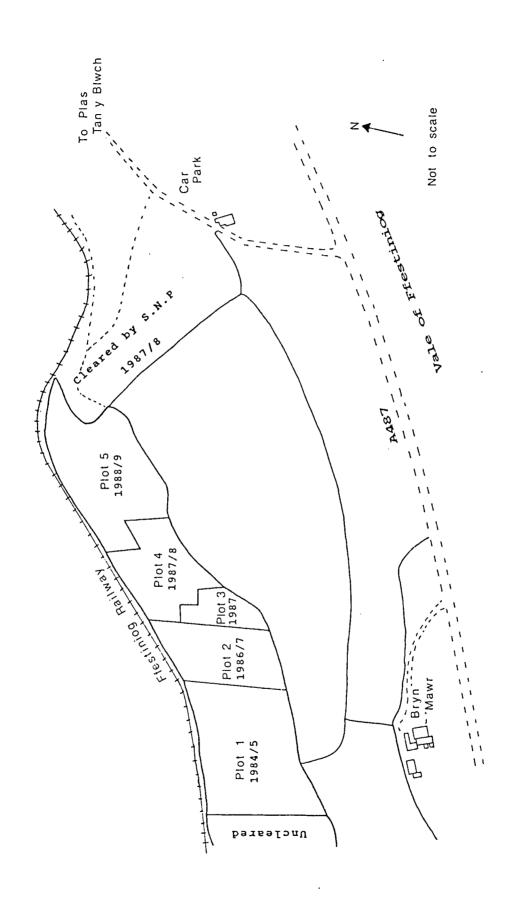
Coed Bryn Mawr (OS 124, Grid ref. 648404), is situated in the Vale of Ffestiniog in North Wales (Map 1). The vale is an important tourist attraction, especially during the summer months when the oakwoods are in leaf and the Rhododendrons in flower. The oakwoods are an intrinsic part of this landscape; their importance is highlighted by the Ffestiniog railway which runs along its northern boundary. The entire site is approximately 3.87 hectares, forming a south facing slope 55 to 95 metres above sea level. The slope, whilst not uniform, is generally steep in the middle part of each plot, with more gentle slopes above and below. Rainfall is estimated at 1700mm per annum. There is no formal grazing pressure, although there are goats in the vicinity. These are only occasional visitors, and no evidence of grazing was found.

Within the wood five plots were defined according to their year of clearance (Map 2).

<u>Plot 1</u>. This is the earliest <u>Rhododendron</u> clearance comprising 1.1 hectares cleared pre-1986, predominantly during the winter of 1984/5. The southern part of this plot has no tree cover. Ammonium sulphamate was used in the initial year of clearance only as there was concern that herbicide might enter the Bryn Mawr water supply. Stumps were cross cut before being treated. Regrowth has been partly removed three times a year for the last three years, although with limited success.

MAP 1 : SITE LOCATION





<u>Plot 2.</u> 0.86 hectares cleared in 1986/7. The <u>Rhododendron</u> was first cut and burnt, followed by brush treatment of stumps with a 50% solution of glyphosate, water, and a chemical dye to show which stumps were treated. Spraying is intended this year.

<u>Plot</u> 3. 0.43 hectares cleared in 1987; the method of clearance was as for plot 2.

<u>Plot 4.</u> 0.58 hectares cleared for approximately £2500 in 1987/8. The method of clearance was as for plot 2.

<u>Plot 5.</u> 0.88 hectares cleared for approximately £3000 in the winter of 1988/9. The method of clearance was as for plot 2.

4.2 Rhododendron cover

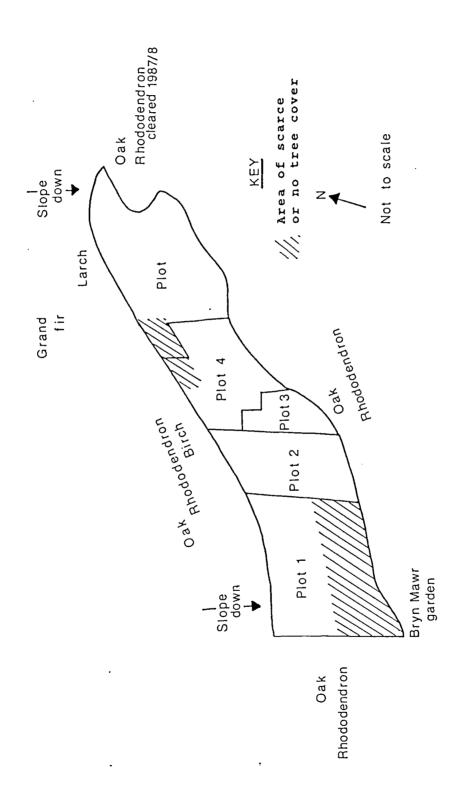
A former <u>Rhododendron</u> cover throughout the site of 75 - 100% was derived from measurement of stumps, and confirmed by the N.C.C warden.

4.3 Tree cover

Basal areas per hectare were derived for the wooded areas of each plot. The areas with scarce, or no tree cover are illustrated in Map 3.

Plot	- <u>-</u>	<u>Basal</u>	<u>area</u>
Plot	1	42.28 31.82	m ² /ha
Plot	2	31.82	m²/ha
Plot	3	27.30	m ² /ha
Plot	4	30.03	m^2/ha m^2/ha
Plot	5	32.45	m²/ha

This gave an average basal area for the site as a whole of $32.78 \text{ m}^2/\text{ha}$.



4.4 Species composition

The wood is a predominantly mature sessile oakwood with some Betula pubescens (Figure 1). The mature Pseudotsuga menziesii, Tilia cordata and Fraxinus excelsior were recorded in plot 5 which appeared more diverse in species than the other plots. Areas in plots 1, 4 and 5 with scarce, or no tree cover are shown at Map 3.

Table 2 shows the stand composition for the site. The dbh classes 21 - 30cm, 31 - 40cm and 41 - 50cm each comprise approximately 23 - 24% of the stand, with 13% of the stand in the dbh range 11 - 20cm. The structure appears sufficiently diverse to allow regeneration with the majority of the stand capable of producing seed.

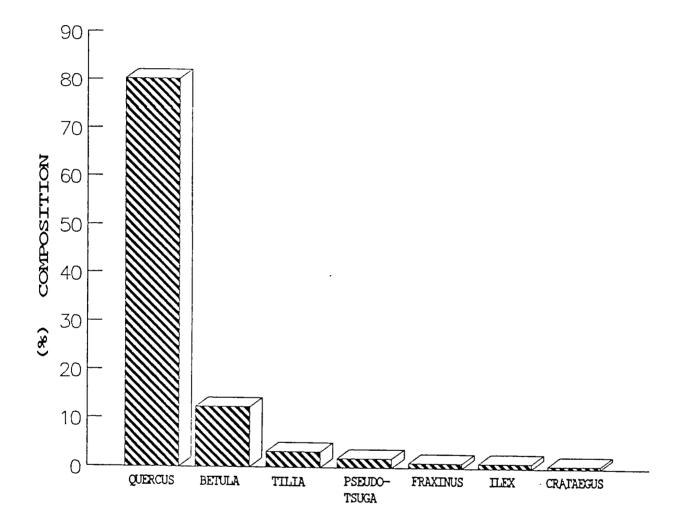
TABLE 2 : SITE STAND COMPOSITION (DBH CLASSES)

DBH Class (cm)	Number of trees	<u>%</u>
7 - 10	8	3.9
11 - 20	28	13.4
21 - 30	48	23.0
31 - 40	50	23.8
41 - 50	49	23.4
51 - 60	12	5.6
61 - 70	8	3.9
71 - 80	1	0.5
81 - 90	2	1.0
91 - 100	2	1.0
100 - 111	1	0.5
	209	100

4.5 Adjacent vegetation

The vegetation adjoining the site (Map 3), is predominantly sessile oak with birch and impregnable <u>Rhododendron</u>, reminiscent of the infestation that occurred within Coed Bryn Mawr before clearance. To the east is the National Park's Plas Tan y blwch Estate, the area adjoining the wood also being predominantly sessile oakwood (Domin score 9).

FIGURE 1 : TREE SPECIES COMPOSITION FOR THE SITE



This is the only adjoining area in which the <u>Rhododendron</u> has been cleared. The clearance was undertaken in 1986. Birch and ash (Domin scores 3 and 5 respectively), are also present.

To the North of plot 5 is a larch plantation, running along approximately 50% of the boundary to the plot. Beyond this a plantation of grand fir extends along the boundaries of both plots 4 and 5.

Immediately to the south of plot 1 is Bryn Mawr garden. The species present include sycamore, oak, ash, wild cherry, holly, hawthorn, giant sequoia and <u>Cryptomeria japonica</u>.

4.6 Geology

A management report for the Maentwrog National Nature Reserve (Howells 1967), gives details of the geology of the reserve which covered an area extending east from Coed Bryn Mawr. The geology of the Maentwrog National Nature Reserve, and Coed Bryn Mawr appears similar, although no independent analysis was undertaken for Coed Bryn Mawr. The area has a shale rock outcrop running east to west, dividing Coed Bryn Mawr along its long axis. The soil is largely shallow brown earth, with a transition to brown podzolic soil formed from frost shattered shale. The pH in the surface horizon ranges from 4.0 to 4.6.

5.0 EXPERIMENTAL DESIGN AND COLLECTION OF DATA

5.1 Germination trials

Twenty soil samples were collected in early May from 10 locations, 5 from plot 5 (cleared in winter 1988/9) and 5 from plot 4 (cleared in winter 1987/8). Only a limited number of trials could be undertaken in the time available and the location of samples was therefore subjective. Samples 1/1a to 6/6a were in an area dominated before clearance by dense Rhododendron thickets and would show the latent effects, if any, of the Rhododendron infestation. The area was sparsely populated with trees and samples 7/7a to 10/10a were chosen as a contrast, being within the mature, predominantly oak canopy. All samples were within an area of 75-100% former Rhododendron cover.

Twigs, branches and leaf litter were removed before the samples were collected. The top 5cm layer of soil was cut to the size of a seedtray (185mm x 270mm) to give approximately equal quantities of soil in each sample.

The samples were taken to the University's Nant Porth Nursery, and treated with the aim of ensuring early and maximum germination. A small quantity of Moist Irish Shamrock peat was placed on the base of each seedtray to prevent excessive moisture loss, and the soil samples were scattered on top of the peat. Three separate treatments were then undertaken as described below.

i) Germination of seed within the seedbank

One seedtray from each sample plot was covered with a thin layer of horticultural gravel (2-3 mm grain size, neutral pH). This had a similar function to the layer of peat in preventing excessive moisture loss and in view of the

exceptionally hot weather proved valuable.

ii) Trial sowings

The remaining seedtrays were sown with <u>Festuca ovina</u> and <u>Betula pendula</u> seed, each species being sown into separate halves of the seedtrays. <u>Betula pendula</u> seed (Kincardine Forest provenance, 1987, E.Scotland), was obtained from the Forestry Commission Seed Branch, no <u>Betula pubescens</u> being available at the time. For each seedtray 0.5g of <u>Betula pendula</u> seed was stratified for three weeks, being mixed with moist Irish Shamrock peat and placed in plastic bags in a refrigerator kept at between 2° and 5° C.

In order to sow similar quantities of <u>Betula</u> and <u>Festuca</u> seeds, a number of samples of both seed were counted (Appendix III). The average number of <u>Betula</u> seeds stratified and sown, and therefore the number of <u>Festuca</u> seeds sown, was 2157. In retrospect this proved too many, and it would have been prudent to have initially measured a smaller quantity of <u>Betula</u> seeds.

The seed once scattered in the seedtrays was covered with a thin layer of horticultural gravel.

iii) Control sowings

To provide a control for the trial sowings, <u>Festuca</u> and <u>Betula</u> seed stratified as above were sown into separate halves of a seedtray filled with moist Irish Shamrock peat. In common with the other experiments the seeds were then covered with a thin layer of horticultural gravel.

All seedtrays were placed on a bench over three layers of capillary matting, intended to provide a moisture retaining surface. The seedtrays were watered as required.

During the stages of abundant germination (weeks 2 to 5), the seeds germinating in the test sowing experiment were counted and removed twice a week. As the rate of germination declined (by week 6), this was reduced to once a week. Germination was very limited in the experiments designed to assess the viable seedbank, and the majority of seedlings were left in the seedtrays throughout the experiment.

Although the method described above allowed comparison of germination rates between treatments, a measure of dry weight (combined with a smaller number of seeds sown), may have provided a better comparison of the ability of the soil to support growth beyond germination. A measure of dry weight would therefore be recommended in future experiments.

5.2 <u>Collection of vegetation succession data</u>

5.2.1 Tree density and Rhododendron cover

To assess tree basal area and <u>Rhododendron</u> cover nineteen 20 x 20m quadrats were surveyed, four in plots 1, 2, 4 and 5, and three in the smaller plot 3. This gave an area surveyed of 0.76ha, equivalent to approximately 19.7% of the site. The nineteen quadrats were placed randomly; the areas of scarce tree cover in plots 1, 4 and 5 were ignored. The species and diameter at breast height (dbh), of each tree in the quadrats was recorded, and a basal area derived for plots 1 to 5 and for the site as a whole.

It had been intended to assess <u>Rhododendron</u> density by recording stump diameter within each quadrat. This proved impracticable for several reasons. Stumps cut close to the ground were buried under litter, smaller ones being easily missed, whilst others were destroyed where bonfires had been

sited. Mature bushes with large stools but few stems also made accurate measurement difficult.

As an alternative the strategy adopted by the R.S.P.B (Becker 1988) was used (Figure 2). Three categories of density are specified, with the corresponding cover, height, root ball and trunk diameter identified. This method allowed a number of stumps to be measured and a classification derived.

5.2.2 <u>Vegetation</u> survey

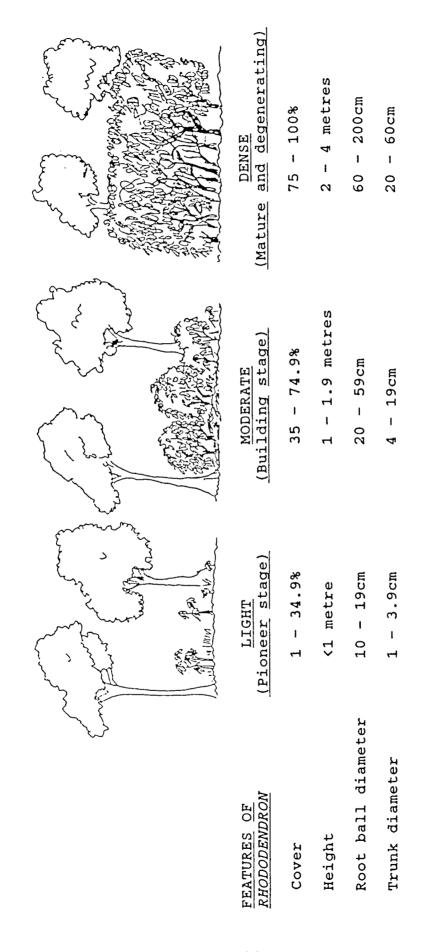
The vegetation was surveyed in mid July, using 35 randomly placed 1 x 1m quadrats in each plot. Whilst most individual grasses were identified, sedges were recorded as <u>Carex spp</u> and bryophytes as bryophyte spp. Abundance for each species was assessed using the Domin scale (Table 3).

TABLE 3 THE DOMIN SCALE

Cover	Domin scale
Cover > 100% Cover > 75%	10
Cover 51 - 75%	9 8
Cover 34 - 50%	7
Cover 25 - 33%	6
Abundant, cover approximately 20	% 5
Abundant, cover approximately 5%	4
Scattered, cover small	3
Very scattered, cover small	2
Scarce, cover small	1
Isolated, cover small	X

An assessment of mean percentage crown cover was made at the centre of each quadrat using a spherical densiometer (Mayhead 1969). Four readings were taken for each quadrat (north, south, east and west) and the average taken as the mean percentage crown cover. The data for each plot is shown at Appendix IV to VIII.

FIGURE 2: CATEGORIES OF RHODODENDRON DENSITY



5.2.3 Adjacent vegetation

Adjacent vegetation was assessed visually from the perimeter of the site and by access to adjacent areas where possible. Only species in the shrub and canopy layer were recorded.

6.0 ANALYSIS OF DATA

6.1 Data analysis techniques

Multivariate techniques of classification and ordination, followed by statistical analyses were undertaken. The VESPAN package (Malloch, 1988) was produced for the N.C.C, and specifically designed to handle and analyse vegetation data. Within VESPAN the programmes RECORD, PREPARE, TWINSPAN and DECORANA were used.

Data is entered using RECORD, which allows simultaneous checking of data for accuracy. Rhododendron coppice and seedlings were entered using separate codes, as were Quercus petraea seedlings and stumps. PREPARE converts the data from RECORD into a form acceptable for the TWINSPAN and DECORANA programmes, and also allows amalgamation of species - such as Rhododendron coppice and seedling to Rhododendron species. Whilst the amalgamation of grasses and of Rhododendron coppice and seedling was tried, in the final analysis no amalgamations were used as given the small number of species recorded the result was more comprehensive.

TWINSPAN produces a dichotomy based on the principal axes of an ordination, giving two groups of samples which are themselves re-ordinated and re-divided. This continues until the number of divisions requested is reached. Each division is the result of a three stage analysis.

- i) <u>Primary ordination</u>. Species indicative of a group of samples, known as differential species, are identified.
- ii) <u>Refined ordination</u>. Each of these differential species is assigned a value of +1, and the cumulative score for each sample is derived. The samples are divided at an appropriate

point and depending on whether they have a score greater or less than a critical value are assigned to the positive or negative group, thereby producing a dichotomy.

iii) <u>Indicator ordination</u>. With the exception of borderline cases, it is the refined ordination that is used to determine the dichotomy. By using the differential species that are most characteristic of a division, the indicator ordination provides a simple criterion for re-identification of the groups, and using this assigns borderline cases to the positive or negative side of the dichotomy.

The dichotomy not only relies on frequency, but by the use of pseudospecies takes account of quantitative data. Using the Domin scale to record abundance results in four pseudospecies levels being used.

Pseudospecies level	<u>Domin</u> <u>scale</u>
1	> 0
2	> 2
3	> 4
4	> 7

As only 31 species were recorded, three levels of division were used giving eight sample classifications (Figure 3). A fourth or fifth division would result in 16 or 32 classifications respectively, and to achieve this with so few species TWINSPAN would divide groups into artificial associations using poor indicator species. Whilst this study aims to analyse the vegetation that succeeds following the clearance of Rhododendron, this species has not been omitted from the analysis as it is an integral part of the vegetation, and its presence may ultimately dictate future species composition. Species names are shortened to eight letters in the TWINSPAN output, the abbreviated names being

shown at Appendix IX. <u>Rhododendron</u> seedling is abbreviated to RHOD PONS, and coppice to RHOD PONC. Likewise <u>Quercus</u> <u>petraea</u> seedling becomes QUER PETS, and the tree stump QUER PETT.

Having ordinated the samples, TWINSPAN then ordinates and divides the species, using the sample classification as a basis (Figure 4).

The final output from TWINSPAN is an ordered two way table (Table 4) combining the sample and species divisions. The samples are arranged from left to right in the order of the sample division with the species being arranged from top to bottom in the order of the species division. The sample and species groups are shown in binary coding along the bottom and right hand margin of the table. The table can be used as a visual illustration of the homogeneity of samples within each group. If a fourth or fifth division had been requested, a further 8 or 24 groups would be created, and the homogeneity of the associations lost.

DECORANA is also an ordination technique, using the same input from PREPARE as TWINSPAN. Down-weighting of rare species is possible whereby any species that has a frequency in the data of less than 20% of that of the commonest species is down-weighted in proportion to its frequency. Species such as <u>Umbilicus rupestris</u> (occurring only once), whilst still retained in the analysis is therefore given less weight than the frequently occurring species. DECORANA produces axes of ordination for both samples and species: the final output is a two dimensional scatter diagram in which similar entities are located close together and dissimilar entities far apart. Environmental gradients can

usually be associated with the axes. An eigenvalue is calculated for each axis and where this is much less than that of axis 1 the axis is unlikely to be of much significance. For both the sample and species ordination only the first two axes were of significance.

VESPAN ordinates quadrats according to the vegetation parameters defined. No measure of bare ground was made, and consequently where a quadrat has no vegetation it is not included in either the TWINSPAN or DECORANA ordinations. Their are 6 quadrats with no vegetation, 1 in plot 2, 4 in plot 4 and 1 in plot 5.

Statistical analyses were then undertaken to assess the relationship between crown cover and vegetation parameters, and vegetation and the number of years since clearance. Statistical analyses were also used for the germination trials to assess germination rates between different treatments.

FIGURE 3 : TWINSPAN SAMPLE CLASSIFICATION

DIVISION 1	# 0 # 8rva	Negative (-ve)	(-۸۵)	N = 169		Positive (+ve)	 ; (48) ubu frut (3)	
DIVISION 2	Rhod	Rhod-ponc (1) Rhod-pons (1)				Hede Pter	Hede heli (2) Pter aqui (1 ¦	
# 00 (25) Rhad ponc (2)	+ 00 (25) od ponc (2)	4	• 01 (36)		* # # % # % % %	† 10 (32) Hede heli (2) Bryo spp (1) Rubu frut (4)	→ ₹ 5	+ 11 (16) Pter aqui (1) Cham angu (1)
# 000 (2) Ruer pett (1) 432 415	# 001 (23)	# 010 (8) Betu pube (1) Digi purp (1) Rhod pons (1) 105 215 112 205 510 104 502 535 128 523 520 214 523 520 214 427 427 427 427 427 524 524 524 524 525 525 527 526 527 527 527 527 527 527 527 527 527 527	+ 011 (88) 1)	2) 2) 102 204 217 227 227 227 233 333 411 411 422 530 531	# 100 (14) Rubu frut (4) Bryo spp (1) [111 109 123 129 134 108 135 222 324 305 320 313	# 101 (18) Hede heli (3) 125 122 117 226 221 332 315 326 303 327 331 501 506 528 519 503 525 534	# 110 (12) Rubu fruc (1) 107 114 113 116 120 211 210 335 531 511 512 508	# 111 (4) 133 119 223 229 233 229

FIGURE 4: TWINSPAN SPECIES CLASSIFICATION

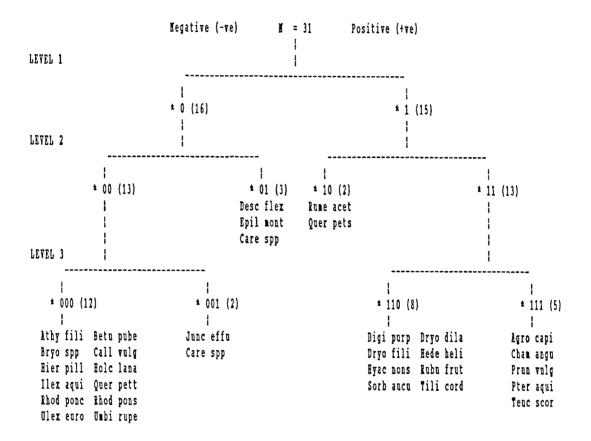


TABLE 4: TWINSPAN TWO WAY TABLE

-2	
3	
7	
11211122 1-2	
3. 3332223232323333333 1-[-	-1
44243233433222142333444	
2-2-3	
- - - - - - - - - - - - -	
	232- 44443342344344433312-
234444444444444444444444444444444444444	4444-444 -3234341 31333334334
	22-2
-3	-1

6.2 Data analysis

6.2.1 <u>Vegetation</u> <u>associations</u>

The TWINSPAN stand classification (Figure 3), and species classification (Figure 4) can be read in conjunction, since species occurring on the left hand side of Figure 3 will occur on the left hand side of Figure 4. Given the relatively small data set however, Figure 3 alone provides sufficient information.

The first division uses <u>Rhododendron</u> coppice, seedling, and bryophytes as the basis of the negative division, whilst bramble, ivy and bracken form the basis of the positive division. This concurs with the literature reviewed (section 3.4), in that the <u>Rhododendron</u> seedling is associated with the most favourable safe site of a low level of bryophytes (pseudospecies level 1). In contrast the dense vegetation on the positive side of the division provides far fewer safe sites.

The negative group is then divided at level 2 by Rhododendron coppice (pseudospecies level 2). The positive group is divided on the negative side by ivy (pseudospecies level 2), bryophytes (pseudospecies level 1), and bramble (pseudospecies level 4), whilst the positive group is characterised by bracken (pseudospecies level 1), and rosebay willowherb (pseudospecies level 1).

The structure and abundance of vegetation within each TWINSPAN group can be derived by accumulating the Domin score of each species in each group, and converting this score to a percentage. The Domin score is used for this purpose as the TWINSPAN dichotomy is itself derived from quantitative data.

I) <u>Group *000</u>

Number of quadrats: 2

Species present	<pre>% Domin</pre>
Quer pett (stump) Bryo spp	50 27
Rhod ponc (coppice)	23

TWINSPAN has separated the two quadrats which partly fell on mature oak stumps; the remaining ground was covered by bryophytes and <u>Rhododendron</u> coppice.

II) Group *001

Number of quadrats: 23

Species present	<pre>% Domin</pre>
Bryo spp Rhod ponc (coppice)	46 44
Betu pube	5

Species occurring at 2-3%: Rhod pons (seedling)

Species occurring at 1-2% : Athy fili

Species occurring at <1% : Care spp; Desc flex; Quer pets

The group can be classified as dominated by **bryophytes** and **Rhododendron coppice**.

III) Group *010

Number of quadrats: 8

Species pre	<u>esent</u>	<pre>% Domin</pre>
Betu pube Bryo spp		38 31
Digi purp Rhod ponc	(coppice)	12 7 7

Species occurring at 2-3%: Care spp; Rhod pons (seedling)

Dominated by <u>Betula pubescens</u> and <u>bryophytes</u>, with <u>Digitalis</u>

<u>purpurea</u>, <u>Rubus fruticosus</u>, and <u>Rhododendron</u> coppice also present.

IV) Group *011

Number of quadrats: 88

<u>Species</u>	<pre>% Domin</pre>
Bryo spp	60
Betu pube Rubu frut	7 7
Rhod pons (seedl	ing) 5

Species occurring at 2-3%: Care spp; Hede heli; Epil mont; Rhod ponc (coppice).

Species occurring at 1-2%: Holc lana; Junc effu; Quer pets; Rume acet.

Species occurring at <1%: Agro capi; Athy fili; Call vulg; Cham angu; Desc flex; Digi purp; Dryo dili; Hier pill; Pter aqui; Umbi rupe; Ulex euro.

Dominated by bryophytes , with <u>Betula pubescens</u>, <u>Rubus</u> fruticosus and Rhododendron seedlings also present.

V) <u>Group *100</u>

Number of quadrats: 14

<u>Species</u>	<pre>% Domin</pre>
Rubu frut Bryo spp	50 23
Hede heli	6

Species occurring at 4-5%: Rume acet. Species occurring at 3-4%: Betu pube.

Species occurring at 2-3%: Care spp; Dryo fili; Epil mont.

Species occurring at 1-2% : Digi purp; Dryo dili.

Species occurring at <1% : Agro capi; Cham angu; Junc effu; Sorb auc.

The group can be classified as <u>Rubus fruticosus</u> and bryophytes, with to a lesser extent <u>Hedera helix</u>.

VI) <u>Group *101</u>

Number of quadrats: 18

<u>Species</u>	<pre>% Domin</pre>
Hede heli	61
Rubu frut Bryo spp	14 6

Species occurring at 3-4%: Quer pets.

Species occurring at 2-3%: Dryo fili; Rhod ponc (coppice).

Species occurring at 1-2%: Betu pube; Digi purp; Frax exce;

Hyac nons; Pter aqui; Rhod pons (seedling).

Species occurring at <1%: Care spp.

Dominated by Hedera helix, with Rubus fruticosus and bryophytes.

VII) Group *110

Number of quadrats: 12

<u>Species</u>	<pre>% Domin</pre>
Rubu frut	38
Pter aqui	26
Cham angu	10
Agro capi	6
Bryo spp	6

Species occurring at 4-5%: Digi purp; Epil mont.

Species occurring at 2-3%: Desc flex.

Species occurring at 1-2%: Betu pube; Hede heli; Quer pets.

Dominated by <u>Rubus fruticosus</u> and <u>Pteridium aquilinum</u>, with a substantial presence of <u>Chamaenerion angustifolium</u>, <u>Agrostis capillaris</u> and <u>bryophytes</u>.

VIII) Group *111

Number of quadrats: 4

<pre>% Domin</pre>
65
15 6

Species occurring at 4-5%: Bryo spp; Prun vulg. Species occurring at 2-3%: Epil mont; Quer pets.

Although dominated by <u>Pteridium aquilinum</u>, the proportion of <u>Teucrium scorodonia</u> is misleading as this TWINSPAN group comprises only four quadrats, in which this species occurs only once.

The distribution of quadrats between TWINSPAN groups is

shown in Table 5, and the vegetation associations summarised in Table 6.

TABLE 5: DISTRIBUTION OF QUADRATS BETWEEN TWINSPAN GROUPS

Plot		TWINSPAN stand group Negative (-ve) Positive (+ve)								
	*000	*001	*010	*011	*100	*101	*110	*111		
1	_	1	1	14	9	3	5	2		
2	-	3	2	22	1	2	2	2		
3	-	5	-	19	4	6	_	_		
4	2	11	_	18	_	-	_	_		
5	_	3	5	15	_	7	4	-		
Total	2	23	8	88	14	18	11	4		

Over 50% of the quadrats fall into group *011, the association dominated by 60% bryophytes with birch, bramble, and Rhododendron seedlings. This is therefore the predominant association for the site. Of the 88 quadrats in this group the individual plots are represented as follows, each plot having a greater proportion of its quadrats within this vegetation association than any other association.

Group *001 Plot	Number of quadrats
1	14
2	22
3	19
4	18
5	15

The second most frequent association is group *001, with approximately 13% of the quadrats. This is the bryophyte (46%) and Rhododendron coppice (44%) association, with birch (5%). The occurrence of individual plots within the group is in contrast to group *011 not evenly distributed, plot 4 being the only plot substantially present.

TABLE 6: SUMMARY OF VEGETATION ASSOCIATIONS

	*111 (4)	Pter aqui	!	 	; ;	; ;	Teuc scor	Hede heli
irats)	*110 (12)	1	‡ †	<u> </u>	Rubu frut	Pter aqui	1	Cham angu Agro capi Eryo spp
	*101 (18)	 	Hede heli	!	 -	 	Rubu frut	Bryo spp
(number of quadrats)	*100 (14)	¦ ¦	 	Rubu frut	1	Bryo spp	 -	Hede heli
TWINSPAN groups (n	*011 (88)	} }	Bryo spp	! !	} !	; ;	!	Betu pube Rhod pons Rubu frut
TVINSPA	*010 (8)	;	!	! !	Betu pube Bryo spp	{ {	Digi purp	Rubu frut Rhod ponc
	*001 (23)	} {	1	Bryo spp Rhod ponc	} 	{ 	1	Betu pube
	<u>*000</u>	;	1	Quer pett	!	Bryo spp Rhod ponc	} }	} }
& Domin		> 61	51.0 - 60.9	41.0 - 50.9	31.0 - 40.9	21.0 - 30.9	11.0 - 20.9	5.0 - 10.9

Group *011 Plot	Number of quadrats
1	1
2	3
3	5
4	11
5	3

The only other association on the negative side of the dichotomy with several quadrats is group *010, again with byrophytes substantially present (38%), and also birch (31%). There are only 8 quadrats within this group, 5 from plot 5, 2 from plot 2 and 1 from plot 1.

Three of the associations on the positive side of the dichotomy comprise over 25% of the quadrats. Group *100, the bramble (50%) and birch (23%) association, principally contains plots 1 (9 quadrats) and 3 (4 quadrats). Group *110 is also a bramble association (38%) with bracken (26%), mainly consisting of plots 1 (5 quadrats) and 5 (4 quadrats). Group *101 whilst having a substantial presence of bramble (14%), is dominated by ivy (61%), with the principal plots being plots 3 (5 quadrats) and 5 (7 quadrats).

Of the quadrats from plot 5 that are on the positive side of the dichotomy, all 4 in group *110 are clustered in the northern (upslope) part of the plot and may be a result of disturbance associated with the railway line. The 7 quadrats in group *101 comprise 4 whose vegetation may also be a result of disturbance, and 3 in a small pocket that may have been unaffected by Rhododendron, as indicated by the presence of ivy. Disturbance may therefore account for plot 5 having a more diverse and abundant species composition than plot 4. Plot 4 may also have suffered from a

particularly low mean monthly rainfall in June 1988, when only 27.4mm of rainfall was recorded at Porthmadog (approximately 8km to the west), and 32.0mm at Blaeneau Ffestiniog (approximately 6km to the north east).

The principal associations are on the negative side dominated by bryophytes and <u>Rhododendron</u> coppice, and on the positive side by bramble, ivy, bracken and bryophytes. The diversity of associations within plots also decreases from plots 1 and 2, the most diverse, to plots 3, 4 and 5, plot 4 being the least diverse.

6.2.2 Crown cover and vegetation associations

Further interpretation of the ordination of the TWINSPAN groups can be derived from Figure 5 where the DECORANA sample ordination has been coded to accord with the TWINSPAN groups. Comparing this to the DECORANA species ordination (Figure 6), shows the correlation of species to TWINSPAN groups, with for example group *110 in a corresponding position to the light demanding species bracken and rosebay willowherb, whilst group *101 corresponds with ivy, a more shade tolerant species.

The mean and standard deviation for the percentage absence of crown cover for each TWINSPAN group is shown in Figure 7. Despite the high standard deviation from the mean of most groups, there appears to be a trend between crown cover and vegetation, with the TWINSPAN groups containing light-demanding species such as bracken, rosebay willowherb and foxglove having a greater percentage absence of crown cover than groups dominated by more shade tolerant species such as ivy, Rhododendron and bryophytes.

FIGURE 5 : DECORANA SAMPLE ORDINATION CODED TO ACCORD WITH TWINSPAN GROUPS

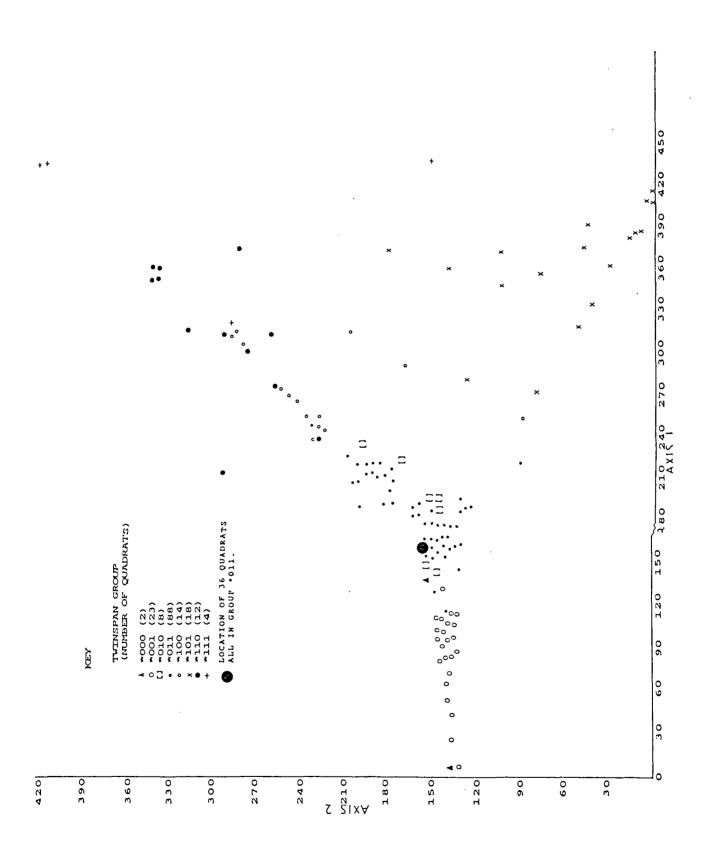


FIGURE 6: DECORANA SPECIES ORDINATION

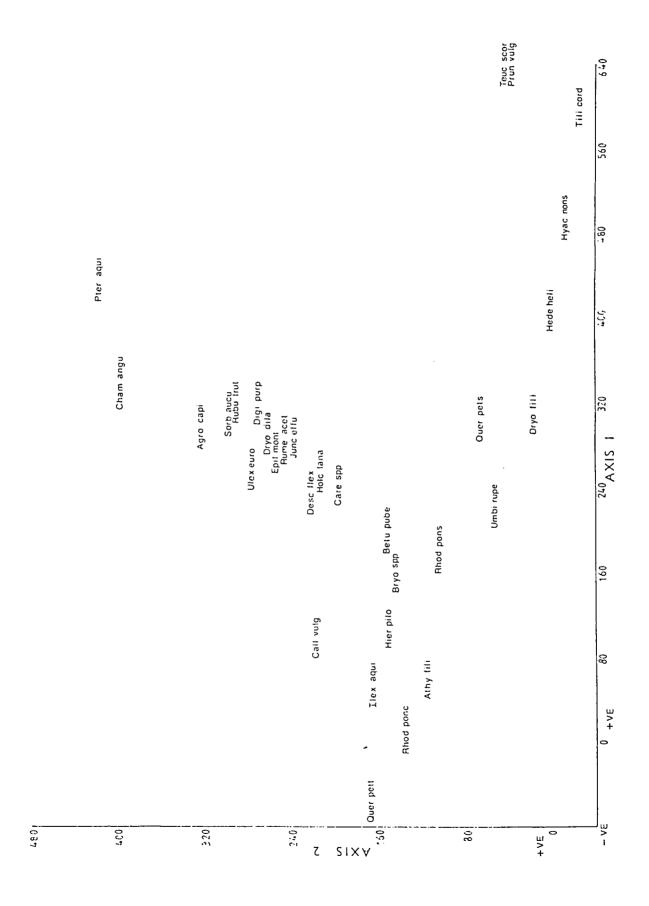
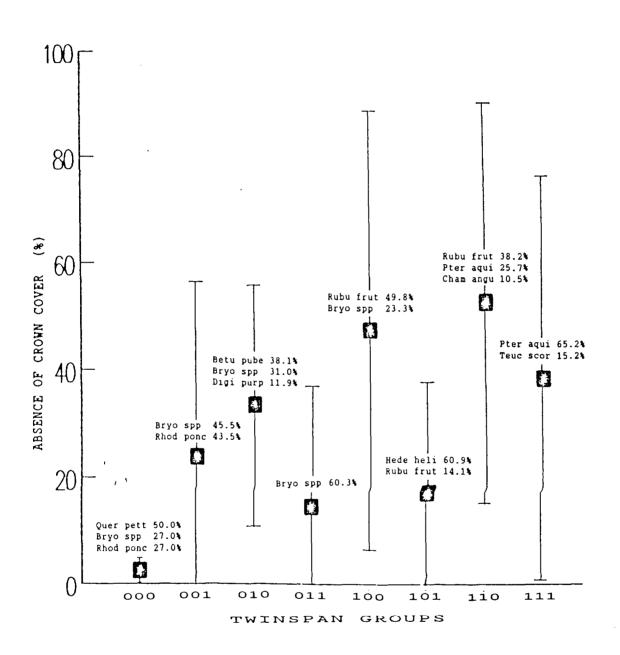


FIGURE 7: MEAN AND STANDARD DEVIATION OF THE ABSENCE OF CROWN COVER FOR EACH TWINSPAN GROUP

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This relationship was statistically analysed using a one-way analysis of variance (Table 7), which identified a significant variation in crown cover between the TWINSPAN groups, with a probability of the variation occurring by chance of less than 0.001%.

TABLE 7 ANALYSIS OF VARIANCE FOR LIGHT INTENSITY BETWEEN
TWINSPAN GROUPS

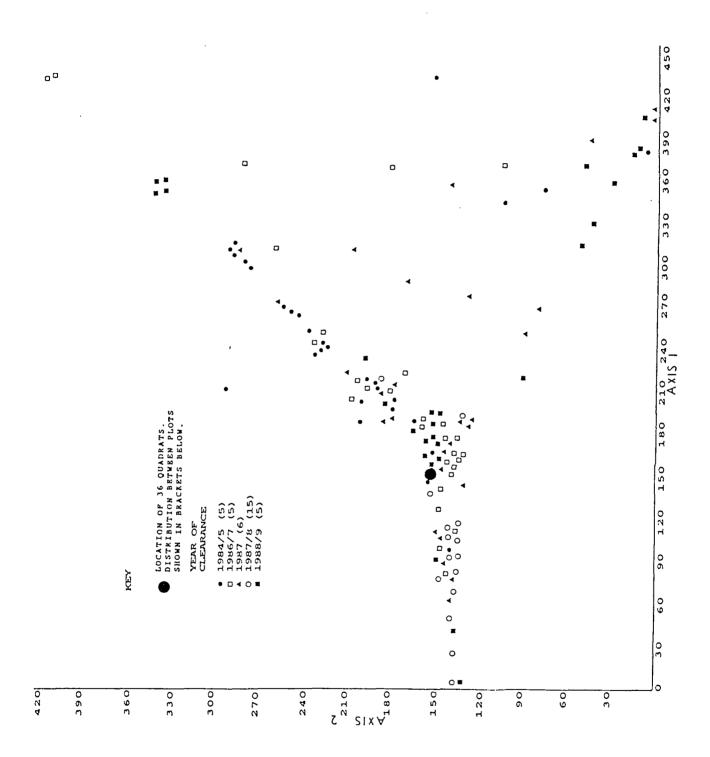
SOURCE	$\underline{D.F}$	<u>s.s</u>	M.S	<u>F</u>	<u>P</u>
FACTOR ERROR TOTAL	7 161 168	28563 122670 151232	4080 762	5.36	0.000

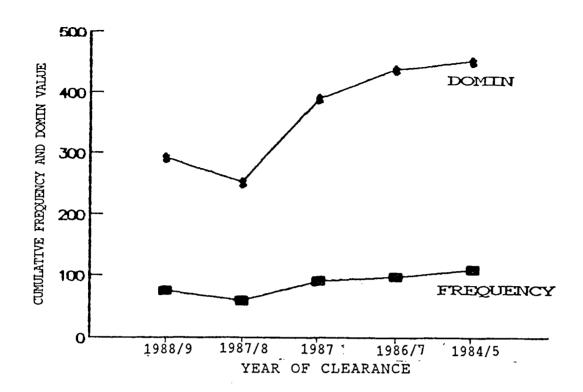
A Scheffy's test was undertaken to produce further analysis (Appendix X) which identifies two principal differences, between groups *011 and *100, and *011 and *110. Group *001 has a low mean percentage crown cover, whilst groups *100 and *110 have the two highest mean percentage crown covers.

6.2.3 <u>Vegetation succession and year of clearance</u>

The Decorana stand ordination (Figure 8), with the year of clearance only shown, illustrates quadrats of the same year of clearance tending to cluster or ordinate in a similar direction, indicating that there is a correlation between the year of clearance and vegetation that succeeds. A comparison of the cumulative frequency and Domin values of all species in each plot (Figure 9), gives a further illustration of this trend, with the abundance of vegetation increasing as the number of years since clearance increases.

FIGURE 8: DECORANA SAMPLE ORDINATION SHOWING YEAR OF CLEARANCE ONLY





To gain a more precise analyse of this relationship a regression analysis was undertaken using frequency data.

TABLE 8: ANALYSIS OF VARIANCE FOR THE REGRESSION
BETWEEN VEGETATION SUCCESSION AND YEAR OF CLEARANCE

The regression equation is Y = 3.82 - 0.335 FY = Year

F = Frequency (number of species in each quadrat)

Analysis of variance

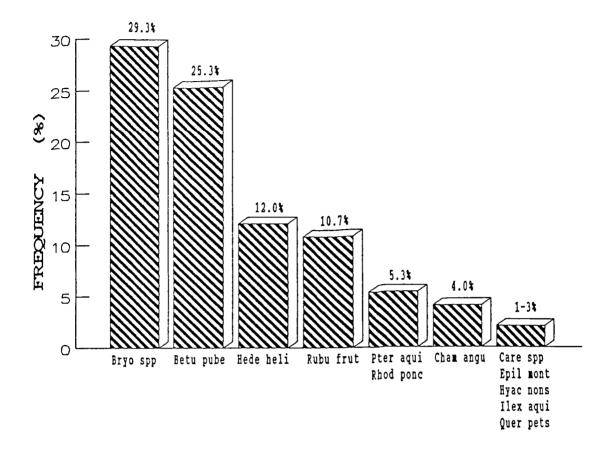
SOURCE	$\underline{D.F}$	<u>s.s</u>	<u>M.S</u>	<u>F</u>	<u>P</u>
REGRESSION ERROR TOTAL	1 173 174	35.197 314.803 350.000	35.197 1.820	19.34	0.000

The regression indicates a significant negative correlation between the year of clearance and the mean number of species within a quadrat, with a probability of the relationship occurring by chance of less than 0.001%. The frequency of species increases as the number of years since clearance increases. The regression was also undertaken using the Domin data, and transformed frequency and Domin data. All regressions produced significant negative correlations, the most significant correlation being derived from the untransformed Domin data.

6.2.4 Species composition within plots

Having established that the abundance of vegetation is correlated to the year of clearance the structure of the vegetation in the different years is of interest. By calculating the percentage frequency of each species in each year, these characteristics can be derived. For ease of interpretation the analysis is set out using the most recent year of clearance as the base year.

I) FIGURE 10 : SPECIES COMPOSITION IN PLOT 5 Cleared in 1988/9

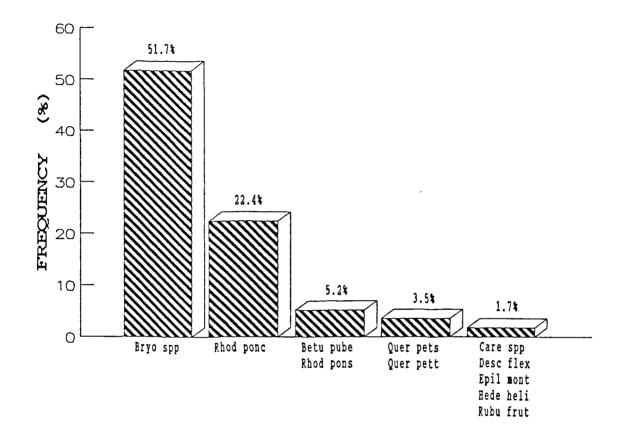


The dominant species are bryophytes and birch seedlings. Although ivy and bracken are also present (12% and 10.7% respectively), their presence is entirely restricted to the northern (upslope) part of the plot, close to the railway line. Bracken, rosebay willowherb, broadleaved willowherb and bluebell are also restricted to this location. Maintenance of the railway line, including the replacement of all sleepers in the last 20 years, has inevitably led to disturbance along the northern boundary of the site; the vegetation adjoining the railway now reflects that

disturbance. Without the species resulting from that disturbance the diversity of vegetation within the plot would be substantially reduced.

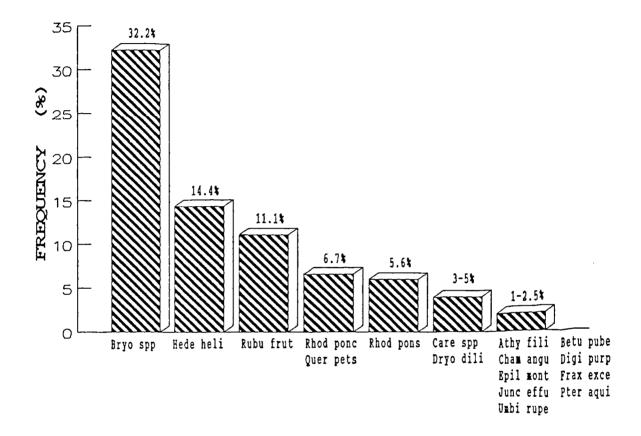
Whilst the colonisation by bryophytes and birch is limited to the area cleared in 1988, their rapid invasion is significant. The difficulty experienced by Rhododendron in finding safe sites for establishment is shown by this being the only plot in which no Rhododendron seedlings were recorded.

II) FIGURE 11: SPECIES COMPOSITION IN PLOT 4 Cleared in 1987/8



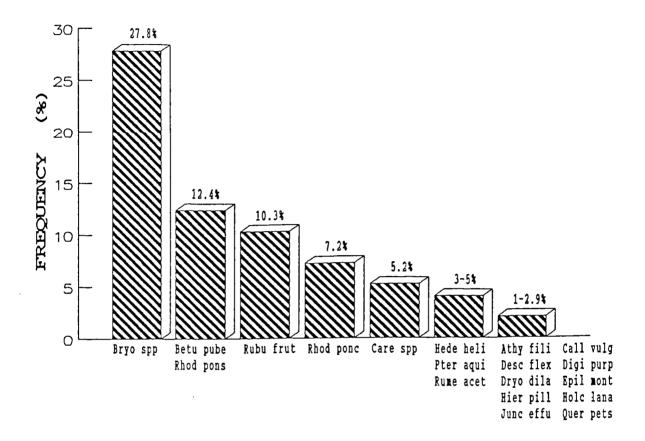
Whilst bryophytes have succeeded to a frequency of over 50%, the presence of <u>Rhododendron</u> coppice (22.4%) highlights the limited success of the initial clearance. For the first time <u>Rhododendron</u> seedlings are beginning to establish as safe sites are generated. Grasses and sedges (<u>Carex</u> spp and wavy hair grass) are beginning to colonise.

III) FIGURE 12: SPECIES COMPOSITION IN PLOT 3 Cleared in 1987



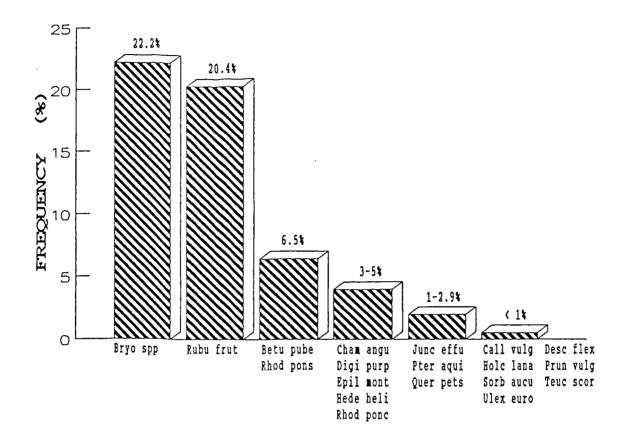
Although bryophytes (32.2%) are the most frequently occurring species, ivy and bramble are present at 14.4% and 11.1% respectively. The substantial presence of ivy and bramble may in part be due to a small area of vegetation in the centre of the plot with little former Rhododendron cover, and to the vegetation recorded in the quadrats in the northern (upslope) boundary of the plot. Rhododendron coppice and seedling are present at a frequency of 6.7% and 5.6% respectively. The number of flowering plants has increased.

IV) FIGURE 13: SPECIES COMPOSITION IN PLOT 2 Cleared in 1986



The plot has a greater diversity of species than the previous plots, with bryophytes (27.9%), still the most frequently occurring species. There is a considerable presence of Rhododendron seedling (12.4%), birch (12.4%), bramble (10.3%), and Rhododendron coppice (7.2%).

V) FIGURE 14: SPECIES COMPOSITION IN PLOT 1 Cleared in 1984/5

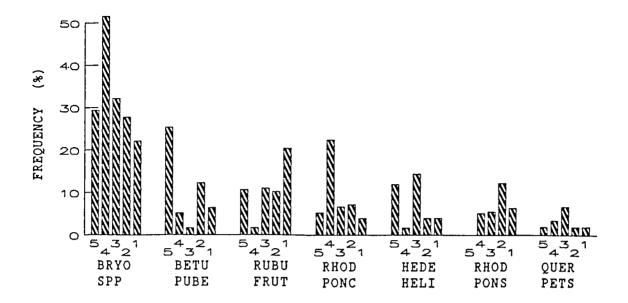


The plot is dominated by bryophytes (22.2%) and bramble (20.4%), although the major part of this bramble is found in the southern part of the plot which is devoid of trees (Map 3). Rhododendron seedling and birch both occur in 6.5% of the quadrats. The grasses and flowering plants are more frequent and diverse than in other years.

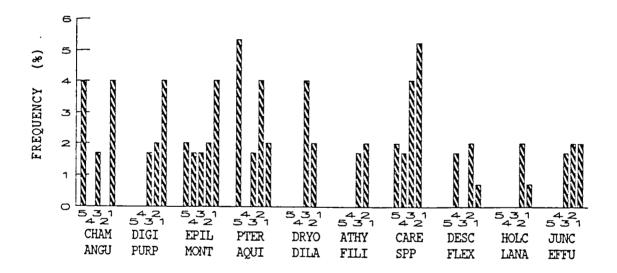
A summary of the species composition for each plot is shown at Figure 15 and Table 9.

FIGURE 15: SUMMARY OF SPECIES COMPOSITION BETWEEN PLOTS

i) Frequently occurring species



ii) Epilobiums, ferns and grasses



iii) Species occurring in 1 or 2 plots only

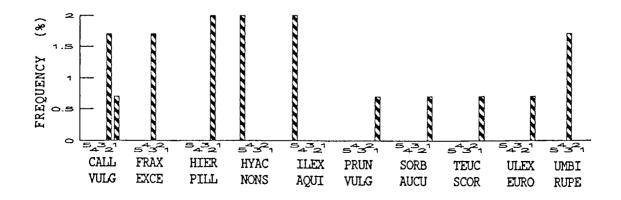


TABLE 9 : SUMMARY OF SPECIES COMPOSITION FOR EACH PLOT

ᆏ	1	!	1	spp frut	1	i	pube	angu purp mont heli ponc	vulg flex lana vulg aucu scor
	}	. !	1	Bryo Rubu	!	}	Betu Rhod	Cham Digi Epil Hede Rhod	Call Desc Holc Prun Sorb Teuc Ulex
•	!	}	ďďs	1	i	pube pons frut	spp	heli aqui acet	fili vulg flex purp dila mont pill lana effu
71	ļ	1	Вгуо	i	1	Betu Rhod Rubu	Care Rhod	Hede Pter Rume	Athy Call Desc Digi Dryo Epil Hier Holc Junc
<u>number</u> <u>3</u>	i	đđs	1	1	1	heli frut	ponc pons pets	spp dila	fili pube angu purp mont exce effu aqui
Plot nu	!	Bryo	}	!	ł	Hede	Rhod Rhod Quer	Care Dryo	Athy Betu Cham Digi Epil Frax Junc Pter
44∣	dds	!	{	poud	1	1	pube	pets pett	spp flex mont heli frut
71	Bryo	;	;	Rhod	-	1	Betu Rhod	Quer Quer	Care Desc Epil Hede Rubu
	!	i	spp pube	1	-	heli frut	aqui ponc	angu	spp mont nons aqui pets
ഥ	1	:	Bryo Betu	1	;	Hede Rubu	Pter Rhod	Cham	Care Epil Hyac Ilex Quer
ency	> 51	34.9	29.9	24.9	29.9	14.9	o.	4. Q	m >
* Frequency		30.0 -	25.0 -	20.0	15.0 -	10.9 -	ئ 0 . 1	3.0	·

6.2.5 Succession of Rhododendron and oak

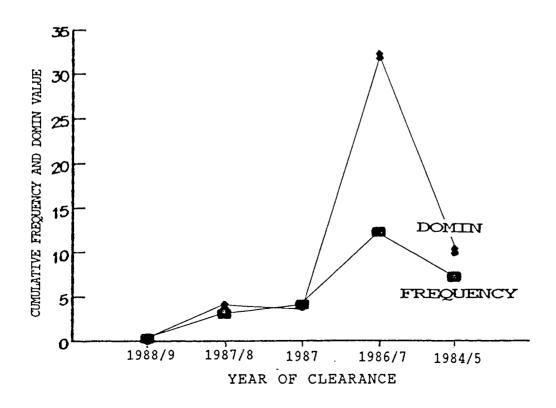
The succession of <u>Rhododendron</u> and oak seedlings is of particular importance, the former undesirable and the latter essential if the characteristics of the woodland are to be maintained in the long term. Statistical analysis of the succession of these species is however complicated by their low frequency of occurrence; in each plot the majority of the quadrats have neither seedling. This was less a reflection of quadrat size (1 x lm) than their low occurrence.

Simple analysis is possible by plotting the cumulative frequency and Domin value for each plot.

i) Rhododendron seedlings

Year of Clearance	Plot	Frequency	<u>Domin</u> <u>value</u>
1988/9	5	0	0
1987/8	4	3	4
1987 [°]	3	4	4
1986/7	2	12	32
1984/5	1	7	10

FIGURE 16 RHODODENDRON SEEDLING ESTABLISHMENT



The establishment of <u>Rhododendron</u> appears to conform with the literature reviewed (section 3.4). No seedlings are able to establish until a suitable site such as a thin bryophyte carpet appears. As time since clearance passes so the number of suitable sites increases. A point should be reached where the grass and herb sward prevents further <u>Rhododendron</u> seeds establishing, and for this particular site this point may have been reached at plot 1, where the lack of tree cover in the southern part of the plot has allowed a considerable grass and herb sward cover to develop. At this point however the <u>Rhododendron</u> seedlings that are already established may begin to grow through the sward despite its presence (section 3.5.1).

ii) Oak seedlings

<u>Year of</u> clearance	<u>Plot</u>	Frequency Number of quadrats	<u>Domin value</u>
1988/9	5	1	5
1987/8	4	2	2
1987	3	6	8
1986/7	2	2	3
1984/5	1	3	3

No significant trend in the establishment of oak seedlings is evident; indeed with the exception of plot 2 there is little variation in frequency, which is low in all plots. The stand composition (section 4.4) should ensure sufficient seed production for natural regeneration, and the low frequencies are not at present cause for concern, especially as establishing a trend over such a short time period is complicated by irregular mast years.

6.3 Germination trials

The germination trials were allowed to continue until the last week of August, by which time the seedtrays had been in the polytunnel for 15 weeks.

6.3.1 Germination of seed within the seedbank

TABLE 10 SPECIES RECORDED IN SEEDTRAYS 1-10

<u>Sample</u>	number	Seedlings recorded	<u>Number of</u> individuals
1988 1988	/ 5	Betula spp	1
1988	•	Betula spp	3
1988	/ /	Rubus fruticosus	3
1988	/ 8	Betula spp	2
1989	/ 1	<u>Betula</u> spp	2
1989	/ 2	<u>Betula</u> spp	13
		<u>Rubus fruticosus</u>	2
1989	/ 3	<u>Betula</u> spp	2
1989	/ 9	<u>Betula</u> spp	18
		<u>Rubus fruticosus</u>	1
1989	/ 10	<u>Betula</u> spp	5

Only two species, birch and bramble, germinated in these trials. Where the seedlings were not too numerous within an individual seedtray they were not removed but allowed to grow. The seedlings developed strongly, reaching up to 40cm in height whilst in the polytunnel; they showed no signs of inhibited growth.

The presence of seed in the soil seedbank depends on many factors, including seed production, dispersal, longevity, dormancy and viability, vulnerability to parasites, and possibly the time necessary to move through the litter layer to the mineral soil (Brown and Oosterhuis, 1981). The apparent lack of a diverse and abundant seedbank at Coed Bryn Mawr may therefore reflect a failure to meet the

requirements of one or more of these factors.

The lack of germination can be contrasted with other seedbank experiments. Abdy (pers.comm, 1989) undertook germination trials in the polytunnel at the same time as the Coed Bryn Mawr experiment, and using similar techniques obtained large numbers of seedlings from second rotation Sitka spruce sites at Clocaenog Forest in North Wales. Hill and Stevens (1981) recorded 13 species from a clearfelled Douglas fir plantation at Gwydyr forest in North Wales. The Douglas fir had been planted by the Forestry Commission in 1927, and at the time of felling there was virtually no ground cover. The 13 species recorded were all present in the 0-5cm soil layer.

Marks (1974) states that seed survives in quantity in soil for up to 50 years, after which there is a rapid decline in viability. Hill and Stevens (1981) substantiate this by referring to Peter (1893, 1894), who states that in plantations established for over 100 years, the flora of the previous land use is largely eliminated. That Rhododendron has dominated the understorey of the wood for perhaps over 100 years (no planting records exist), may have led to the exhaustion of viable seed in the seedbank, which would account for the germination of only two species.

Birch seed is short lived (Hill and Stevens, 1981), and although birch was one of the 13 species recorded by Hill and Stevens (1981) in Gwydyr forest, its presence was attributed to contamination during collection. Granstrom and Fries (1985) found that viability of Betula pubescens in forest soil declined to 6% of the original within one year, and to 1.5% of the original by the third year. The birch

that germinated from the soil collected at Coed Bryn Mawr is almost certainly from seed dispersed subsequent to the Rhododendron clearance.

Bramble is able to remain viable in the seedbank far longer than birch, although Marks (1974) states that substantial viability of the species may last only 50 years. The presence of bramble may therefore also be due to immigration of seed.

The <u>Rhododendron</u> may have been so impregnable that wind dispersal was inhibited, and the reduced fauna (particularly birds) using the wood since infestation would severely reduce potential dispersal by this source.

These inhibitions to the creation of a viable seedbank are in addition to any allelopathic inhibition to germination caused by the plant itself. Brown and Oosterhuis (1981) cite Howard and Ashton (1976) who state that whilst the processes of seed penetrating mineral soil are not well understood, burial by earthworms, ants and invertebrates is a more likely mechanism than the pure accumulation of litter and formation of a fermentation layer which may eventually become part of the mineral soil. Ground up Rhododendron leaves, when added to peaty soil, reduced the number of earthworms (Lumbricus rubellins) present (Doekson, in Cross, 1975). The same may be true of Coed Bryn Mawr, in which case another factor necessary in the formation of a viable seedbank is not satisfied.

It may be concluded that the colonisation of a cleared area will therefore not be from the seedbank, but by immigration. For seeds that are not wind dispersed, effective dispersal will be dependent on chance and may be slow, and

consequently not be capable of large scale colonisation. However wind-borne seed such as birch, rosebay willowherb, Epilobium spp and Compositae are capable of invading in appreciable numbers, and if conditions allow these will therefore be the species most likely to colonise cleared areas (Hill & Stevens, 1981).

6.3.2 <u>Test sowings</u>

TABLE 11 GERMINATION RATES IN SEEDTRAYS 1A-10A AND CONTROLS

1-5

<u>Sample</u>	number	<u>Number</u> <u>of</u> <u>Festuca</u>	<u>seedlings</u> <u>Betula</u>
1988 1988 1988 1988 1988	/ 5a / 6a / 7a	809 409 336 921 <u>920</u> Total 3395 Mean 679	324 128 0 213 290 955 191
1989 1989 1989 1989 1989	/ 2a / 3a	769 848 506 816 <u>539</u> Total 3478 Mean 696	166 70 67 189 <u>15</u> 507 101
Contr Contr Contr Contr	ol 2 ol 3 ol 4	833 889 939 468 <u>710</u> Total 3839 Mean 768	92 154 70 13 <u>115</u> 444 89

A one-way analysis of variance was undertaken for the test sowings in order to test for significant differences between the mean germination rates of the three treatments.

TABLE 12 ONE WAY ANALYSIS OF VARIANCE FOR TEST SOWINGS

I) Festuca germination

SOURCE	DF	SS	MS	F	P
FACTOR	2	22360	11180	0.24	0.793
ERROR	12	568887	47407		
TOTAL	14	591247			

II) Betula germination

SOURCE	DF	SS	MS	F	P
FACTOR	2	31053	15526	1.85	0.200
ERROR	12	100952	8413		
TOTAL	14	132005			

For both species there are no significant differences between the means of samples, and germination rates in this experiment cannot be attributed to year of clearance, or any latent inhibition to germination.

The mean germination of <u>Festuca</u> seedlings per half tray in all experiments was 716, a germination rate of approximately 33%. For the <u>Betula</u> seed this was approximately 127 seeds, or just under 6%. Although the germination of <u>Betula</u> was less than the average expected, namely 26% (Gordon and Rowe, 1982), as there was no significant difference between the mean germination rates for treatments, this cannot be attributed to any inhibiting factor within the soil. The germination of <u>Betula</u> seeds was more erratic than of the <u>Festuca</u>, varying from 0 to 324 per half tray.

Assuming that colonisation is dependent on immigration, then the ability of the soil to support germination and growth of the seed alighting on its surface is critical. From this experiment it appears that there is no significant difference in this ability between soil recently supporting dense Rhododenron thickets, and the test substrate of peat.

The conditions for germination created in the polytunnel are however not a replicate of conditions found in the wood itself. If this were the case the result of the germination trials (Table 10) would suggest that after a short time any cleared site would be substantially invaded by birch, with a presence of bramble.

The vegetation recorded in the quadrats shows that such an infestation has not occurred, and field conditions must therefore be different from conditions created in the polytunnel. These differences may include an allelopathic inhibition to growth, and differences in temperature, water and light.

Rotherham (1988) showed the inhibition of <u>Festuca ovina</u> shoot and root growth to be associated with a presence or recent presence of <u>Rhododendron</u>. However as Newman (1978) states, although an allelopathic substance may be present in a plant and released to the outside environment, it is difficult to know the concentrations of the active substance that reaches the receiver plant. Whilst the germination of seeds in a petri dish has therefore been shown to be restricted, the chemical change undergone by the inhibiting substance in the soil are complex and rapid (Newman, 1978), and the effects in the field may be very different.

7.0 CONCLUSIONS AND RECOMMENDATIONS

7.1 Vegetation associations

Over 50% of the site (88 quadrats) is dominated by bryophytes (60%), with birch, bramble and Rhododendron seedlings each present at less than 10%. If Rhododendron coppice and foxglove (the presence of foxglove is probably due to disturbance in plot 5) are included, over 68% of the quadrats are accounted for. The remaining 32% of the quadrats are contained in groups dominated by bramble, ivy or bracken. The diversity of associations within each plot varies from plot 1 (cleared in 1984/5) and plot 2 (cleared in 1986/7), the most diverse, to plots 3 (cleared in 1987), 4 (cleared in 1987/8) and 5 (cleared in 1988/9).

7.2 <u>Vegetation</u> and <u>light</u> intensity

There is a significant difference between the mean percentage crown cover of each TWINSPAN group. Groups with a high mean percentage crown cover are dominated by shade tolerant species such as bryophytes and ivy, whilst the groups with lower mean percentage crown cover contain more light demanding species such as bracken and foxglove.

7.3 Germination trials

Very little viable seed appears to be present in the seedbank. Although both birch and bramble germinated in soil collected from the site, their presence may be due to seed dispersed subsequent to clearance. In the absence of a viable seedbank colonisation of the site is dependent on immigration of seed; wind dispersed seeds will be able to colonise more rapidly than seed dispersed by other mechanisms. The conditions in the nursery are not replicated

in the field, and consequently an infestation by birch, with substantial bramble is not seen.

The germination in the polytunnel of birch and Festuca ovina seed was erratic but there was no significant difference between the ability of soil recently infested by Rhododendron, or the ability of peat, to support germination. Whilst allelopathy seems to account for inhibition of growth under laboratory conditions (Rotherham, 1988), no inhibiting agent appears to be present in the soil collected from Coed Bryn Mawr. A significant time lapse will usually occur between Rhododendron clearance and the dispersal and germination of seed; it may therefore be that any allelopathic agent has by this time undergone such changes whilst in the soil that any inhibition to growth is substantially reduced.

7.4 <u>Vegetation</u> <u>succession</u>

i) <u>Vegetation</u> and year of clearance

There is a significant difference between the frequency and abundance of vegetation in each plot. With the exception of the area cleared in 1987/8 (plot 4), both frequency and abundance increase with the number of years since clearance. The vegetation within each plot reflects the many factors of succession including the dispersal of seed and ability of the soil to support germination and growth. These factors are in turn affected by extraneous factors such as year of clearance and disturbance.

Bryophytes dominate each plot, with wind dispersed seeds accounting for an increasing diversity and abundance of grasses and flowering plants as the number of years since

clearance increases. The wind dispersed flora plays a dominant role from plot 3 (cleared in 1987), increasing in dominance in plots 2 and 1 (cleared in 1986/7 and 1984/5 respectively). By plot 1 a grass / herb sward has developed which prevents the establishment of Rhododendron seedlings and allows forest trees such as oak to establish. If the Rhododendron already established in this sward is suppressed (either by the vegetation itself or a control programme), clearance would appear successful. However the site being exposed to prolific seed sources will always be vulnerable to Rhododendron re-infestation, as disturbance creates fresh sites for establishment. Until more effective measures of initial control are found continued repeated eradication treatments will therefore be necessary.

ii) Rhododendron

The high frequency of <u>Rhododendron</u> coppice in plot 4 (cleared 1987/8) reflects the limited success of the initial clearance, and the need for subsequent treatments. The establishment of <u>Rhododendron</u> seedlings is shown by TWINSPAN to be strongly associated with the development of a bryophyte cover, and the site therefore provides plenty of opportunity for <u>Rhododendron</u> to re-invade, particularly when the wood is surrounded by seed sources. Despite the site being south facing (section 3.7) <u>Rhododendron</u> seedlings have a substantial presence in all plots with the exception of plot 5 (cleared in 1988/9) in which no suitable sites for germination have yet developed.

iii) Oak

The presence of oak shows no trend of increasing as the number of years since clearance increases. The present low frequency and abundance of oak is not yet cause for concern, as the stand composition should provide sufficient seed to ensure natural regeneration when site conditions are suitable.

7.5 Management implications

The aim of <u>Rhododendron</u> clearance within an oakwood of conservation interest is to end the domination of the understorey by <u>Rhododendron</u>, and to allow the native flora and fauna to re-colonise the wood. This aim is achieved as within five years of clearance the frequency and abundance of vegetation increases significantly; indeed within three years of clearance this increase is apparent.

The expenditure required for the initial clearance and for repeated control measures is substantial. Until more cost effective methods of control are developed, the availability of adequate funds to continue present control programmes will be essential.

7.6 <u>Further research</u>

Further succession studies are necessary to monitor the vegetation composition that results from complete removal of the understorey within an oakwood, and to assess the future inputs needed to prevent infestation reoccurring. Further research into the allelopathy that persists in the field, and into improved eradication and control measures are also needed.

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Appendix I PARASITIC FLORA AND FAUNA ASSOCIATED WITH RHODODENDRON PONTICUM.

I) Animal feeders or parasites

Hemiptera : Stephanitis rhododendri Horv

Dicyphus rhododendri Dolling

Graphocephala coccinea Forster

Dialeurodes chittendeni Laing

Masonaphis lambersi McGillivray

Coleoptera : Melolontha melolonthaL
Otiorhynchus sulcatus F
Otiorhynchus singularis L

II) Plant parasites

Armillaria mellea (Vahl ex Fr) Kummer

Botrytis cinerea Pers. ex Fr. Leaf blast
Cercospora handelii Bub. Leaf spot
Chrysomyxa rhododendri de Bary Root rot
Colletotrichumsp. Leaf spot
Cylindrocarpon destructans
(Zins) Scholten Root rot

Echidnodes aulographoides
(Bomm., Rouss & Sacc.) Robertson
Exobasidium vaccinii (Fuckel) Woron.
Lophodermium rhododendri Chev.
Mycosphaerella rhodoendri Lindau

Phomopsi ericaceana Fairm.

Phytophthora cactorum

(Lebert & Cohn) Schroet.

P.cinnamoni Rands

P.citricola Sawada

P.cryptogea Pethybr. & Laff

<u>Pycnostysanus azalaea</u> (Peck) Mason <u>Rhizoctonia solani</u> DC. ex Fr.

Leaf spot
Stem rot of
young trees

Stem rot
Wilt
Root rot and
ariel lesions
Etiolation

Twig blight Bud - gall

Leaf spot

Necrosis of scion leaves

and death

Bud blast

APPENDIX II BIRD SPECIES ABUNDANCE FOLLOWING THE

REMOVAL OF RHODODENDRON AT COOMBES VALLEY.

Source: Becker (1988)

I) New records

Acanthis flammea Accipiter nisus Certhia familiaris Columba oenas
Dendrocopos major
Dendrocopos minor
Muscicapa striata Parus ater Scolopax rusticola Sitta europaea Sylvia curruca

Redpoll
Sparrowhawk
Treecreeper
Stock dove
Great spotted woodpecker
Lesser spotted woodpecker
Spotted flycatcher
Coal tit

Woodcock

Nuthatch Lesser whitethroat

II) Species showing an increase

Ficedula hypoleuca Pied flyca
Fringilla coelebs Chaffinch
Parus caeruleus Blue tit
Parus major Great tit Parus major Phoenicurus phoenicurus Redstart
Phylloscopus sibilatrix Wood warbler

Pied flycatcher Great tit

III) Species showing a decrease

Garrulus glandariusJayPhylloscopus collybitaChiffchaffSylvia atricapillaBlackcapTurdus merulaBlackbirdTurdus philomelosSong thrush

IV) Species showing no change

Aegithalos caudatus Long talled Columba palumbus Woodpigeon Carrion crow Erithacus rubecula Parus palustris
Picus viridis
Prunella modularis Sitta europaea Strix aluco Sturnus vulgaris
Sylvia borin
Troglodytes troglodytes
Turdus viscivorus

Starling
Garden warbler
Wren
Mistle thrush

Long tailed tit Robin Marsh tit
Green woodpecker
Dunnock
Nuthatch
Tawny owl
Starling Garden warbler

APPENDIX III ASSESSMENT OF THE QUANTITY OF SEED PER SEEDTRAY

I) The seed in eight petri dishes, each containing 0.5 grams of Betula pubescens seed was counted.

<u>Petri</u> <u>dish</u>	<u>Nur</u>	nber of s	eeds (X-X)	$(x-\overline{x})^2$
1		2199	42	1764
2		2305	148	21904
3		1981	- 176	30976
4		2227	70	4900
5		2197	40	1600
6		2000	-157	24649
7		2109	-48	2304
8		<u> 2238</u>	81	<u>6561</u>
	Total	17256		94478
	X =	2157		

This adequately satisfies the required number of counts needed to produce an accurate mean, according to the formulae:-

$$N = t^2 cv^2$$

N = The number of petri dishes required to be counted

t = T table factor at N - 1 degrees of freedom

cv = Coefficient of variation

e = Percentage accuracy reuired

Adopting a 95% probability level, and 5% accuracy required, eight counts were shown to be sufficient :-

$$N = \frac{(1.895^2) * (5.04^2)}{5^2} = 3.65$$

II) Five samples of 2157 <u>Festuca</u> seed were counted into separate petri dishes, and an average weight of seed derived. The remaining quantities of seed were weighed accordingly.

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APPENDIX IX SPECIES LIST AND ABBREVIATED NAMES

BOTANICAL NAME	<u>VESPAN</u> CODE	COMMON NAME
TREES		
Betula pubescens	Betu pube	White birch
Crataegus monogyna		Hawthorn
ILex aquifolium	Ilex aqui	Holly
Pseudotsuga menziessi		Douglas fir
Quercus petraea (Seedling) (stump)	Quer pets Quer pett	Sessile oak
Salix caprea		Goat willow
Sorbus aucuparia	Sorb aucu	Rowan
Tilia x europaea	Tili cord	Common lime
FLOWERING PLANTS		
Buddleja davidii		Buddlea
Calluna vulgaris	Call vull	Heather
Chamaenerion angustifolium	Cham angu	<i>Rosebay</i> Willowherb
Circaea lutetiana		Enchanters Nightshade
Digitalis purpurea	Digi purp	Foxglove
Epilobium montanum	Epil mont	Broadleaved Willowherb
Fragaria vesca		Wild strawberry
Geranium robertianum		Herb Robert
Hedera helix	Hede heli	Ivy
Hieracium pilosella		Hawkweed
Hyacinthoides nonscriptus	Hyac nons	Bluebell
Hypericum androsaemum		Tutsan
Hypochoeris glabra		Smooth cats-ear

BOTANICAL NAME	VESPAN CODE	COMMON NAME
Lysimachia nemorum		Yellow Pimpernel
Oenothera biennis		Common evening primrose
Oxalis acetosella		Wood sorrel
Potentilla sterilis		Barren strawberry
Prunella Vulgaris	Prun vulg	Selfheal
Ranunculus spp		Buttercup
Rhododendron ponticum (Seedling) (Coppice)	Rhod pons Rhod ponc	Rhododendron
Rubus fruticosus	Rubu frut	Bramble
Rubus idaeus		Raspberry
Rumex acetosella	Rume acet	Sheeps sorrel
Senecio jacobaea		Common Ragwort
Solidago virgaurea		Goldenrod
Taraxacum officinale		Dandelion
Teucrium scorodonia	Teuc scor	Woodsage
Umbilicus rupestris	Umbi rupe	Wall pennywort
Veronica officinalis		Heath speedwell
Ulex europaeus	Ulex euro	Gorse
<u>GRASSES</u>		
Agrostis capillaris	Agro capi	Common bent
Anthoxanthum odoratum		Sweet vernal grass
Carex binervis		Green ribbed sedge
Carex demissa		Common yellow sedge
Carex pallescens		Pale sedge
Deschampsia flexuosa	Desc flex	Wavy hair grass
Holcus lanatus	Holc lana	Yorkshire fog
Juncus articulatus		Jointed rush
Juncus effusus	Junc effu	Soft rush

BOTANICAL NAME VESPAN CODE COMMON NAME

BRYOPHYTES

Atrichum undulatum

Bryophyte spp Bryo spp

Campylopus introflexus

Dicranum majus

Funaria hygrometrica

Hypnum cupressiforme

Isothecium myosuroides

Leucobryum glaucum Cushion moss

Mnium hornum

Polytrichum spp

PTERIDOPHYTES

Athyrium filix - femina Athy fili Lady fern

Dryopteris dilatata Dryo dili Broad buckler fern

Dryopteris

filix - mas Dryo fili Male fern

Pteridium aquilinum Pter acqu Bracken

APPENDIX X SCHEFFEY'S TEST - following a one way analysis of variance to assess the relationship between TWINSPAN groups and light intensity.

ROW	TWINSPA	N GROUP	DIFFMEAN	SE	rom_ci	UPP_CI
1	*000	*001	-21.20	20.3502	-98.606	56.2061
2	*000	*010	-30.88	21.8232	-113.889	52.1288
3	*000	*011	-11.92	19.7398	-87.004	63.1643
4	*000	*100	-44.86	20.8669	-124.232	34.5116
5	*000	*101	-14.50	20.5751	-92.761	63.7615
6	*000	*110	-50.08	21.0832	-130.274	30.1142
7	*000	*111	-35.75	23.9061	-126.682	55.1816
8	*001	*010	-9.68	11.3305	- 52 . 778	33.4179
9	*001	*011	9.28	6.4645	-15.309	33.8690
10	*001	*100	-23.66	9.3573	-59.252	11.9324
11	*001	*101	6.70	8.6870	-26.343	39.7427
12	*001	*110	-28.88	9.8301	-66.271	8.5107
13	*001	*111	-14.55	14.9543	-71.432	42.3317
14	*010	*011	18.96	10.1936	-19.813	57.7334
15	*010	*100	-13.98	12.2343	-60.516	32.5557.
16	*010	*101	16.38	11.7296	-28.236	60.9959
17	*010	*110	-19.20	12.5996	- 67 . 125	28.7252
18	*010	*111	-4.87	16.9041	-69.168	59.4284
19	*011	*10	-32.94	7.9428	-63.152	-2.7280
20	*011	*101	-2.58	7.1409	-29.742	24.5819
21	*011	*110	-38.16	8.4947	-70.471	-5.8489
22	*011	*111	-23.83	14.1124	- 77 . 509	29.8493
23	*100	*101	30.36	9.8368	-7.056	67.7761
24	*100	*110	-5.22	10.8595	-46.526	36.0863
25	*100	*111	9.11	15.6502	-50.419	68.6387
26	*101	*110	-35.58	10.2875	-74.711	3.5507