Methods for predicting Sitka spruce natural regeneration presence and density in the UK

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Running title

Predicting Sitka spruce regeneration

Abstract

Natural regeneration is crucial for silvicultural approaches based on the continuous presence of a forest cover, or Continuous Cover Forestry (CCF). Sitka spruce (*Picea sitchensis*), is the main commercial species in the United Kingdom (UK), and its potential for CCF has been demonstrated in various studies. However, there are no quantitative models available to predict its natural regeneration in the country. We describe models for Sitka spruce seedlings presence and density under canopy cover in the UK forests, to be used as a substitution of a regeneration survey.

Using a natural regeneration dataset comprised of 340 plots, a Generalized Linear Mixed Model (GLMM) was calibrated to estimate the likelihood of regeneration presence at plot level. Seedling density was simulated in a subsequent step using only the subset of data with regeneration presence (138 plots): we compared methods based on GLMMs calibrated to the observed seedling density, and the simple generation of random numbers similar in distribution to the observed values. We validated the models with a cross-validation method using the calibration dataset, and with an independent dataset of 78 plots collected in forests already in the process of transformation to CCF.

The best GLMM for regeneration presence included age of the plantation, time after last thinning, favourable ground cover and basal area. After the cross-validation, 73% of the plots were correctly estimated (76% for presence of regeneration and 71% for the absence). After the independent validation process, 82% of the plots were correctly estimated, although 100% for presence of regeneration and only 12% for the absence. Both methods for estimating seedling density had a poor performance, both with the cross-validation and independent validation.

The results showed that the tools here described are appropriate for estimating regeneration presence in traditional Sitka spruce plantations. However, alternative methods are required for forests already in an advanced stage of transformation to CCF systems.

Introduction

Continuous cover forestry (CCF) is a range of silvicultural approaches involving uninterrupted maintenance of forest cover and avoidance of clearcutting (Pommerening & Murphy 2004), is becoming increasingly important worldwide (Schütz et al. 2011). Under this approach, there is a focus on the use of natural regeneration to develop uneven-aged and mixed-species stands (Pommerening & Murphy 2004). Various models have been thus developed to predict the occurrence of such regeneration. The process was often split into two stages: i) determining if regeneration is successfully occurring during the time interval studied, and if so ii) defining the species composition and density of the established seedlings (Miina et al. 2006). For the first stage, logistic equations with binomial distribution are often calibrated on various stand and site characteristics to estimate the probability of regeneration occurrence in a forest plot, considered to have a binary status of absence or presence (Ferguson & Carlson 1991, Hasenauer & Kindermann 2006, Pausas et al. 2006, Schweiger & Sterba 1997). Stochastic approaches are common since forest regeneration, particularly in boreal and temperate regions, tends to be sporadic (Miina et al. 2006). Then, the species composition and density are defined using different statistical approaches, often based on the Weibull or Poisson distribution.

Sitka spruce (*Picea sitchensis* Bong. Carr.) is a prolific seed producer with abundant natural regeneration after clear-cutting both in its natural range (Peterson et al. 1997) and in the UK, where it is the commercial conifer with the highest potential for natural regeneration:

up to 400,000 seedlings per ha on favourable sites after clearcutting or wind-throw, although with high variation between and within sites (Nixon & Worrell 1999). Various reviews of the factors influencing the natural regeneration of Sitka spruce in the British Isles have been carried out, focusing on obtaining natural regeneration as a substitute for artificial planting in clear-felled areas (Clarke 1992, von Ow et al. 1996, Nixon & Worrell 1999). Sitka spruce also proved to have the potential for regeneration under canopy cover in the UK, and more recent studies researched how to obtain and use natural regeneration to transform even-aged, mono-specific conifer forests into irregular stands (Malcolm et al. 2001, Mason & Kerr 2004). Mason (2015) recently carried out an exhaustive review especially focused on Sitka spruce natural regeneration under canopy cover in the UK, and summarized the main factors involved (Table 1).

Seed availability is undoubtedly the first crucial factor: Sitka spruce seeds, like those of most temperate forest tree species, have a low survival rate in the forest soil and do not produce a viable seed bank. Sitka spruce in the UK starts to have a good seed crop at 25-35 years, after which the seed production increases with age and can reach high levels already at 35-40 years, depending also on the stand density (Nixon & Worrell 1999). Years of heavy cone production tend to be synchronised amongst trees and to happen at periodic intervals called mast years, that in the UK can happen every 3-6 years (Clarke 1992, Mason 2015).

Seed germination is highly dependent on the seedbed characteristics. Nixon & Worrell (1999) indicated as the most favourable seedbed soils with low fertility (because of less competing vegetation), with the presence of adequate moisture (neither too dry nor too wet), and without too much brash or needle litter (considered unfavourable due to low water retention). On the contrary, von Ow et al. (1996) found litter favourable to germination in Ireland. Low-growing mosses are generally considered favourable for regeneration (Mason 2015) due to good water retention, while taller mosses seemed to have a negative effect likely because they prevent the seedlings' root from reaching the mineral soil (von Ow et al. 1996). In the coastal forests of North America, decayed logs are considered the most favourable seedbed for Sitka spruce seedlings (Harmon & Franklin 1989, Taylor 1990).

Stand structure can affect regeneration through different mechanisms. A certain level of overstorey cover was found to be beneficial for Sitka spruce regeneration both in its natural range (Burton 2016, Greene et al. 1999) and in the UK (Mason et al. 2004), likely thanks to the control of the growth of competing ground vegetation (Nixon & Worrell 1999), and the influence on the microsite temperature and moisture (Fairbairn & Neustein 1970). On the other hand, the presence of overstorey trees reduces the light availability for seedlings. Light-growth functions for the growth of Sitka spruce seedlings in the UK have been developed by Bianchi et al. (2018).

Thinning interventions have been shown to have a positive effect by creating a favourable light environment. Studies carried out both in the UK and in North America generally found more Sitka spruce seedlings in the stands with lower densities, which were either more recently or more heavily thinned (Deal & Farr 1994, Page et al. 2001, Herd 2003, Glendinning 2014), but differences were not always observed between silvicultural treatments (Bertin et al. 2011). Most of the studies in the UK considered stands originated from artificial planting. Regeneration density in such pure Sitka spruce stands after thinning varied from 4,500 to 70,000 seedlings per ha, but when small germinants under 20 cm height were considered the density could go up 270,000 seedlings per ha (Page et al. 2001, Herd 2003, Bertin et al. 2011, Glendinning 2014). In contrast, studies in North America focused on natural mixtures of Sitka spruce and western hemlock, and Sitka spruce regeneration occurred with lower densities (1,900-22,000 seedlings per ha) (Deal & Farr 1994). When comparing local overstorey variables to seedling density, contrasting results were found: either no relationship was observed (Glendinning 2014), or only a weak positive correlation with basal area (Page et al. 2001), or a weak negative correlation with stems per ha (Deal & Farr 1994).

Objectives

The UK has been defined "data-poor" regarding natural regeneration (Kerr et al. 2011), and even if the qualitative information is extensive, there are no existing models to quantitatively predict the regeneration occurrence of Sitka spruce under canopy cover. The aim of this research was to prepare such models by investigating as main predictors the factors considered more affecting such processes. We also put emphasis on analysing the methodological approaches available given the constraints of the UK situation.

In the absence of studies following the development of regeneration over time, the dataset generated by Kerr et al. (2011) is the most comprehensive regeneration survey of coniferous forests available in the UK to date, covering a wide range of forest structures and geographical areas. We thus decided to use this dataset for calibration. However, there were some limitations. The dataset was produced by a one-off sampling, including neither detailed information on the timing of the regeneration establishment nor on its size. The age of the regenerating trees could have been highly variable, and so could the biological processes they had been through, and/or the stand characteristics at the regeneration event could have been very different from the survey data. The only possible approach using such a dataset was to model the regeneration "presence", and not the regeneration "occurrence", the latter defined as the seedling establishment within a time interval. We thus calibrated models that could generate a regeneration tally like one produced from a field survey, for stands which do not have this information. First, we modelled the likelihood of Sitka spruce seedling presence, then its density. For each stage we identified the significant variables within the wide range of those included in the original survey. We considered plots as modelling units to allow the predictions to be sensitive to within-stand variations, as recommended by Miina et al. (2006). The models prepared were then validated with an independent dataset.

Methodology

Calibration dataset

Kerr et al. (2011) carried out multi-level sampling during 2008/09 in 129 stands of coniferous species located in 38 forests across most of Great Britain. From this, we extracted information on 34 artificially–planted, Sitka-spruce-dominated stands, located in 13 forests evenly distributed across most areas of Great Britain where Sitka spruce is present (see original research for more details). In the original survey, ten 0.01 ha circular plots (radius 5.6 m) were laid out in each stand, recording diameter at breast height (DBH, measured at 1.30 m above ground) and species for all trees more than 7 cm DBH. In a 2 m x 2 m square located at the centre of the circular plot, the number and species of all trees less than 7 cm DBH were recorded, differentiating between seedlings (height less than 1.30 m) and saplings (height more than 1.30 m). From the 340 plots retrieved, 138 showed at least one Sitka spruce seedling or sapling (40% of the total). We considered those plots to have presence of regeneration. Since saplings occurred in only four plots, in which seedlings were also present, we decided not to differentiate between them. From now on, we will refer to all regenerating trees as seedlings. The main characteristics of the calibration dataset are indicated in Table 2.

Age of the plantation in years (from now on simply Age), Soil Nutrient Regime (SNR), time after last thinning, and Deer Impact Index (DII) were recorded at stand level. We calculated from the original inventory the plot level values for basal area (BA), stems per ha (SPH), and the maximum DBH (maxDBH). From those values we calculated at plot level the quadratic mean diameter (QMD, the diameter of a tree considered as having the average basal area); and the Global Site Factor (GSF) , an indication of the canopy light transmittance, using the relationship established from (Hale et al., 2009).

As an indication of seed availability, we investigated the use of Age and two possible alternatives. Hasenauer & Kindermann (2006) for MOSES used maxDBH (at plot level) to represent a mother-tree effect, while Schweiger & Sterba (1997) used QMD as a substitute for age; both were positively correlated with regeneration occurrence in mixed-species, uneven-aged forests. However, in this dataset both maxDBH and QMD were negatively correlated with regeneration presence (preliminary results not shown). For this reason, maxDBH was considered as a possible indicator of local overstorey competition (see later) while QMD was discarded.

The SNR was estimated by the original field surveyor from analysis of the ground vegetation following the Ecological Site Classification criteria (Pyatt et al. 2001). Most of the stands were located on sites with either medium or poor SNR (respectively 38% and 53% of the total plots). Those two classes did not show a significant difference from each other in terms of regeneration presence frequency (Fisher's exact test, two-sided: $p=0.556$, $n=310$), and only 9% of the plots were in other SNR classes, so we excluded this factor from further analysis. The SNR class indirectly influences regeneration due to its effect on ground vegetation, as described previously. Since the dataset included for the 2 m x 2 m plots the percentage of ground covered by different classes of vegetation, we decided to use as candidate variables the favourable ground cover classes of Mosses and Bare Ground, instead of SNR, consistent with the model prepared by Kerr et al. (2012).

We considered the plot-level stand density measures of BA, SPH and maxDBH as a negative proxy for the light regime under the forest cover (higher stand density, lower light level) and so expected to be negatively correlated with regeneration presence. On the other hand, GSF is a direct indication the light regime under the forest cover, expected to be positively correlated with regeneration presence. The time since the last thinning was estimated for each stand using both historical records and evidence on the ground; the expected effect was a negative correlation between the time since the intervention and the likelihood of regeneration. We divided the stands in the present study into three different Thinning Classes (TC) as in Kerr et al. (2011): TC 1, thinned in the last 1-5 years; TC 2, thinned 6-10 years before; TC 3, thinned more than 10 years before or never. We used discrete classes since there was often an uncertainty in the precise timing of the thinning. In some cases, it was observed that a thinning was carried out only in a fraction of the stand. Since we could not identify which specific plots were affected, we assigned an approximate thinning class to the whole stand with a subjective decision (for example, when only half of the stand was reported to be affected by a recent thinning as in TC1, and the rest by none, a TC2 was assigned to all the plots). We considered this variable as numeric.

The Deer Impact Index (DII) was visually estimated as low (no browsing observed), moderate (browsing damage on up to 25% of the regeneration) and high (browsing damage on more than 25% of the regeneration). Because of the unbalanced distribution (see Table 1) and the lack of significant differences in regeneration presence frequency between the moderate and high impact classes (Fisher's exact test, two-sided: p=0.611, n=330), this factor was discarded from further analysis.

Additionally, we retrieved stand-level geographical variables from topographic maps, namely northing, easting, elevation and aspect, and stand-level climatic variables from the Forestry Commission's decision support system ESC-DSS (Pyatt et al. 2001), namely accumulated temperature above 5 ˚C, moisture deficit, Conrad continentality index and total summer and winter rainfall. Preliminary analysis (not shown) revealed that none of those variables was significant when included in a model and they were all discarded.

The density of Sitka spruce seedlings per plot was very different between the Thinning Classes (Figure 1). Sitka spruce contributed to 97% of the seedlings in the study areas and different species were sporadic (present in only 2% of the plots); for simplicity the latter was ignored during the analysis. At stand level, considering all the plots with or without regeneration, there were on average of 20,740 seedlings per ha, with a minimum of 0 and a maximum of 250,000.

Independent validation dataset

For independent validation, we assessed in 2016 four Sitka-spruce-dominated stands in Clocaenog forest, Denbighshire, Wales (53° 04′ N, 3° 25′ W, 390-430 m altitude), and four in Kielder forest, Northumberland, England (55° 10′ N, 2° 29′ W, 200-250 m altitude). Both forests were originally artificial plantations that have been managed in recent years according to different CCF principles, using silvicultural systems ranging from irregular shelterwood to group selection. All stands belonged to Thinning Class 2, but most of them were thinned more frequently or with higher intensity in the past than stands in the calibration dataset. The situation in all stands was generally a lower tree density than under the traditional management (as defined by Edwards & Christie 1981), leading to a larger amount of natural regeneration. For each stand, we drew random non-parallel transects on a desktop map and placed on them 10 evenly spaced plots, later located in the field using a GPS receiver. The distance between plots varied with the size of the stand. We followed the same data collection protocol used for the calibration dataset and collected in this way 78 plots. The main characteristics of this dataset are shown in Table 3 for a comparison with the calibration dataset. SNR, DII and QMD were not considered, as in the calibration dataset. Again, we considered all seedlings and saplings as “seedlings”, and a total of 62 plots (about 80% of the total) had at least one of these. The density of Sitka spruce seedlings per plot is shown inFigure Figure 1. At stand level, considering all the plots with or without regeneration, there were on average 46,940 seedlings per ha, with a minimum of 4,500 and a maximum of 171,800.

Statistical analysis

Regeneration presence

We carried out all the analyses using R Statistical Software (R Core Team 2017). To estimate the probability of regeneration presence, we used a Generalized Linear Mixed Model (GLMM) fit by maximum likelihood (Laplace Approximation) with Binomial function and Logit link, from the package lme4 (Bates et al. 2014). Possible autocorrelation effects were considered using the stand and forest levels as random nested effects. The candidate fixed effects for the model were Age, BA, SPH, GSF, maxDBH, Thinning Class, Mosses and Bare Ground. We included a quadratic term for BA, SPH, and maxDBH to check if the relationship between stand density and regeneration was non-linear, as a certain level of canopy cover can be beneficial to natural regeneration. Then we removed non-significant parameters using a step-wise approach aimed at reducing the Akaike Information Criterion (AIC) to select the best model (Yamashita et al. 2007). We re-calibrated the best model structure on standardized variables (rescaled so that their new mean is equal to zero and the standard deviation to 1). This process transforms all the variables with different orders of magnitude to a similar scale, still maintaining their variability, making the magnitude of the model coefficients directly comparable.

We assessed the accuracy of the best model with a cross-validation technique (Bennett et al. 2013). Using the same model structure, we re-calibrated the coefficients by removing all the plots belonging to one stand from the calibration dataset. Then we validated it on the plots belonging to the left-out stand and calculated their likelihood of regeneration presence. We repeated the process 34 times, once for each stand. After we estimated in such a way the likelihood of regeneration for each plot, to determine which ones the model would predict to have regeneration, we used two methods.

In the first method, we defined a cut-off likelihood value using the Receiver Operator Characteristics (ROC) curve method with the package pROC (Robin et al. 2011). We assigned the presence of regeneration to all plots with a likelihood above the cut-off, and otherwise the absence of regeneration. We estimated this cut-off as the likelihood value that would maximise the sum of sensitivity (the proportion of correctly identified positive plots, that is in this case with presence of regeneration) and specificity (the proportion of correctly identified negatives, that is with absence of regeneration). Once each plot was assigned its simulated status, we built a contingency table to compare the predictions with the observations. In the second method, we used a stochastic approach (Hasenauer & Kindermann 2006). We generated for each plot a pseudo-random number between 0 and 1. If that number was lower than the regeneration likelihood, the plot was considered to have regeneration, and otherwise without regeneration. We ran the simulation 10,000 times, averaged the results, and built another contingency table. For both methods, we analysed the results also at stand level in the following way. For each stand, we calculated the difference between the total of all simulated regeneration plots minus the total observed ones. We checked the field notes to subjectively investigate why predictions were in error for the stands with the worst results (as in Ferguson et al. 1986). For this analysis, we did not consider it important if individual plots were wrongly simulated if the overall predictions at stand level were accurate.

Regeneration density

We used two approaches. First, we investigated GLMMs using the same random and fixed effects as described above, using the sub-dataset for plots with presence of regeneration (n = 138), and a Gamma distribution with log-link to approximate the seedling distribution. No preliminary model based on all plots with presence of regeneration (n = 138) could converge (results not shown). The importance of the Thinning Class was evident from the sharp difference in seedling distribution amongst the classes, so we decided to calibrate separate models for TC1 and TC 2 & 3 (pooled together due to the lower number of observations). For those two subsets of data, we prepared GLMMs using the same random and fixed effects as described above (excluding Thinning Class). Then we removed non-significant parameters using a step-wise approach aimed at reducing the AIC to select the best model.

We evaluated its accuracy through comparing predicted and observed values at plot and stand level.

For the second approach, we simulated the seedling density simply by generating random numbers that approximated the observed density distribution for each Thinning Class (Ferguson & Carlson 1993, Schweiger & Sterba 1997). We fitted Weibull distribution functions to simulate the distribution pattern of the seedlings in each Thinning Class group using the package MASS (Venables & Ripley 2002). We used the values of seedlings per ha observed at plot level transformed to units of 1,000 for simplifying the calculations. For validation, in each plot observed with regeneration, we generated a random number 10,000 times from the resulting functions and averaged the results. I then compared the observations and simulations averaged at stand level. I did not compare results at plot level analysis since the random generation of numbers makes this analysis impossible.

Independent validation

We calculated the likelihood of regeneration presence in the independent validation plots using the best model above selected (calibrated on the full dataset). Then, we used the same two methods as before to assign the presence of regeneration. First, we considered the same cut-off likelihood value previously determined with the ROC method, assigning the status of presence of regeneration to all plots above that threshold. Second, we used the stochastic method to randomly determine the presence or absence of regeneration. For both methods, we built contingency tables at plot level and examined the performance at stand level by comparing the total numbers of simulated and observed plots with regeneration, with the same procedures described above for the cross-validation. Then, we used both seedling density modelling methods prepared with the calibration dataset to simulate the density in the plots of the independent datasets with observed presence of regeneration. The simulated seedling density was compared with the observed values.

Results

Regeneration presence

The model structure after the step-wise AIC reduction process is shown in Model (1), with more details of the coefficients shown in **Error! Reference source not found.**.

Model (1): $p_{\text{regen}} = \frac{1}{1+e^{2.693+1.864*TC-0.087*Age-0.020*Mosses+1.569*(BA/100)^{\wedge}2}}$

The model did not converge when the forest-level random effect was included, so we maintained only the stand-level effect. The effect of bare ground was not significant, and it had a weak negative relationship with regeneration, contrary to the hypothesis. Only the quadratic term for BA remained in the best model structure amongst the stand density indicators. Note that values of BA were divided by 100 since they were on a different scale from the other variables.

Figure 2 displays how the probability of regeneration changes according to variation in the model variables. Using Model (1), we calculated the likelihood of regeneration presence for new virtual datasets. In Figure 2a, we used a dataset where we allowed only TC to vary (from 1 to 3) while the other fixed effects were kept at the mean values of the calibration dataset (as seen in Table 4). In Figures 2b, 2c and 2d, we allowed respectively Age, BA and Mosses to vary across the full range observed in the calibration dataset, while we kept the other fixed effects at their means except for TC. We repeated the analysis changing the Thinning Class, represented by the different lines (decreasing from 1 to 3 from top to bottom). Generally, from TC 1 to 2 there was a stronger decrease in regeneration likelihood than from TC 2 to 3. For TC 1, regeneration probability decreases more sharply for Age less than 60 years and BA more than 60 m^2 ha^{-1}. For TC 2, only in old stands (more than 70 years old) was the probability of regeneration above 0.5, while for TC 3 the likelihood was always low. The effect of mosses on regeneration likelihood was more linear.

Figure 3 shows the coefficient values for the model shown in Equation (1) when it was calibrated on the standardized variables. TC had the highest coefficient (i.e. most influential) in absolute terms (1.522), followed by Age (1.255), Mosses (0.701) and BA (0.533).

After the cross-validation analysis, with the ROC method, the cut-off likelihood value for the regeneration presence probability was 0.3. Figure 4 shows the ROC curve, that is all the combinations of specificity and sensitivity values obtained by using all the possible cut-off values. The chosen cut-off was the one that maximised their sum and corresponded to the point on the curve closest to the upper left corner, which would be to the ideal case of both specificity and sensitivity equal to 1. For the ROC method, the plots that had an estimated likelihood above 0.3 were considered by the model to have presence of regeneration. For the stochastic method, the pseudo-random generated numbers were checked with the likelihood values for each plot. Table 4 shows the contingency table of using both methods. For the ROC method, the plots correctly predicted (true positives plus true negatives) amounted to 73% of the total. The model estimated with similar accuracy plots with or without presence of regeneration (respectively 76% and 71%). For the stochastic method, there was a markedly lower accuracy in sensitivity (55%) and only a slightly better specificity (74%), bringing the overall accuracy lower than in the ROC method (66%).

When the results were aggregated at stand level for the ROC method, 21 stands out of 34 had a difference between total observed and predicted regeneration plots equal to or lower than 20% (11 with no difference), while five had a difference equal to or larger than 50% (worse than chance). For the stochastic method, very similar results were obtained: 22 stands out of 34 had a difference between total observed and predicted regeneration plots equal to or lower than 20% (10 with no difference), while five had a difference equal to or larger than 50%.

The worst simulated stands were almost the same stands in both methods. The field notes provided additional insights about them, showing that they were generally the ones subjected to heterogeneous thinning interventions within the same stand, suggesting that the TC class was inaccurate. In stand with fewer simulated regenerating plots than observed, it was also observed that windblow events had opened gaps comparable to a thinning, or that there was precocious cone production in young stands. In stands with more simulated regenerating plots than observed, it was noted that in stands favourable for regeneration according to all the model variables, the limiting factors were likely to be: competing ground vegetation; presence of deer browsing; and lack of cone production. In the two worst over-simulated stands for both methods, the field notes declared that everything seemed suitable for regeneration and its total absence was inexplicable for the surveyor too.

Regeneration density

In the GLMMs calibrated for TC 1 and TC 2 & 3, only the effect of BA was significant, but with a positive relationship with seedling density in the former class (TC 1) and a negative relationship for the latter group (TC 2 & 3). However, both models showed a very poor fit between the simulated and observed density values and they were discarded (results not shown).

The Weibull distributions fitted to seedling density distribution in each TC are described by the parameters in Table 5. Figure 5 shows the comparison between the distribution of simulated values of seedlings per ha and the distribution of the observed values, considering all plots with regeneration. While the fit was adequate at whole-population level for each Thinning Class, at stand level it did not provide good results. Generally, there was a poor correspondence between those values: only two stands had a simulated density ± 20% of the observed density. On average, the difference between simulated and observed values was 770 seedlings ha^{-1}, but with extremes of -177,500 and 59,000 seedlings ha^{-1}.

Independent validation

We used Model (1) to calculate the likelihood of regeneration presence in the independent dataset. With the ROC method, we considered regeneration to be present only in the plots with a likelihood greater than the same cut-off likelihood value of the cross-validation process ($p = 0.3$). The resulting contingency matrix is shown in Table 4, together with the results of the stochastic method. For the ROC method, while the total accuracy was 82%, this was because almost all plots (76 out of 78) were predicted to have regeneration, giving a sensitivity of 100% and a specificity of only 12%. For the stochastic method, the overall accuracy was again lower than for the ROC method (64%), although sensitivity and specificity were more even. After aggregating the results at stand level, however, worse results were found for the ROC method than for the stochastic method: out of eight stands, respectively four for the ROC method and six for the stochastic method had a difference between total observed and predicted plot with regeneration equal to or lower than 20%. In both methods, two stands had no difference between total observed and predicted plots with regeneration, and none had a difference equal to or larger than 50%.

Regeneration density was then estimated in the plots with observed regeneration presence (n=62). Only the Weibull distribution approach was used, with the function previously calibrated for Thinning Class 2. The GLMM approach was already deemed too inaccurate. After averaging the results, there was no good correspondence between the simulated and observed values, and no stands had a simulated density ± 20% of the observed value. On average, the difference between simulated and observed values was -34,570 seedlings ha^{-1}, with extremes of -155,800 and 4,500 seedlings ha^{-1}.

Discussion

The model predicting regeneration presence was based on the established knowledge of the biological and ecological characteristics of Sitka spruce. The effect of time since the last thinning showed the strongest significance in the model, and the largest coefficient after standardization. Consistently with Kerr et al*.* (2012), the model showed that probability of regeneration presence is high after an intervention, but it decreases rapidly and there is no positive effect after 10 years. If the operations are not repeated, the canopy can revert quickly to a closed status and small seedlings die off (Hale 2003). The field notes showed that inaccuracies in the thinning regime information, or the presence of windblown gaps not considered in the model, were likely causes of the errors in the worst-simulated stands. To improve the accuracy, it is necessary for the model to know which plots are affected by a tree removal, irrespective of whether it is due to natural mortality or timber extraction.

The age of the plantation emerged as the second most important factor. Such a positive effect in the artificial plantations of the present study can be explained by the larger seed production of older trees, and possibly also by the higher number of gaps that can naturally occur in a mature canopy past the self-thinning stage. We tested the use of maximum DBH (at plot level) and quadratic mean diameter (at stand level) as possible alternatives to age, but in this research, they were both negatively correlated with regeneration presence. For maximum DBH, it is likely that large trees present in the small study plots (5.6 m radius) were shading the ground and dispersing their seed outside the plots. Schweiger & Sterba (1997) considered quadratic mean diameter to be a compound measure of age, density and site quality, and here it seems the density effect was predominant. Sitka spruce is a prolific seeding species (up to 20 million seed per ha released under canopy) with an estimated dispersal distance of 60-80 m (von Ow et al. 1996, Nixon & Worrell 1999). In pure, even-aged stands seed availability is likely to be a factor not associated with the trees present at local level but with the general production at stand level, with little spatial variation (Malcolm et al. 2001). This may change in mixed-species, uneven-aged stands. In those situations, especially since age will not anymore be a suitable measure to describe the stand correctly, better studies on the role of mother trees and seed availability will be necessary. After checking the field notes, cone production that was exceptionally higher or lower than expected for that age of stand was a possible cause of error in the worst-simulated stands, suggesting that seed availability is not only controlled by age, even in single-species plantations.

Mosses showed a positive effect on regeneration consistent with previous findings. A thin layer of mosses cover is favourable for germination due to their water retention capacity, but heavy mosses can prevent roots from reaching the mineral soil (von Ow et al. 1996). LePage et al. (2000) found that the same ground cover can have different effects on regeneration according to the overstorey characteristics: for example, the positive effects of moss cover decreased with an increase in canopy cover. These various aspects could be the cause of the relatively low effect of mosses in the model. Further, in some stands the combined presence of competing ground vegetation (such as bramble, shrubs and tall grasses) and mosses seems to have affected the accuracy of the simulation. Additional studies may be necessary, considering the use of more specific classes (such as light and heavy mosses, deadwood in various stage of decomposition).

Increasing competition from the overstorey, expressed here as the quadratic term of basal area, influenced the regeneration negatively. However, for Thinning Class 1, at low overstorey levels the effect was relatively low and almost flat, likely confirming the benefit of a certain amount of shading. The same levels of basal area can be obtained with different numbers of trees, resulting in different canopy structures and thus light availability on the ground. When the number of trees is lower for a given basal area, there are likely to be more gaps between crowns and significantly more light at ground level (Hale et al. 2009). However, it is possible that the number of stems per ha was not significant in the present study because both age and Thinning Class were already partially describing the reduced number of trees resulting from natural mortality and anthropic removals.

None of the topographic and climatic variables tested showed significance. The climatic data were interpolations for 10 km grid squares of average climatic data collected during 1960-90. They had already been found in another study to lack the precision needed for stand analysis (Moore et al. 2009). The under-canopy climate is also generally different from the climate of open sites, with a degree of variation according to the stand characteristics (Sellars 2005). Significant differences between forest districts were not identified in this study. Foresters have not observed regional differences in the occurrence of Sitka spruce natural regeneration across the UK (Mason pers. comm.).

The cross-validation process with the use of the Response Operator Characteristics curve showed satisfactory statistical results at plot level: 73% of plots were correctly simulated, with similar values for specificity and sensitivity. The stochastic method, such as is employed by various models, showed worse results: 66% of total plots correctly simulated, with a larger difference between sensitivity and specificity. However, when aggregating the results at stand level and considering the difference between the total simulated and total observed plots with regeneration, the results were similar between methods: around two-thirds of the stands showed an acceptable error (simulated values within ± 20% of observed values). In a non-spatial forest growth simulator (*sensu* Robinson & Ek 2000) such as MOSES_GB, the accuracy at stand level may be more important than at plot level since the actual positions of the trees are not known.

The results of the independent validation with the Response Operator Characteristics curve method were not satisfactory since the model predicted regeneration in almost all plots, even if the total accuracy was 82%. Using the stochastic method, the total accuracy was worst (64%), although there was a slightly better balance between sensitivity and specificity. It is evident that the independent dataset is describing a situation largely different from the calibration dataset, noting the differences both in the stand variables (Tables 1 and 2) and the high frequency of plots with regeneration presence (about 80% in the independent dataset versus 40% of the calibration). The independent validation stands surveyed have been managed specifically to obtain natural regeneration. All the stands belonged to the Thinning Class 2, but most of them had been thinned more regularly and with higher intensity than those in the calibration dataset. When we aggregated the results at stand level and considered the difference between the total simulated and total observed plots with regeneration, the results were better for the stochastic method: two-thirds of the stands showed an acceptable error (simulated values within ± 20% of observed values), against half for the Response Operator Characteristics curve method. It seems that the cut-off calculated for the cross-validation process cannot be applied to the independent dataset, and although the model still presents problems in its application to continuous cover forestry situations, the stochastic method gave better results in this case.

The models tested here for regeneration density did not give results of acceptable accuracy. Generating random numbers from Weibull distributions was, in the present study, the only option found and still produced inadequate results both during the auto-validation and independent validation. Nonetheless, even if the models were deemed too inaccurate, it is interesting to note that the effect of basal area was significant and positive in the seedling density model based only on plots belonging to Thinning Class 1, suggesting a possible mother-tree positive effect. In the model for Thinning Class 2 & 3, basal area had a negative effect, maybe because the already-lower light availability is aggravated by bigger tree size and the overstorey competition effect becomes predominant. Similar results were observed by Page et al. (2001) in Sitka spruce forests in the UK.

A very important limitation of both models was the lack of data on the regeneration size or age. Both the regeneration presence and density model did not consider the possibility of other tree species germinating and competing with Sitka spruce, likely another crucial limitation of the use of these models in mixed forest stands resulting from continuous cover forestry practices. Presence of deer browsing, although not statistically significant in this analysis, was found in the field notes as a possible cause of limiting factor for regeneration in some sites where all the model variables were at a beneficial level for regeneration.

Concluding, the tools here described can be used to simulate regeneration presence in traditional Sitka spruce plantations in the UK. Then, the growth of the regeneration can be predicted with the light-growth models presented by Bianchi et al. (2018). However, the regeneration occurrence tools are not adequate for forests already in an advanced stage of transformation to CCF systems, and the density results must be treated with caution.

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Tables

Table 1. Some of the crucial factors influencing Sitka spruce natural regeneration, the general conclusions drawn in the literature about them, and the evidence quality of such conclusions. Adapted from Mason (2015).

Factor	Conclusions	Evidence Quality
Seed availability	Mast years very important, in British Sitka spruce stands happening every 4-5 years	Good-Moderate
Germination conditions	Favourable seedbed conditions: moist soils with needle litter or light moss cover	Moderate-Poor
Vegetation competition	Avoid fertile sites or competition from ericaceous vegetation	Moderate
Understorey microclimate	Retain some canopy cover to limit frost damage but provide adequate light	Moderate
Light requirements for growth	At least 20% of full light, plus an overstorey with basal area of 30 m^2/ha and reduced tree density	Good
Browsing pressure	Keep deer population below 5 animals per 100 ha	Moderate

Table 2. Details of calibration dataset. Values at stand (Age, Quadratic Mean Diameter, Soil nutrient regime, Time after last thinning and Deer Impact Index) and plot level (the remaining parameters).

Variable	**Min.**	**1st Qu.**	**Mean**	**3rd Qu.**	**Max.**
Age (years)	32	39	54.5	64	85
Basal area (m^2 ha^{-1})	1.6	43.6	58.0	70.0	196.0
Stems per hectare (n ha^{-1})	0	400	700	900	2,200
Quadratic mean diameter (cm)	0	27.0	36.3	43.1	83.0
Maximum diameter breast height (cm)	0	36.0	45.3	52.0	90.0
Global Site Factor	0.02	0.16	0.21	0.26	0.55
Bare ground (%)	0	0	1.2	0	85.0
Mosses (%)	0	5.0	41.6	80.0	95.0
Seedling density (ha^{-1})	0	0	20,780	10,000	450,000
Soil Nutrient Regime	**Very Rich**	**Rich**	**Medium**	**Poor**	**Very poor**
Plots (n)	10	10	130	180	10
Time after last thinning	**Class 1 (1-5 years)**		**Class 2 (6-10 years)**		**Class 3 (10+ years)**
Plots (n)	170		90		80
Deer Impact Index	**Low**		**Moderate**		**High**
Plots (n)	10		290		40

Table 3. Details of validation dataset. Values at stand (Age and Time after last thinning) and plot level (the remaining parameters).

Variable	**Min.**	**1st Qu.**	**Mean**	**3rd Qu.**	**Max.**
Age (years)	60	65	69	77	80
Basal area (m^2 ha^{-1})	7.6	28.8	41.4	53.8	107.2
Stems per hectare (n ha^{-1})	50	200	284	400	1,100
Maximum diameter at breast height (cm)	35	44	50.6	55.8	85
Global site factor	0.08	0.22	0.28	0.34	0.49
Bare ground (%)	0	0	0.1	0	4
Mosses (%)	0	70.6	82.3	99.7	100
Seedling density (ha^{-1})	0	2,500	48,460	52,500	417,500
Time after last thinning	**Class 1 (1-5 years)**		**Class 2 (6-10 years)**		**Class 3 (10+ years)**
Plots (n)	0		78		0

Table 4. Contingency tables for both the cross-validation and the independent validation results, using both the Response Operator Curve (ROC) method and stochastic method. YES indicates the presence of regeneration, NO the absence.

ROC method Cross-validation		**Predicted** YES	NO	Total	**Partial Accuracy**
	YES	105	33	138	0.76
Observed	NO	58	144	202	0.71
	Total	162	178	340	
Overall accuracy					0.73

Stochastic method Cross-validation		**Predicted** YES	NO	Total	**Partial accuracy**
	YES	76	62	138	0.55
Observed	NO	52	150	202	0.74
	Total	128	212	340	
Overall accuracy					0.66

ROC method Independent validation		**Predicted** YES	NO	Total	**Partial accuracy**
	YES	62	0	62	1.00
Observed	NO	14	2	16	0.12
	Total	76	2	78	
Overall accuracy					0.82

Stochastic method Independent validation		**Predicted** YES	NO	Total	**Partial accuracy**
	YES	44	18	62	0.71
Observed	NO	10	6	16	0.39
	Total	54	24	78	
Overall accuracy					0.64

Table 5. Parameters for the Weibull distributions fitted to seedling density per ha (thousands)

	Shape	Rate
Thinning Class 1	0.696	52.555
Thinning Class 2	0.871	14.134
Thinning Class 3	1.834	4.651

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Figures (caption)

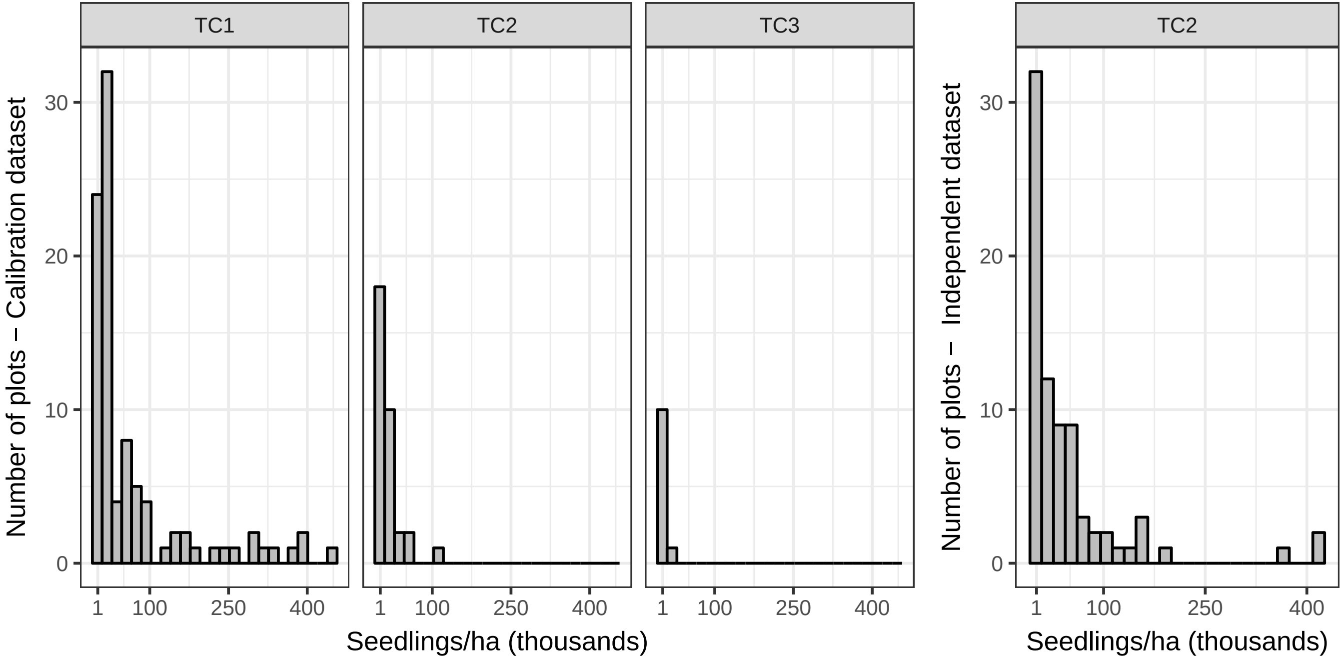
Figure 1. Frequency of seedlings per hectare in different Thinning Classes (TC1, TC2 and TC3), only plots with presence of regeneration, for the calibration dataset (left) and the independent validation dataset (right).

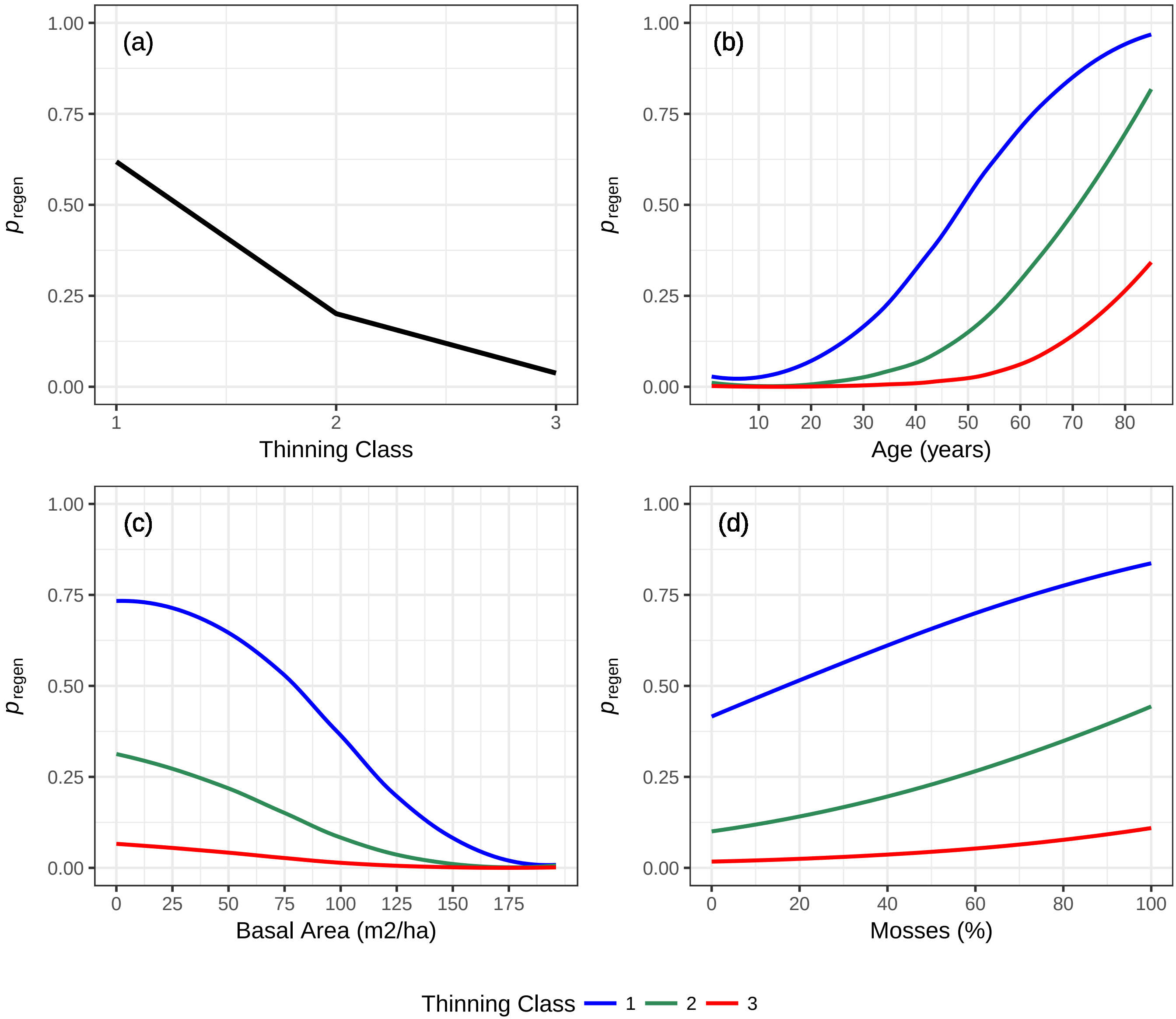
Figure 2. Regeneration presence likelihood (p_{regen}) as a function of the model variables. In each graph, the likelihood was estimated with only one variable varying across all its range (plotted on the x-axis), while the others were kept at the calibration population mean. Multiple lines indicate the analysis used different values of Thinning Class

Figure 3. Coefficient values after standardization of the model variables (BA = Basal area, TC= Thinning class). The dot corresponds to the mean values, the wider blue line to the 90% confidence interval, the narrower blue line to the 95% confidence interval

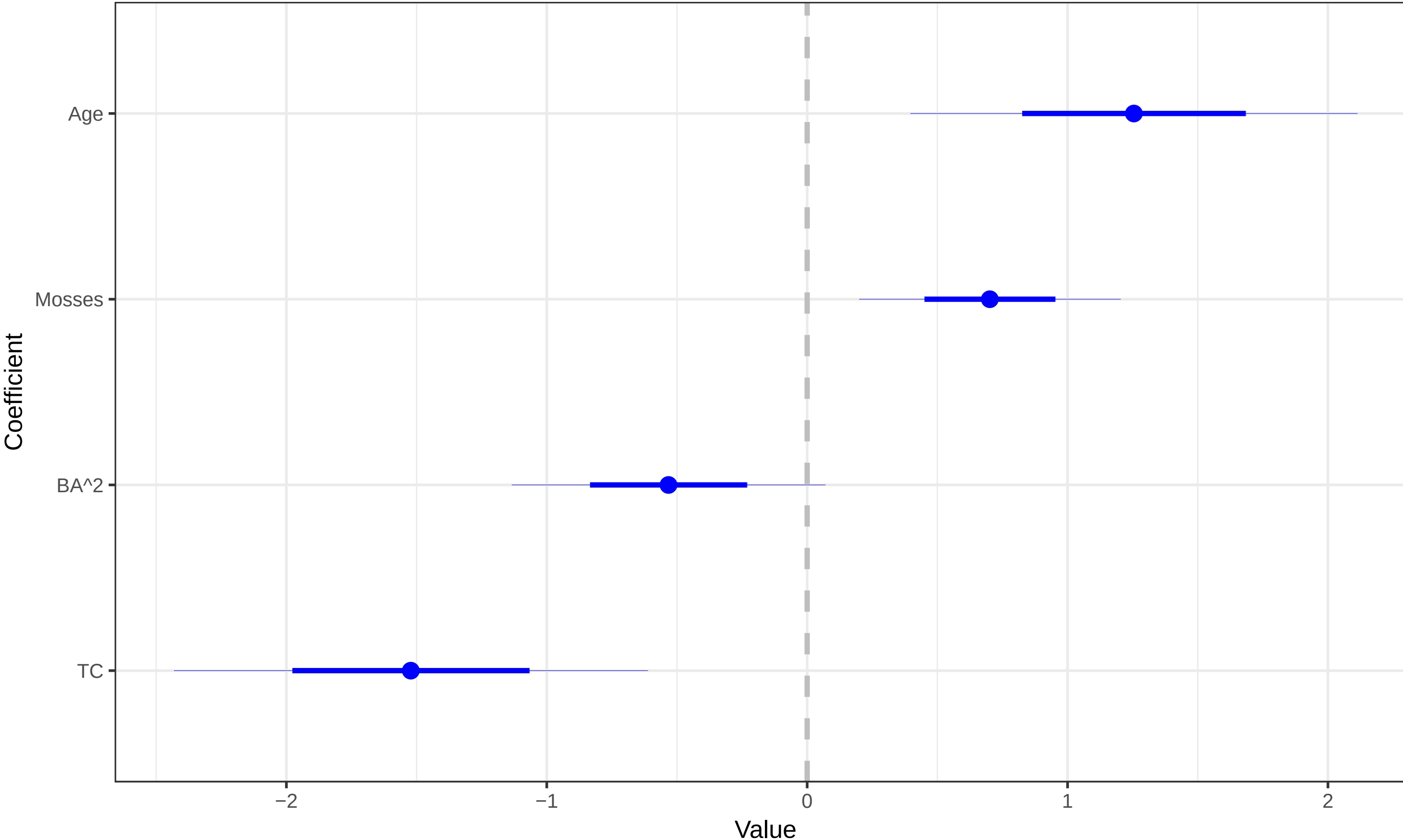
Figure 4. Receiver Operator Characteristics (ROC) curve for the cross-validation method. The dot represents the point with the highest sum of the specificity and sensitivity values (presented between parentheses) and shows the corresponding cut-off likelihood value.

Figure 5. Probability densities of the fitted Weibull distributions (lines) vs probability densities of the observed number of seedlings per ha at plot level (bars), according to Thinning Class (TC).

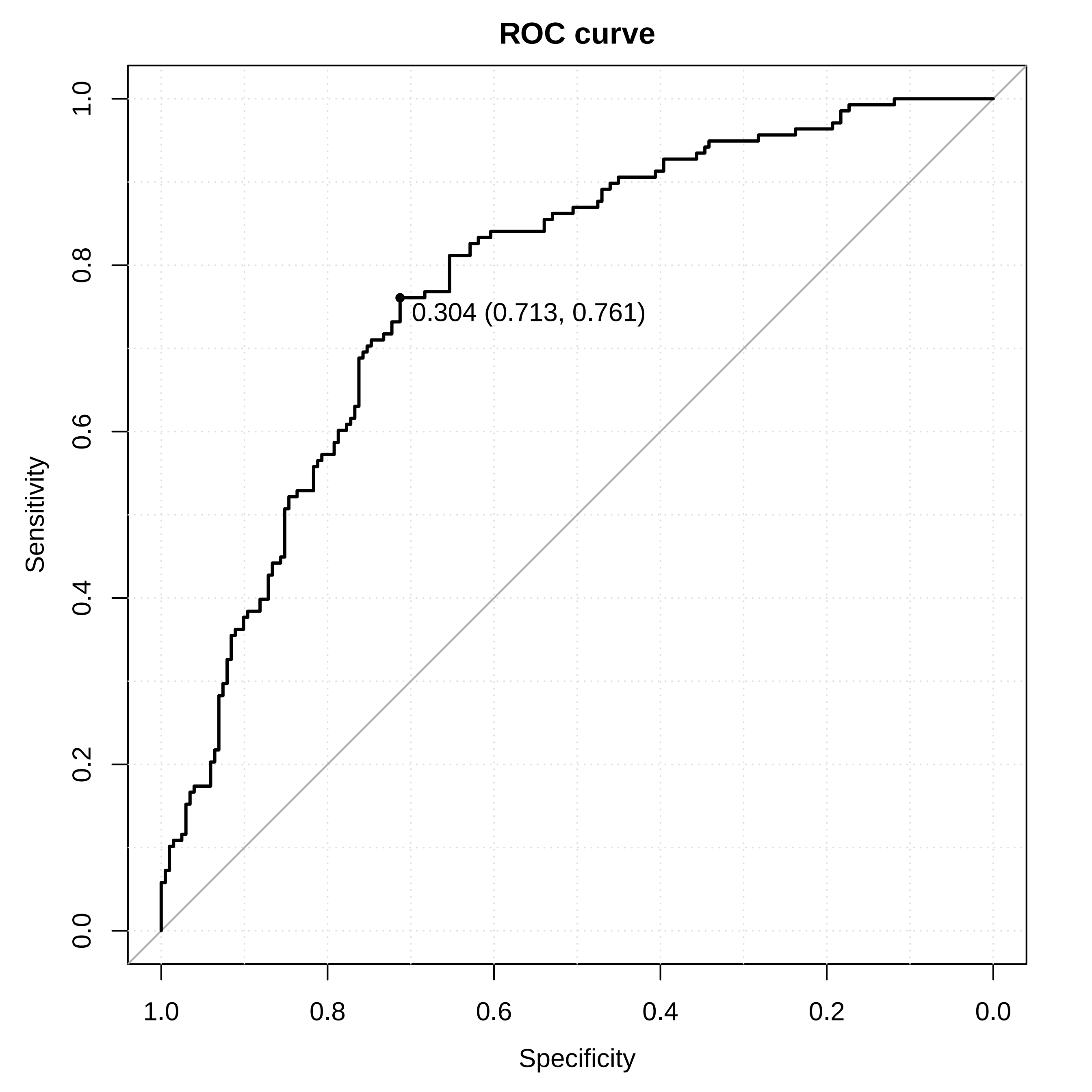

TC1
TC2
TC3
TC2
Number of plots – Calibration dataset
Number of plots – Independent dataset
0
10
20
30
0
10
20
30
1
100
250
400
Seedlings/ha (thousands)
Seedlings/ha (thousands)

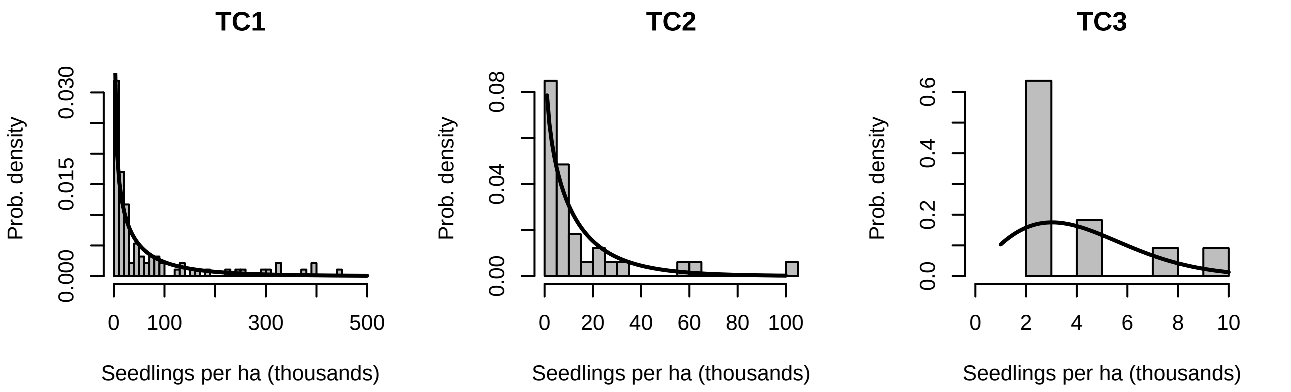


(a)
(b)
(c)
(d)
p_{regen}
1.00
0.75
0.50
0.25
0.00
Thinning Class
1
2
3
10
20
30
40
50
60
70
80
Age (years)
0
25
50
75
100
125
150
175
Basal Area (m2/ha)
0
20
40
60
80
100
Mosses (%)
Thinning Class
1
2
3



Age
Mosses
BA^2
TC
Coefficient
−2
−1
0
1
2
Value


ROC curve
Sensitivity
Specificity
0.304 (0.713, 0.761)
1.0
0.8
0.6
0.4
0.2
0.0


TC1
TC2
TC3
Prob. density
0.000
0.015
0.030
0.00
0.04
0.08
0.0
0.2
0.4
0.6
0
100
300
500
20
40
60
80
2
4
6
8
10
Seedlings per ha (thousands)