

## Moving from biodiversity offsets to a target-based approach for ecological compensation

Simmonds, Jeremy S.; Sonter, Laura J. ; Watson, James E. M. ; Bennun, Leon; Costa, Hugo M.; Dutson, Guy; Edwards, Stephen ; Grantham, Hedley; Griffiths, Victoria F.; Jones, J.P.G.; Kiesecker, Joseph ; Possingham, Hugh P.; Puydarrieux, Philippe; Quétier, Fabien; Rainer, Helga ; Rainey, Hugo; Roe, Dilys ; Savy, Conrad E.; Souquet, Mathieu ; ten Kate, Kerry ; Victurine, Ray; von Hase, Amrei ; Maron, Martine

### Conservation Letters

DOI:

[10.1111/conl.12695](https://doi.org/10.1111/conl.12695)

Published: 01/04/2020

Publisher's PDF, also known as Version of record

[Cyswllt i'r cyhoeddiad / Link to publication](#)

*Dyfyniad o'r fersiwn a gyhoeddwyd / Citation for published version (APA):*

Simmonds, J. S., Sonter, L. J., Watson, J. E. M., Bennun, L., Costa, H. M., Dutson, G., Edwards, S., Grantham, H., Griffiths, V. F., Jones, J. P. G., Kiesecker, J., Possingham, H. P., Puydarrieux, P., Quétier, F., Rainer, H., Rainey, H., Roe, D., Savy, C. E., Souquet, M., ... Maron, M. (2020). Moving from biodiversity offsets to a target-based approach for ecological compensation. *Conservation Letters*, 13(2), Article e12695. <https://doi.org/10.1111/conl.12695>

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**REVIEW**

# Moving from biodiversity offsets to a target-based approach for ecological compensation

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**Funding information**

Australian Research Council, Grant/Award Numbers: DE170100684, FT140100516; National Environmental Science Programme; Science for Nature and People Partnership; COMBO Project

**Abstract**

Loss of habitats or ecosystems arising from development projects (e.g., infrastructure, resource extraction, urban expansion) are frequently addressed through biodiversity offsetting. As currently implemented, offsetting typically requires an outcome of “no net loss” of biodiversity, but only relative to a baseline trajectory of biodiversity decline. This type of “relative” no net loss entrenches ongoing biodiversity loss, and is misaligned with biodiversity targets that require “absolute” no net loss or “net gain.” Here, we review the limitations of biodiversity offsetting, and in response,

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propose a new framework for compensating for biodiversity losses from development in a way that is aligned explicitly with jurisdictional biodiversity targets. In the framework, targets for particular biodiversity features are achieved via one of three pathways: Net Gain, No Net Loss, or (rarely) Managed Net Loss. We outline how to set the type (“Maintenance” or “Improvement”) and amount of ecological compensation that is appropriate for proportionately contributing to the achievement of different targets. This framework advances ecological compensation beyond a reactive, ad-hoc response, to ensuring alignment between actions addressing residual biodiversity losses and achievement of overarching targets for biodiversity conservation.

#### KEYWORDS

averted loss, biodiversity loss, Convention on Biological Diversity, counterfactual, environmental impact assessment, environmental policy, infrastructure development, mitigation hierarchy, net gain, no net loss

## 1 | INTRODUCTION

The 196 Parties to the Convention on Biological Diversity (CBD) are currently setting ambitious post-2020 biodiversity targets (Mace et al., 2018; Visconti et al., 2019). Yet, despite widespread recognition of the need to slow and ultimately halt biodiversity loss, transformation of the natural world for infrastructure, industry, commercial agriculture, urbanization, and resource extraction (hereafter, “development”) continues to drive declines (Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services [IPBES], 2019). Ceasing all such transformation is not feasible in the face of desirable development imperatives (Griggs et al., 2013; United Nations, 2018). Governments, developers, and civil society therefore need tools for reconciling development and conservation to reduce the rate of biodiversity loss.

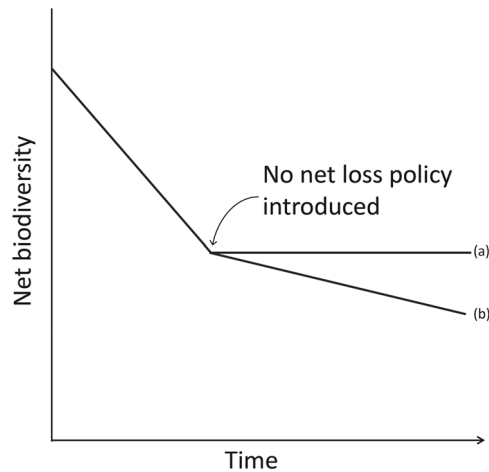
The mitigation hierarchy is an approach for responding to biodiversity losses arising from development. It has been embedded into numerous government, lender, and corporate policies (Business and Biodiversity Offsets Programme [BBOP], 2012; Gardner et al., 2013; International Finance Corporation [IFC], 2012; IUCN, 2018a; Rainey et al., 2014). Proponents of development projects—where these are mandated by policy (“regulated sectors”)—are required to reduce adverse biodiversity outcomes through sequentially following four steps. Only after completing avoidance, and then restoration/rehabilitation of disturbed areas onsite, should the fourth step be taken—compensating for any residual losses through biodiversity offsetting. When applied as the final step of the mitigation hierarchy, biodiversity offsets are typically intended to achieve an outcome in which there is (at least) “no net loss” of the impacted biodiversity due to a particular project (BBOP, 2012; Bull, Gordon, Watson, & Maron, 2016; IUCN, 2016).

Biodiversity offsetting, however, is almost never designed to align with the achievement of national or sub-national

(“jurisdictional”) biodiversity targets that aim to halt species and ecosystem decline, or achieve biodiversity recovery. In large part, this is because no net loss of biodiversity at the level of individual development projects can mean something quite different to no net loss at the jurisdictional level (Maron et al., 2018). When framed in relation to a jurisdictional biodiversity target, no net loss implies that the amount of a particular biodiversity feature (e.g., forest) should not fall below what we have now; in other words, it means no net loss relative to a “fixed reference scenario” (Maron et al., 2018). Under such a scenario, any lost forest (for example) would need to be replaced to achieve absolute no net loss—that is, to maintain the amount of forest at its current level (Figure 1).

This is rarely the intended meaning of no net loss in offset policies that guide compensation for residual losses at the development project level, not the jurisdictional level. Project-level no net loss is often framed relative to a counterfactual scenario of decline, in which biodiversity is expected to be lost even without the development (and its offset) (IUCN, 2016; Maron, Bull, Evans, & Gordon, 2015). The rationale is that the protection provided by the offset action achieves a benefit by averting a loss or decline that would otherwise have occurred. Such “averted loss” offsetting (also called avoided loss or protection offsetting) is one of the two main forms of biodiversity offsetting (the other being restoration). It is referenced as a key approach to offsetting in policies and standards espoused by financial institutions (IFC, 2012; World Bank Group, 2016), multistakeholder platforms (BBOP, 2012; IUCN, 2016), and jurisdictions (Australia, Columbia, and Chile; Maron et al., 2018). In a global review of over 12,000 individual offsets projects, Bull and Strange (2018) found that approximately 66% used averted loss offsetting, either exclusively or in combination with other measures.

When framed this way, even best-practice offsets result in less biodiversity over time, as protection of already-existing biodiversity, which is expected to decline in the future, can be



**FIGURE 1** ‘No net loss’ relating to different reference scenarios. No net loss at the jurisdictional level implies that loss is stopped in absolute terms compared to a fixed reference scenario—that is, that all biodiversity losses are addressed by gains of the same size, thus maintaining biodiversity at the same level compared to before the loss occurred (a). However, in reality, no net loss commitments frequently only require that individual projects achieve no net loss relative to a declining counterfactual, by protecting biodiversity that might otherwise be lost in the future due to unregulated impacts (“averted loss”) (b). Such project-level no net loss results in ongoing loss of biodiversity at the jurisdictional level, albeit at a slower rate (figure adapted from Maron et al., 2018)

exchanged for biodiversity losses at the development site(s) (Bekessy et al., 2010; Buschke, Brownlie, & Manuel, 2017; Maron et al., 2018; Moilanen & Laitila, 2016). Across multiple projects, offsetting that achieves no net loss relative to a counterfactual scenario of biodiversity decline maintains the declining trend, and corresponds with a net loss at the jurisdictional level (Figure 1; Gibbons & Lindenmayer, 2007; Quétiér, van Teeffelen, Pilgrim, von Hase, & ten Kate, 2015).

Relative no net loss of biodiversity at the project level does not equate with the achievement of absolute no net loss at the jurisdictional level. This mismatch causes conceptual confusion and ambiguity about the meaning and intention of no net loss as a policy objective. It also makes it hard to assess the contribution that project-level compensatory actions (e.g., biodiversity offsetting) are making to broader conservation goals, such as the achievement of jurisdictional biodiversity targets (Maron et al., 2018). We are aware of only one national policy that links compensatory actions to the achievement of a target (limiting ecosystem loss to pre-defined thresholds)—South Africa’s Draft National Biodiversity Offset Policy (Republic of South Africa, 2017). If offsetting continues to occur in isolation from broader conservation imperatives, the risk is that at best, offsetting will contribute minimally to conservation objectives, and at worst, will detract from achieving such goals (e.g., where counterfactual-based approaches entrench ongoing declines; Maron et al., 2018). An overar-

ching framework is therefore needed to align project-level actions under the mitigation hierarchy, particularly of ecological compensation for residual losses, with the biodiversity targets that a jurisdiction may strive to achieve.

Here, we propose such a framework, and review its suitability in applied conservation policy. We refer throughout to “ecological compensation” to distinguish our proposed approach as an alternative to the narrower concept of biodiversity offsetting, which has strict rules about like-for-like trades in biodiversity and aims to achieve at least no net loss relative to a counterfactual scenario (BBOP, 2012; Bull et al., 2016; IUCN, 2016). We discuss the consequences of different approaches to ecological compensation, and provide guidance on how, where, and when the framework we present could be operationalized. This framework entails several advantages over current practice. First, it makes explicit the contribution of ecological compensation toward meeting jurisdictional biodiversity targets. Second, it avoids the need for complex (and highly uncertain) calculations of the counterfactual scenario. Third, it strengthens the focus on avoidance, because it explicitly identifies instances where biodiversity losses require proportionate increases through actions such as restoration, which will not always be a feasible option. Fourth, it provides conceptual clarity; the net outcome across impact and compensation sites for a particular project would align with the desired net outcome at the jurisdictional level.

## 1.1 | Jurisdictional-level biodiversity targets

The framework we propose is general, and can apply to any biodiversity targets that describe a desired state of biodiversity (“outcome-based targets”) at any jurisdictional scale. Target-setting is not a part of the framework, but the existence of quantifiable targets is a pre-requisite for its implementation. Indeed, the targets that we refer to in this framework should be set independently of, and have primacy over, policy relating to the mitigation hierarchy and compensation. This is to prevent targets being designed to facilitate a particular policy approach.

Biodiversity targets are a familiar concept. Under the CBD Strategic Plan for Biodiversity 2010–2020 (CBD, 2010), more than 160 Parties to the CBD already have targets for biodiversity conservation laid out in their National Biodiversity Strategy and Action Plans (a response to the 20 global Aichi Targets agreed in 2010; UNEP, 2019). However, these are often not outcome-based targets (IUCN, 2018b)—a reflection of the fact that the Aichi Targets themselves are predominantly non-quantifiable, and lack focus on desired outcomes (Barnes, Glew, Wyborn, & Craigie, 2018; Butchart, Di Marco, & Watson, 2016).

As Parties to the CBD negotiate the post-2020 global biodiversity framework, there are increasing calls for clear,

quantifiable science-based targets for the retention and recovery of biodiversity and nature (Dinerstein et al., 2019; Mace et al., 2018; Maron, Simmonds, & Watson, 2018; Visconti et al., 2019; Watson & Venter, 2017). Such targets should be incorporated in national plans and actions, and linked to the achievement of broader global goals (IUCN, 2018b; Mace et al., 2018). Plentiful guidance on target-setting is available (Butchart et al., 2016; Carwardine, Klein, Wilson, Pressey, & Possingham, 2009; Di Marco, Watson, Venter, & Possingham, 2016; Doherty et al., 2018; Maron et al., 2018; Maxwell et al., 2015; Watson & Venter, 2017). The framework we present requires that targets are measurable, and explicitly reflect the desired state (outcome) of the biodiversity feature (e.g., species population, ecosystem extent) on which the target focuses, rather than a desired rate of change, or a mechanism for achieving the target. Examples of such targets that already exist include the French Government's pledge to support and maintain a population of 500 wolves for the years 2018 to 2023 (Republique Francaise, 2018), and ecosystem-specific retention thresholds that are incorporated into South Africa's Draft National Offset Policy (Brownlie et al., 2017; see Supporting Information 1).

## 2 | FRAMEWORK OVERVIEW

### 2.1 | Aligning ecological compensation with biodiversity targets

In this framework, targeted conservation outcomes such as desired species populations or minimum ecosystem extents are set in absolute terms at the jurisdictional level. The required trajectory needed to achieve a target for a particular species, assemblage, or ecosystem (hereafter, "biodiversity feature") depends on the level (e.g., number, amount, area) of the biodiversity feature when the jurisdictional-level target for that biodiversity feature was set (Figure 2).

When a biodiversity feature is approximately at the target level, ongoing "No Net Loss" is required. All losses of the biodiversity feature need to be balanced by proportionate gains in order to maintain the biodiversity feature at the target level. It follows that when a biodiversity feature is below the target level, "Net Gain" is needed to achieve the target, whereby the biodiversity feature increases in absolute terms to (at least) the point where the target is met. "Managed Net Loss" may be appropriate in exceptional circumstances when a biodiversity feature is above its target. Setting a target below current levels might require that: (a) the particular biodiversity feature is very common and widespread; (b) some losses at the jurisdictional level can occur without compromising the ecological integrity and function of the feature (e.g., population viability, intactness); and (c) continued, strictly managed drawdown to a predetermined target level is socially acceptable.

Once a jurisdiction has established targets, and thus specified the required trajectory for its biodiversity features, project-level actions under the mitigation hierarchy can be designed to contribute to achieving these targets. The approach to compensating for residual losses at the project level depends upon several factors. The type of compensatory action depends on whether achievement of the jurisdictional biodiversity target requires Net Gain, No Net Loss, or occasionally in specific situations allows for Managed Net Loss. The amount of compensation required for any given project is guided by the amount of residual loss, how much of the affected biodiversity feature remains relative to its particular target, and policy decisions regarding the share of responsibility among sectors. Below, we set out each consideration.

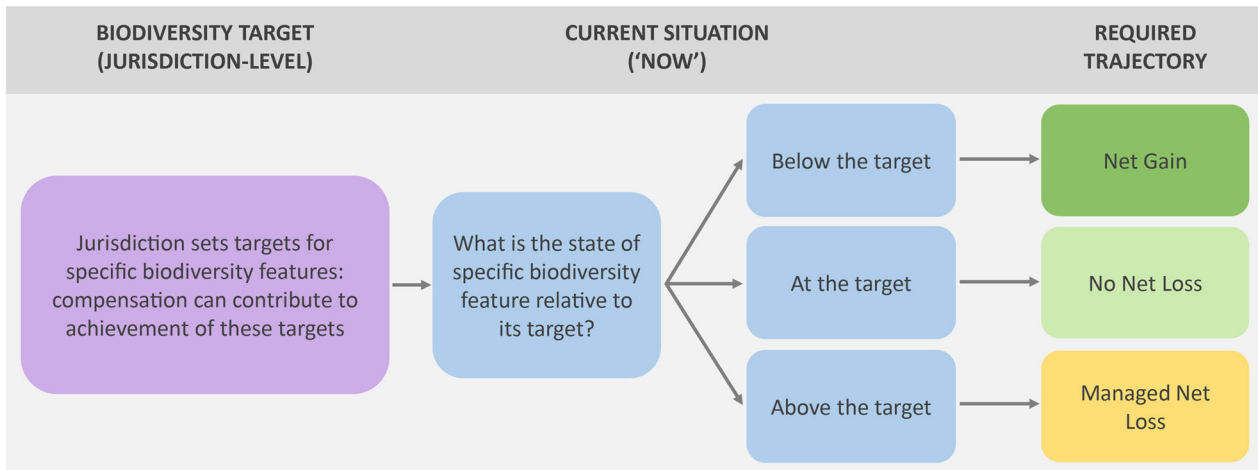
### 2.2 | Achieving jurisdictional outcomes—improvement, maintenance, and avoidance

There are two broad types of ecological compensation in this framework: Maintenance and Improvement. By "Maintenance" we mean preventing a threat to ensure persistence of a biodiversity feature at its current condition, extent, or population (and conservation status), for example, by legally securing existing biodiversity at a compensation site. The aim of Maintenance is to prevent existing biodiversity from being lost at a site in the future (i.e., avert future losses). The net result of Maintenance interventions across a jurisdiction is a reduction in the biodiversity feature, because the loss from development is compensated for by securing the persistence of the biodiversity feature at another site(s), where it already exists.

This contrasts with "Improvement," which involves producing a quantifiable increase in the biodiversity feature. Improvement can take a range of forms, and result from a variety of interventions such as habitat enhancement (e.g., improving condition of native vegetation) or removal of pervasive pressures to allow populations to increase (e.g., invasive species control). In reality, the interventions that achieve Maintenance and Improvement at a site can overlap—legally securing a site and managing it at a moderate intensity might preserve that site's condition (Maintenance), but if management intensity is increased it might achieve Improvement; similarly, legal protection of a degraded site might over time allow its recovery (Improvement). Generally, Improvement will require complementary Maintenance as a necessary prerequisite (e.g., securing a site containing the focal biodiversity feature or its habitat, with a view to improving it).

Enhancing biodiversity, including Improvement compensation actions, is ultimately essential for achieving jurisdictional-level No Net Loss or Net Gain—only by increasing the extent and/or condition or amount of a biodiversity feature can No Net Loss (or Net Gain) be achieved under





**FIGURE 2** Aligning ecological compensation with jurisdictional biodiversity targets starts with establishing the trajectory required to achieve net target outcomes. The required trajectory depends on whether a biodiversity feature is below, at, or above its jurisdictional biodiversity target at the time the target is set (“now”)

this framework (Figures 3a and 3b). When carefully linked to biodiversity targets, Maintenance can be used to contribute to Managed Net Loss, until such time that the target is reached, after which Improvement becomes an essential response to any permitted losses (Figure 3c). Further, while Maintenance alone cannot achieve No Net Loss or Net Gain at the jurisdictional level, it may be a necessary transitional intervention to ultimately achieving these outcomes in the common situation where a biodiversity feature is (a) below its target; and (b) experiencing rapid and ongoing loss from unregulated pressures, where the mitigation hierarchy is not fully applied. In these circumstances, compensation through Maintenance may be appropriate for a transitional period alongside or in advance of compensation through Improvement (Figure 3d). However, for such an approach to be a step toward No Net Loss or Net Gain, transition phases with strict limits must be set (see Supporting Information 1).

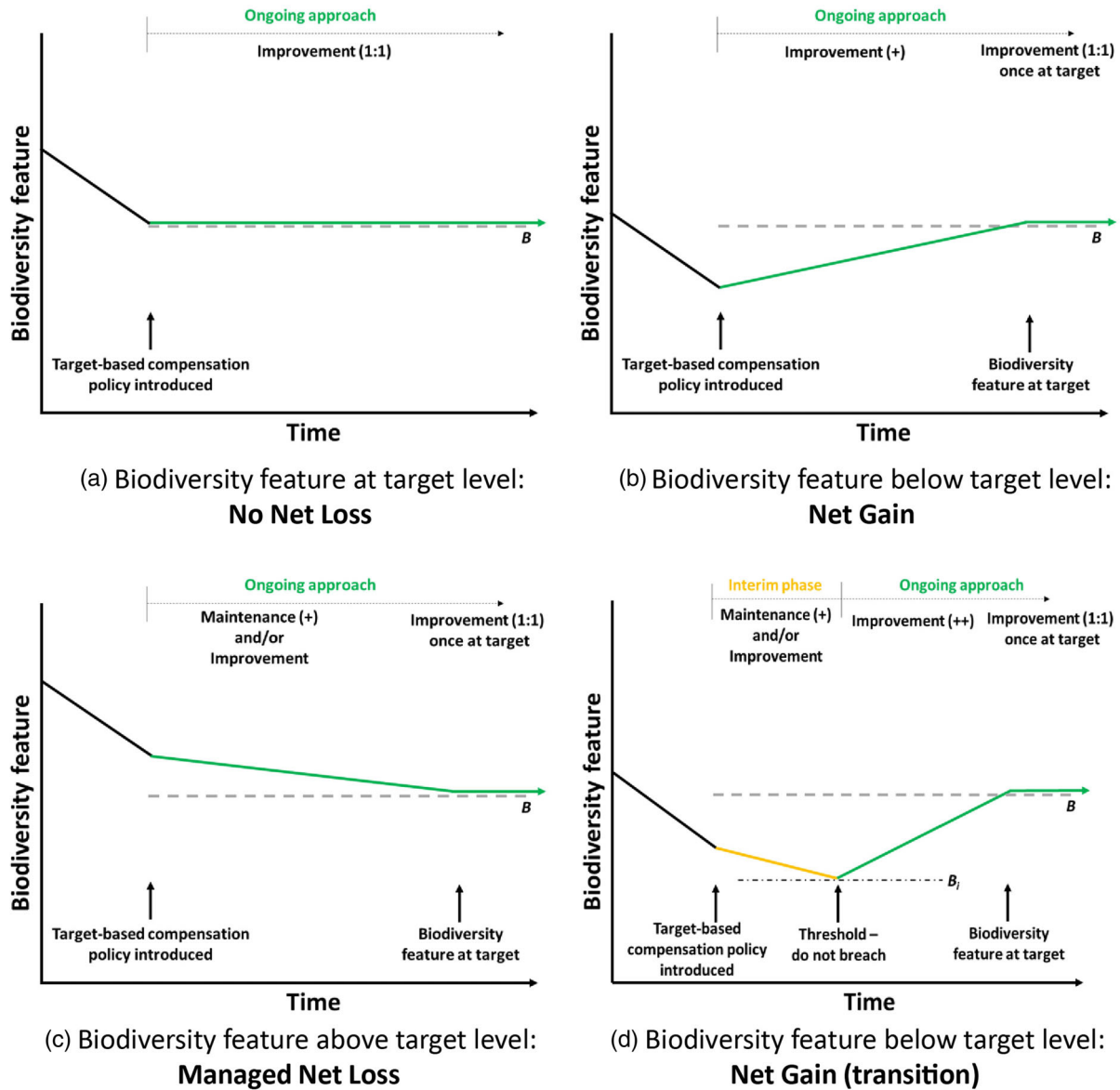
Because this framework explicitly links ecological compensation requirements with jurisdictional-level target outcomes, it strengthens the focus on rigorously applying the earlier steps of the mitigation hierarchy. Jurisdictional No Net Loss or Net Gain cannot occur without losses being compensated by Improvement actions such as restoration or increases in species’ populations. However, for some biodiversity features, achieving gains through actions such as restoration is either hampered by great uncertainty, or is simply not possible (given, e.g., substantial time lags; Curran, Hellweg, & Beck, 2014; Gibbons et al., 2016; Maron et al., 2012; Moilanen, van Teeffelen, Ben-Haim, & Ferrier, 2009; Pilgrim et al., 2013). This reality limits considerably the types of biodiversity features for which No Net Loss or Net Gain are feasible. Losses of irreplaceable biodiversity features simply cannot be managed through a compensation approach, unless the jurisdictional target involves Managed Net Loss. If an outcome of

further (managed net) loss is unacceptable, the only option is more rigorously to apply the earlier steps in the mitigation hierarchy, and avoid losses entirely.

### 2.3 | The amount of compensation required for a given loss

This target-based framework no longer depends upon the complex and often counterintuitive process of defining dynamic counterfactual scenarios to establish what type of action, and how much, is required to compensate for a given loss (as offsetting does). This is because instead of a dynamic counterfactual scenario, a reference point fixed at a particular level—the target—is used. The amount of compensation required for any given project is determined by both how much residual loss a particular biodiversity feature experiences as a result of a development project, and the pathway (e.g., No Net Loss) required to achieve a target, along with several additional considerations (outlined below) that are factored into the calculation of a compensation ratio. The compensation ratios (sometimes called a “multiplier”) detailed here only need to be established once—at the inception of a compensation scheme—and should be applied consistently to all projects.

The compensation ratio sets the amount of Improvement or Maintenance required per unit of residual loss to contribute to the achievement of a target, as depicted in Figure 3. The first step in calculating the compensation ratio is to estimate how much of the affected biodiversity feature ( $x$ ) exists relative to its target (at time  $t = 0$ , when the target,  $B$ , was set). The current amount of  $x$  comprises two parts: how much of what exists is already considered effectively protected from adverse impacts (e.g., fully resourced protected areas) or planned to be so protected ( $x_p(0)$ ); and how much of what exists could



**FIGURE 3** Illustration of the target-based ecological compensation approach for contributing to the achievement of (a) No Net Loss; (b) Net Gain; (c) Managed Net Loss; and (d) Net Gain using a transitional approach in which Maintenance actions can be undertaken for a period of time to help stem unregulated losses, before Improvement actions become the default requirement. The dashed line on each plot represents the target level ( $B$ ) for the biodiversity feature. The indicative amount of Improvement and/or Maintenance (denoted by “+”) depends on the difference between the level of the biodiversity feature and the target (and in the case of the transitional approach [d]), the threshold ( $B_i$ ) below which the biodiversity feature cannot decline; see Supporting Information 1). Importantly, compensation for residual losses from development is one of a suite of complementary measures to achieve the desired trajectory and ultimately achieve a target. At such time that the target is met, maintaining the biodiversity feature at this level requires losses to be compensated for by Improvement at a ratio of 1:1 (or targets could be revised towards ambitious new objectives)

still conceivably be lost (including because of development projects;  $x_a(0)$ ). Places identified as being under effective protection ( $x_p$ ) are not available to be used for compensation.

Where No Net Loss or Net Gain is needed to achieve a target, the amount of compensation (gain via Improvement) required for a given unit of loss to a particular biodiversity feature is:

$$\text{Compensation ratio (Improvement)} = \frac{(B - x_p(0))}{x_a(0)} \quad (1)$$

Where Managed Net Loss is appropriate, the amount of compensation (securing existing biodiversity via Maintenance) required for a given unit of loss to a particular biodiversity feature is:

$$\text{Compensation ratio (Maintenance)} = \left( \frac{B - x_p(0)}{x(0) - B} \right) \quad (2)$$

For the transitional approach (Figure 3d), Equation (2) is used to set Maintenance requirements to ensure that an interim target (threshold) of  $B_i$  is not breached, before switching

to Improvement using Equation (1) to achieve the desired target. More details on calculating the Improvement and Maintenance compensation ratios (including for transitional approach) are provided in Supporting Information 1 and 2.

To exemplify these ratios, compensation for a project-level loss of 100 ha of habitat, consistent with Net Gain linked to a target of doubling the currently available habitat for a species, requires an Improvement ratio of 2:1. This is based on assumptions that none of the biodiversity feature is currently protected, and all adverse impacts to this biodiversity feature are regulated (i.e., follow the mitigation hierarchy). Here, a ratio 2:1 requires that 200 ha of “new” equivalent habitat must be successfully created (and maintained) to compensate for the loss. Similarly, Managed Net Loss in which 90% of a remaining ecosystem is to be retained would require a Maintenance ratio of 9:1, wherein nine times the area of residual loss is secured and retained into the future. Again, this assumes no current protection of the ecosystem, and no unregulated losses. If, say, half the remaining ecosystem was already effectively protected, the ratio would be 4:1.

These compensation ratios can vary with policy settings. For example, the ratios presented above are based on a proportionate contribution toward the achievement of the target. In other words, a unit of loss caused by a regulated sector requires the same amount of compensation as would a unit of unregulated loss (the liability for which accrues, in effect, to the jurisdictional government) in order to progress toward the target. However, in some instances a jurisdiction may require sectors that are regulated to contribute disproportionately toward a target’s achievement. For example, the jurisdiction may require that some sectors make additional contributions toward a biodiversity target, beyond just compensating for their own impacts. Alternatively, the government may shoulder some of the responsibility for compensation to limit the requirements on certain sectors. Government decisions about proportionate or disproportionate responsibility and policy scope (which sectors or type of impact are regulated) can affect both compensation ratios for regulated sectors and the amount of responsibility that falls on governments to address losses that are contrary to the required trajectory needed to achieve target commitments. Therefore, they must be made and factored in at the point of policy development when ratios are calculated (i.e., prior to the policy’s implementation; Supporting Information 2). This allows for transparency and clarity about which actor must do what action, how much of it, and why, to compensate for residual impacts in line with meeting desired targets.

Time lags in and uncertainty about achievement of compensatory outcomes are also often dealt with by adjusting ratios. These factors can be incorporated in this approach by increasing the ratios as appropriate (Bull, Lloyd, & Strange, 2017; Laitila, Moilanen, & Pouzols, 2014; Moilanen & Koti-aho, 2018). This particularly applies to Improvement, where

the unadjusted ratio assumes full and certain compensation instantly. The compensation ratio for Improvement thus gives the minimum compensation required for a particular unit of loss (to contribute to achievement of the target), and would need to be increased accordingly to account for time lags and uncertainties (e.g., restoration not being fully successful; Maron et al., 2012; Moilanen et al., 2009).

## 2.4 | Contrast with counterfactual-based offsetting

Both target-based ecological compensation, as described in this framework, and counterfactual-based offsetting, require strict adherence to the mitigation hierarchy, quantification of residual losses, and determination of compensatory requirements for these losses. The fundamental difference lies in how the compensation required for a particular biodiversity feature is calculated—now based on the overall jurisdictional biodiversity target and on policy choices for how to achieve it, rather than a project-specific assessment underpinned by complex counterfactual scenarios. This, and other differences, are summarized in Table 1. We note that some jurisdictions may lack the enabling environment to (a) develop and implement compensatory policy; and (b) determine and enact either targets for biodiversity conservation, or mechanisms for their achievement. In circumstances such as these, counterfactual-based offsetting may be more appropriate, although this should be considered a temporary solution given its inherent propensity for the uncapped drawdown of biodiversity. As long as appropriate, scientifically robust biodiversity targets can be set, we propose that a move toward a target-based approach is desirable.

## 3 | IMPLEMENTATION CONSIDERATIONS

While in its totality, target-based ecological compensation represents a novel alternative to the prevailing biodiversity offsetting paradigm, its component parts are familiar, with most aspects of existing standards remaining applicable (BBOP, 2012; Gardner et al., 2013; Gelcich, Vargas, Carreras, Castilla, & Donlan, 2017; IUCN, 2016). A target-based system involves changes only to the final step of the well-established mitigation hierarchy, primarily relating to the sizing of compensatory requirements. The on-ground actions (improving or maintaining biodiversity in a particular place) are no different to those in current offsetting practice, and are subject to the same challenges that affect these, and indeed most, applied conservation activities. Biodiversity targets are already central in international and jurisdictional policy. Target-based ecological compensation simply helps to connect project-level responses to these broad biodiversity targets to achieve desirable outcomes for stakeholders and



**TABLE 1** Comparison between counterfactual-based offsetting and target-based ecological compensation

	Advantages	Risks and challenges
<b>Counterfactual-based offsetting</b> (aiming for no net loss relative to a counterfactual scenario)	<ul style="list-style-type: none"> <li>• Can be implemented in the absence of any articulated conservation targets</li> <li>• Increases the attention on the difference made by a conservation intervention</li> <li>• Can be implemented for individual projects in poorly-regulated settings</li> <li>• Main concepts and approaches familiar to many practitioners / policy makers</li> </ul>	<ul style="list-style-type: none"> <li>• Outcomes are relative to a dynamic counterfactual trajectory that cannot be known in advance, only estimated</li> <li>• Biodiversity decline continues even though a project may achieve no net loss relative to a declining counterfactual</li> <li>• Constructing robust counterfactuals is conceptually complex and can be data-hungry</li> <li>• The type and amount of offset action required is highly sensitive to assumptions about the counterfactual trajectory</li> <li>• The end point of the biodiversity trajectory is implicit or unknown</li> <li>• Relatively easy to manipulate the counterfactual and thus undermine the net outcome</li> </ul>
<b>Target-based ecological compensation</b> (aiming for net jurisdictional outcomes aligned with specific biodiversity targets)	<ul style="list-style-type: none"> <li>• Aligns outcomes of actions regulated by compensatory policy with overarching conservation objectives</li> <li>• Outcomes are explicit and relative to a fixed, known point in time</li> <li>• ‘No Net Loss’, ‘Net Gain’ and ‘Managed Net Loss’ have intuitive meanings</li> <li>• Standardises calculation of the type and amount of compensation required</li> <li>• Complex, dynamic counterfactual scenarios are not required</li> </ul>	<ul style="list-style-type: none"> <li>• Requires articulation of conservation targets, potentially creating incentive to ‘set bar low’ to facilitate ‘business as usual’ compensatory policy (not advocated by this framework)</li> <li>• Requires estimate of the difference between the target state and current state of impacted biodiversity features</li> <li>• When targets are at odds with actions occurring or planned outside the scope of the compensatory policy, target-based actions can be suboptimal</li> <li>• Target-based ecological compensation is a relatively new concept (although similar approaches exist in some jurisdictions) and will take adjustment</li> </ul>

biodiversity. It should be implemented synergistically with other conservation and sustainable development considerations—trading up, landscape-level planning, and impacts to people (see Supporting Information 3).

A shift to the approach we propose carries risks. First, changing existing regulations, which (currently) promote averted loss offsetting, may result in sub-optimal biodiversity outcomes if the biggest gains (in the short-term) can be made by protecting highly threatened biodiversity from unmanaged pressures. Our framework deals with this by incorporating a “phased approach” (see above; Supporting Information 1). Second, having outcome-based targets places a level of accountability on those who set the target, and those who are required to contribute to its achievement. This may encourage the setting of “easy” or unambitious targets, which may lead to small compensatory requirements. This underscores the need for science-based targets that are established independently of the design of the compensatory scheme. As long as such targets exist, the simplicity of calculating compensatory requirements and the transparency of the contribution this makes to a specific goal lend itself to higher certainty for all stakeholders, and more straightforward regulatory monitoring and compliance auditing.

Operationalizing target-based ecological compensation can draw on lessons from other policy frameworks. For example, REDD+ is a mechanism under the UNFCCC where local forest protection contributes to achieving broader goals (carbon emissions targets). Challenges have been identified regarding multilevel governance, relating to accounting (e.g., carbon crediting, incentives) and implementation (e.g., decision-making; Cortez et al., 2010; Ravikumar, Larson, Duchelle, Myers, & Gonzales Tovar, 2015). This has prompted the development of implementation frameworks (e.g., “nested” approach proposed by Cortez et al., 2010), from which a key lesson is that the achievement of national targets is reliant on actors operating at multiple scales, thus necessitating protocols for their engagement, including in decision-making and benefit sharing. In light of the REDD+ experience, coordination among actors, and especially those undertaking projects “on the ground,” to contribute to the achievement of jurisdictional biodiversity targets, will be crucial for successful implementation of target-based ecological compensation.

In Brazil, requirements for the protection of a minimum proportion of native vegetation on private properties (legal reserves under the “Forest Code”) aim to help achieve bioregional vegetation retention targets. Brazil’s overall approach

has the benefit of transparency in desired outcomes, with mechanisms designed explicitly to achieve it (Metzger et al., 2019). However, criticism of its restrictiveness for business and landholders have led to relaxations of its requirements over time (e.g., amnesty for illegal deforestation on small properties; Soares-Filho et al., 2014), and even calls for it to be extinguished. This underscores the risk of implementing any environmental regulation that is reliant on contributions from industry and private individuals to achieve a broader public-good goal (e.g., explicit environmental targets).

A target-based ecological compensation approach would be most effective when developed as a coordinated jurisdictional policy, with both jurisdictional net outcomes set and Improvement/Maintenance compensation ratios calculated at the outset. The main enabling conditions (or conversely, barriers to implementation, where these conditions are lacking) for embedding the approach at the jurisdictional level include basic information on the extent/amount and condition of the biodiversity features that would be the focus of the policy, including how much is considered to be already effectively protected, and regulatory control of at least some sectors that cause biodiversity loss. Taken together, these would allow for the calculation of compensation ratios and identification of valid locations for compensation. Once this (nontrivial) work is done, the project-level process of identifying suitable ecological compensation would be greatly simplified.

In addition to government policy, most multilateral finance institutions reference “no net loss” and even “net gain” requirements in relation to escalating biodiversity risks. For example, IFC Performance Standard 6 requires no net loss where feasible in natural habitats, while net gain is required for critical habitats (IFC, 2012). The simplified ratio-based protocol that is embedded in the target-based approach could facilitate investment by these institutions, and, represents a desirable objective for those multilateral finance institutions with mandates to engage the public sector on policy reform to facilitate sustainable development.

Regardless of whether embedded in government policy or industry/corporate standards, this framework does not imply that proponents of development projects are expected to bear the entire burden of a jurisdiction achieving its particular biodiversity targets, nor that compensation alone be used to achieve targets. Indeed, the share that falls on developers is a policy decision for governments (See Results; Supporting Information 2). Fundamentally, it offers a systematic approach to determining project-level compensation that is consistent with the achievement of jurisdictional biodiversity targets. The more comprehensive the policy’s scope—that is, the more sectors that are regulated and required to compensate for losses to biodiversity arising from their activities—the greater the contribution of proponents of development to meeting a jurisdiction’s biodiversity targets.

However, it will rarely, if ever, be the case that a compensatory policy is broad enough in scope to capture all processes that result in the loss of biodiversity. This means that actors other than proponents of development projects (e.g., governments) will need to address losses to biodiversity that are beyond the scope of compensatory policy—the unregulated losses—in combination with a wide suite of other complementary conservation actions that are implemented to contribute to meeting targets. This ecological compensation framework involves setting out clearly the expectation for both proponents of development and jurisdictional authorities as this relates to how to address losses of biodiversity, whereby compensatory actions alongside other conservation investment can contribute to achieving biodiversity targets.

Ecological compensation should always be an option of last resort. In instances where the biodiversity features that are exposed to residual project losses are imperiled and irreplaceable—in other words, they cannot be feasibly improved or recreated—ecological compensation is not acceptable, and losses must be avoided altogether. Where residual losses can be reasonably addressed through compensatory interventions, this target-based framework provides a pathway toward more transparent and effective outcomes. It explicitly links compensatory actions to broader biodiversity targets, and clarifies and simplifies the expectations on and requirements of developers. In this regard, it represents a step toward the coordinated planning and integrated actions that will be crucial to stem and reverse biodiversity losses in the face of ongoing development pressures.

## ACKNOWLEDGMENTS

This research was conducted by the Compensatory Conservation working group supported by the Science for Nature and People Partnership (SNAPP), a collaboration of The Nature Conservancy, the Wildlife Conservation Society, and the National Center for Ecological Analysis and Synthesis (NCEAS). It was also supported by: COMBO Project (Agence Française de Développement, Fonds Français pour l’Environnement Mondial and the MAVA Foundation); and received funding from the Australian Government’s National Environmental Science Program through the Threatened Species Recovery Hub Project 5.1. MM is supported by ARC Future Fellowship FT140100516. LJS is supported by ARC DECRA DE170100684. G.D. and L.B. declare that they receive income from commercial contracts for consultancy services related to the development and implementation of biodiversity offsets policies.

## AUTHOR CONTRIBUTIONS

This framework was developed in a working group led by M.M. and J.E.M.W. All authors contributed to the


development of the framework. J.S.S. led the writing of the manuscript, and all authors contributed to its preparation, and approved the final version for submission.

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## REFERENCES

- Barnes, M. D., Glew, L., Wyborn, C., & Craigie, I. D. (2018). Prevent perverse outcomes from global protected area policy. *Nature Ecology & Evolution*, 2(5), 759–762. <https://doi.org/10.1038/s41559-018-0501-y>
- Bekessy, S. A., Wintle, B. A., Lindenmayer, D. B., Mccarthy, M. A., Colyvan, M., Burgman, M. A., & Possingham, H. P. (2010). The biodiversity bank cannot be a lending bank. *Conservation Letters*, 3(3), 151–158. <https://doi.org/10.1111/j.1755-263X.2010.00110.x>
- Brownlie, S., von Hase, A., Botha, M., Manuel, J., Balmforth, Z., & Jenner, N. (2017). Biodiversity offsets in South Africa—Challenges and potential solutions. *Impact Assessment and Project Appraisal*, 35(3), 248–256. <https://doi.org/10.1080/14615517.2017.1322810>
- Bull, J. W., Gordon, A., Watson, J. E. M., & Maron, M. (2016). Seeking convergence on the key concepts in ‘no net loss’ policy. *Journal of Applied Ecology*, 53(6), 1686–1693. <https://doi.org/10.1111/1365-2664.12726>
- Bull, J. W., Lloyd, S. P., & Strange, N. (2017). Implementation gap between the theory and practice of biodiversity offset multipliers. *Conservation Letters*, 10(6), 656–669. <https://doi.org/10.1111/conl.12335>
- Bull, J. W., & Strange, N. (2018). The global extent of biodiversity offset implementation under no net loss policies. *Nature Sustainability*, 1, 790–798.
- Buschke, F. T., Brownlie, S., & Manuel, J. (2017). The conservation costs and economic benefits of using biodiversity offsets to meet international targets for protected area expansion. *Oryx*, 53, 732–740. <https://doi.org/10.1017/S0030605317001521>
- Business and Biodiversity Offsets Programme (BBOP). (2009). *Biodiversity offset design handbook*. Washington, DC: BBOP.
- Business and Biodiversity Offsets Programme (BBOP). (2012). *Standard on biodiversity offsets*. Washington, DC: BBOP.
- Butchart, S. H. M., Di Marco, M., & Watson, J. E. M. (2016). Formulating smart commitments on biodiversity: Lessons from the Aichi Targets. *Conservation Letters*, 9(6), 457–468. <https://doi.org/10.1111/conl.12278>
- Carwardine, J., Klein, C. J., Wilson, K. A., Pressey, R. L., & Possingham, H. P. (2009). Hitting the target and missing the point: Target-based conservation planning in context. *Conservation Letters*, 2(1), 4–11. <https://doi.org/10.1111/j.1755-263X.2008.00042.x>
- Convention on Biological Diversity. (2010). The Strategic Plan for Biodiversity 2011–2020 and the Aichi Biodiversity Targets. Retrieved from: <https://www.cbd.int/kb/record/decision/12268>
- Cortez, R., Saines, R., Griscom, B., Martin, M., De Deo, D., Fishbein, G., ... Marsh, D. (2010). *A nested approach to REDD+*. Arlington, VA: The Nature Conservancy.
- Curran, M., Hellweg, S., & Beck, J. (2014). Is there any empirical support for biodiversity offset policy? *Ecological Applications*, 24(4), 617–632. <https://doi.org/10.1890/13-0243.1>
- Di Marco, M., Watson, J. E. M., Venter, O., & Possingham, H. P. (2016). Global biodiversity targets require both sufficiency and efficiency. *Conservation Letters*, 9(6), 395–397. <https://doi.org/10.1111/conl.12299>
- Dinerstein, E., Vynne, C., Sala, E., Joshi, A. R., Fernando, S., Lovejoy, T. E., ... Wikramanayake, E. (2019). A Global Deal For Nature: Guiding principles, milestones, and targets. *Science Advances*, 5(4), eaaw2869. <https://doi.org/10.1126/sciadv.aaw2869>
- Doherty, T. S., Bland, L. M., Bryan, B. A., Neale, T., Nicholson, E., Ritchie, E. G., & Driscoll, D. A. (2018). Expanding the role of targets in conservation policy. *Trends in Ecology & Evolution*, 33(11), 809–812. <https://doi.org/10.1016/j.tree.2018.08.014>
- Gardner, T. A., von Hase, A., Brownlie, S., Ekstrom, J. M. M., Pilgrim, J. D., Savy, C. E., ... ten Kate, K. (2013). Biodiversity offsets and the challenge of achieving no net loss. *Conservation Biology*, 27(6), 1254–1264. <https://doi.org/10.1111/cobi.12118>
- Gelcich, S., Vargas, C., Carreras, M. J., Castilla, J. C., & Donlan, C. J. (2017). Achieving biodiversity benefits with offsets: Research gaps, challenges, and needs. *Ambio*, 46(2), 184–189. <https://doi.org/10.1007/s13280-016-0810-9>
- Gibbons, P., & Lindenmayer, D. B. (2007). Offsets for land clearing: No net loss or the tail wagging the dog? *Ecological Management & Restoration*, 8(1), 26–31. <https://doi.org/10.1111/j.1442-8903.2007.00328.x>
- Gibbons, P., Evans, M. C., Maron, M., Gordon, A., Le Roux, D., von Hase, A., ... Possingham, H. P. (2016). A loss-gain calculator for biodiversity offsets and the circumstances in which no net loss is feasible. *Conservation Letters*, 9(4), 252–259. <https://doi.org/10.1111/conl.12206>
- Griggs, D., Stafford-Smith, M., Gaffney, O., Rockström, J., Öhman, M. C., Shyamsundar, P., ... Noble, I. (2013). Sustainable development goals for people and planet. *Nature*, 495, 305. <https://doi.org/10.1038/495305a>
- Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services (IPBES). (2019). Summary for policy-makers of the global assessment report on biodiversity and ecosystem services of the Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services. Retrieved from [https://www.ipbes.net/sites/default/files/downloads/spm\\_unedited\\_advance\\_for\\_posting\\_htn.pdf](https://www.ipbes.net/sites/default/files/downloads/spm_unedited_advance_for_posting_htn.pdf)
- International Finance Corporation (IFC). (2012). Performance standards. Retrieved from [https://www.ifc.org/wps/wcm/connect/Topics\\_Ext\\_Content/IFC\\_External\\_Corporate\\_Site/Sustainability-At-IFC/Policies-Standards/Performance-Standards](https://www.ifc.org/wps/wcm/connect/Topics_Ext_Content/IFC_External_Corporate_Site/Sustainability-At-IFC/Policies-Standards/Performance-Standards)

- International Union for Conservation of Nature (IUCN). (2016). IUCN policy on biodiversity offsets. Retrieved from <https://www.iucn.org/theme/business-and-biodiversity/our-work/business-approaches-and-tools/biodiversity-offsets>
- International Union for Conservation of Nature (IUCN). (2018a). Global inventory of biodiversity offset policies (GIBOP). Retrieved from <https://portals.iucn.org/offsetpolicy/>
- International Union for Conservation of Nature (IUCN). (2018b). Strengthening global biodiversity governance post-2020: Lessons from the climate regime? Retrieved from <https://www.iucn.org/files/information-paper-global-biodiversity-governance-and-lessons-climate-regime>
- Laitila, J., Moilanen, A., & Pouzols, F. M. (2014). A method for calculating minimum biodiversity offset multipliers accounting for time discounting, additionality and permanence. *Methods in Ecology and Evolution*, 5(11), 1247–1254. <https://doi.org/10.1111/2041-210X.12287>
- Mace, G. M., Barrett, M., Burgess, N. D., Cornell, S. E., Freeman, R., Grooten, M., & Purvis, A. (2018). Aiming higher to bend the curve of biodiversity loss. *Nature Sustainability*, 1(9), 448–451. <https://doi.org/10.1038/s41893-018-0130-0>
- Maron, M., Brownlie, S., Bull, J. W., Evans, M. C., von Hase, A., Quétier, F., ... Gordon, A. (2018). The many meanings of no net loss in environmental policy. *Nature Sustainability*, 1(1), 19–27. <https://doi.org/10.1038/s41893-017-0007-7>
- Maron, M., Bull, J. W., Evans, M. C., & Gordon, A. (2015). Locking in loss: Baselines of decline in Australian biodiversity offset policies. *Biological Conservation*, 192, 504–512. <https://doi.org/10.1016/j.biocon.2015.05.017>
- Maron, M., Hobbs, R. J., Moilanen, A., Matthews, J. W., Christie, K., Gardner, T. A., ... McAlpine, C. A. (2012). Faustian bargains? Restoration realities in the context of biodiversity offset policies. *Biological Conservation*, 155, 141–148. <https://doi.org/10.1016/j.biocon.2012.06.003>
- Maron, M., Simmonds, J. S., & Watson, J. E. M. (2018). Bold nature retention targets are essential for the global environment agenda. *Nature Ecology & Evolution*, 2(8), 1194–1195. <https://doi.org/10.1038/s41559-018-0595-2>
- Maxwell, S. L., Milner-Gulland, E. J., Jones, J. P. G., Knight, A. T., Bunnefeld, N., Nuno, A., ... Rhodes, J. R. (2015). Being smart about SMART environmental targets. *Science*, 347(6226), 1075–1076. <https://doi.org/10.1126/science.aaa1451>
- Metzger, J. P., Bustamante, M. M. C., Ferreira, J., Fernandes, G. W., Libran-Embid, R., Pillar, V. D., ... Overbeck, G. E. (2019). Why Brazil needs its legal reserves. *Perspectives in Ecology and Conservation*, 17(3), 91–103.
- Moilanen, A., & Kotiaho, J. S. (2018). Fifteen operationally important decisions in the planning of biodiversity offsets. *Biological Conservation*, 227, 112–120. <https://doi.org/10.1016/j.biocon.2018.09.002>
- Moilanen, A., & Laitila, J. (2016). Indirect leakage leads to a failure of avoided loss biodiversity offsetting. *Journal of Applied Ecology*, 53(1), 106–111. <https://doi.org/10.1111/1365-2664.12565>
- Moilanen, A., van Teeffelen, A. J. A., Ben-Haim, Y., & Ferrier, S. (2009). How much compensation is enough? A framework for incorporating uncertainty and time discounting when calculating offset ratios for impacted habitat. *Restoration Ecology*, 17(4), 470–478. <https://doi.org/10.1111/j.1526-100X.2008.00382.x>
- Pilgrim, J. D., Brownlie, S., Ekstrom, J. M. M., Gardner, T. A., von Hase, A., Kate, K. T., ... Ward, G. (2013). A process for assessing the offsetability of biodiversity impacts. *Conservation Letters*, 6(5), 376–384. <https://doi.org/10.1111/conl.12002>
- Quétier, F., van Teeffelen, A. J. A., Pilgrim, J. D., von Hase, A., & ten Kate, K. T. (2015). Biodiversity offsets are one solution to widespread poorly compensated biodiversity loss: A response to Curran et al. *Ecological Applications*, 25(6), 1739–1741. <https://doi.org/10.1890/14-1217.1>
- Rainey, H. J., Pollard, E. H. B., Dutson, G., Ekstrom, J. M. M., Livingstone, S. R., Temple, H. J., & Pilgrim, J. D. (2014). A review of corporate goals of No Net Loss and Net Positive Impact on biodiversity. *Oryx*, 49(2), 232–238. <https://doi.org/10.1017/S0030605313001476>
- Ravikumar, A., Larson, A. M., Duchelle, A. E., Myers, R., & Gonzales Tovar, J. (2015). Multilevel governance challenges in transitioning towards a national approach for REDD+: Evidence from 23 sub-national REDD+ initiatives. *International Journal of the Commons*, 9(2), 909–931. <https://doi.org/10.18352/ijc.593>
- Republic of South Africa. (2017). *National Environmental Management Act 1998 (ACT NO. 107 OF 1998): Draft National Biodiversity Offset Policy*.
- Republique Francaise. (2018). *Plan national d'actions 2018–2023 sur le loup et les activités d'élevage*.
- Soares-Filho, B., Rajão, R., Macedo, M., Carneiro, A., Costa, W., Coe, M., ... Alencar, A. (2014). Cracking Brazil's forest code. *Science*, 344, 363–364.
- United Nations. (2018). *The sustainable development goals report 2018*. New York, NY: United Nations.
- United Nations Environment Programme (UNEP). (2019). *National Biodiversity Strategies and Action Plans (NBSAPs)*. Retrieved from <https://www.cbd.int/nbsap/>
- Visconti, P., Butchart, S. H. M., Brooks, T. M., Langhammer, P. F., Marnewick, D., Vergara, S., ... Watson, J. E. M. (2019). Protected area targets post-2020. *Science*, 364(6437), 239–241. <https://doi.org/10.1126/science.aav6886>
- Watson, J. E. M., & Venter, O. (2017). A global plan for nature conservation. *Nature*, 550, 48. <https://doi.org/10.1038/nature24144>
- World Bank Group. (2016). *Biodiversity offsets: A user guide*. Washington D.C.: World Bank Group.

## SUPPORTING INFORMATION

Additional supporting information may be found online in the Supporting Information section at the end of the article.

**How to cite this article:** Simmonds JS, Sonter LJ, Watson JE, et al. Moving from biodiversity offsets to a target-based approach for ecological compensation. *Conservation Letters*. 2019;e12695. <https://doi.org/10.1111/conl.12695>