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Ge, Zhuang; Li, Shuangyi; Bol, Roland; Zhu, Ping; Peng, Chang; An, Tingting; Cheng, Na; Liu, Xu; Li, Tingyu; Xu, Zhiqiang; Wang, Jingkuan

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Differential long term fertilization changes residue-derived labile organic carbon fractions and microbial community during straw residue decomposition --Manuscript Draft--

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Corresponding Author:	Shuangyi Li College of Land and Environment, Shenyang Agricultural University, China Shenyang, China
First Author:	Zhuang Ge
Order of Authors:	Zhuang Ge
	Shuangyi Li
	Roland Bol
	Ping Zhu
	Chang Peng
	Tingting An
	Na Cheng
	Xu Liu
	Tingyu Li
	Zhiqiang Xu
	Jingkuan Wang
Abstract:	Straw residue amendment is a key global management strategy to achieve more sustainable agriculture. Straw residue returns affect the labile soil organic carbon (SOC) fractions and microbial community in soils. We examined temporal changes in dissolved organic carbon (DOC), microbial biomass carbon (MBC), particulate organic carbon (POC) and microbial community structure in relation to the overall straw-derived residue decomposition. The topsoil (0-20 cm) from three fertilizer management strategies (no fertilization control, CK; inorganic fertilizer, IF; inorganic fertilizer plus manure, IFM) was collected from a unique 29-year long-term field experiment (Mollisols) in Northeast China. An in-situ micro-plot incubation experiment with 13C-labeled maize straw residue (δ13C = 246.9‰) (i.e., no fertilization control + straw, CKS; inorganic fertilizer + straw, IFS; inorganic fertilizer plus manure + straw, IFMS) and without straw residue was conducted. We used the 13C isotope technique to measure the content of residue-derived labile SOC fractions and used high-throughput sequencing-based amplification of bacterial 16S rRNA and fungal internal transcribed spacer (ITS) rRNA to evaluate the dynamic changes of bacterial and fungal community structure and diversity at different times (the 1st day, 60th day and 150th day) under all six treatments. We found that residue-derived POC was significantly increased, but residue-derived MBC content was higher in the fertilized (IFS and IFMS) compared to the unfertilized (CKS) treatment. The Linear discriminant analysis Effect Size (LEfSe) revealed changes after adding straw of soil microbes in CK and IF were significantly higher than in IFM soil. Network analysis showed that straw residue addition decreased the bacterial microbial network complexity in all treatments, but increased fungal network complexity in IFS and IFMS). Proteobacteria (IFMS) were keystone taxa in the bacterial community, and Ascomycota was the keystone taxon in the fungal community. The straw residue was r

	POC in labile SOC fractions and was further enhanced in fertilizer management with manure addition. Less pronounced effects on the microbial community following straw addition were found for the combined inorganic and organic fertilizer treatment compared to those without or with only inorganic fertilizer addition. The observed temporal changes in the microbial community suggested that in these Mollisols, independent of agricultural fertilizer management, straw residue-derived POC and DOC promoted fungal C processing, whereas for bacterial this was facilitated only via residue-derived MBC.
Suggested Reviewers:	Anning Zhu anzhu@issas.ac.cn He is an expert on this topic.
	Jie Zhuang jzhuang@utk.edu He did this very well in the related feild.
	Hongtu Xie xieht@iae.ac.cn He is an expery in the related field.
	Sean Schaeffer sschaef5@utk.edu He is active doing this type of research.
	Lujun Li lilujun@iga.ac.cn He does this research field.

Cover Letter

Dear Prof. Noellemeyer,

We would like to submit our manuscript 'Differential long term fertilization changes residue-derived labile organic carbon fractions and microbial community during straw residue decomposition' prepared by Zhuang Ge, Shuangyi Li, Roland Bol, Ping Zhu, Chang Peng, Tingting An, Na Cheng, Xu Liu, Tingyu Li, Zhiqiang Xu, Jingkuan Wang for evaluation as Research Article in the journal *Soil & Tillage Research*.

The work in this paper can be summarized as follows, the topsoil (0-20 cm) from three fertilizer management strategies (no fertilization control, CK; inorganic fertilizer, IF; inorganic fertilizer plus manure, IFM) was collected from a unique 29-year long-term field experiment (Mollisols) in Northeast China. An in-situ micro-plot incubation experiment with ¹³C-labeled maize straw residue ($\delta^{13}C = 246.9\%$) (i.e., no fertilization control + straw, CKS; inorganic fertilizer + straw, IFS; inorganic fertilizer plus manure + straw, IFMS) and without straw residue was conducted. We used the ¹³C isotope technique to measure the content of residue-derived labile SOC fractions and used high-throughput sequencing-based amplification of bacterial 16S rRNA and fungal internal transcribed spacer (ITS) rRNA to evaluate the dynamic changes of bacterial and fungal community structure and diversity at different times (the 1st day, 60th day and 150th day) under all six treatments. We found that residue-derived POC was significantly increased, but residue-derived DOC was significantly decreased during straw residue decomposition. The residue-derived MBC content was higher in the fertilized (IFS and IFMS) compared to unfertilized (CKS) treatment. The Linear discriminant analysis Effect Size (LEfSe) revealed changes after adding straw of soil microbes in CK and IF were significantly higher than in IFM soil. Network analysis showed that straw residue addition decreased the bacterial microbial network complexity in all treatments, but increased fungal network complexity in IFS and IFMS. The random forest model predicted that during straw decomposition, Chloroflexi (CKS), Actinobacteria (IFS), Proteobacteria (IFMS) were keystone taxa in the bacterial community, and Ascomycota was the keystone taxon in the fungal community. The straw residue was retained as POC in labile SOC fractions and was further enhanced in fertilizer management with manure addition. Less pronounced effects on the microbial community following straw addition were found for the combined inorganic and organic fertilizer treatment compared to those without or with only inorganic fertilizer addition. The observed temporal changes in the microbial community suggested that in these Mollisols, independent of agricultural fertilizer management, straw residue-derived POC and DOC promoted fungal C processing, whereas for bacterial this was facilitated only via residue-derived MBC.

This manuscript has neither been published and or is under consideration for publication elsewhere. We have no conflicts of interest to disclose. All authors have read and approved the final version of the manuscript. Thank you

in advance the consideration of our manuscript, and we look forward to hearing from you at your earliest convenience.

Yours sincerely,

Shuangyi Li

Associate Professor of Soil Science, College of Land and Environment, Shenyang Agricultural University, Shenyang, Liaoning 110866, China. Tel: +86(24)88487155. E-mail: shy li@syau.edu.cn.

Roland Bol

Professor of Soil Science, Institute of Bio- and Geosciences, Agrosphere (IBG-3), Forschungszentrum Jülich GmbH, Jülich, Germany. Tel: +49 2461 61-6653. E-mail: r.bol@fz-juelich.de

- Straw addition decreased the bacterial network complexity in all treatments
- Straw addition increased fungal network complexity in fertilized soils
- Key species in straw decomposition changed under different fertilizer regimes
- Residue-derived POC and DOC increased fungal carbon processing
- Bacterial carbon processing was facilitated only via residue-derived MBC

1 2 3	Differential long term fertilization changes residue-derived labile organic carbon fractions and microbial community during straw residue decomposition
4	Zhuang Ge ^{a, b} , Shuangyi Li ^{a*} , Roland Bol ^{b,c*} , Ping Zhu ^d , Chang Peng ^d , Tingting An ^a ,
5	Na Cheng ^a , Xu Liu ^a , Tingyu Li ^a , Zhiqiang Xu ^e , Jingkuan Wang ^a
6	^a Northeast Key Laboratory of Conservation and Improvement of Cultivated Land
7	(Shenyang), Ministry of Agriculture, College of Land and Environment, Shenyang
8	Agricultural University, Shenyang, Liaoning 110866, China
9	^b Institute of Bio- and Geosciences, Agrosphere (IBG-3), Forschungszentrum Jülich
10	GmbH, Wilhelm-Johnen-Straße, 52428 Jülich, Germany
11	^c Environment Centre Wales, Deiniol Road, Bangor University, LL57 2UW Bangor,
12	Gwynedd, UK
13	^d Jilin Academy of Agricultural Sciences, Gongzhuling, Jilin 136100, China
14	^e Liaoning Agricultural Development Service Center, Shenyang 110034, China
15	
16	*Corresponding author:
17	Shuangyi Li (Tel: +86(24)88487155; E-mail address: shy_li@syau.edu.cn; Address:
18	No.120 Dongling Road, Shenhe District, College of Land and Environment, Shenyang
19	Agricultural University, Shenyang, Liaoning 110866, China)
20	Roland Bol (Tel: +49 2461 61-6653; E-mail address: r.bol@fz-juelich.de; Address:
21	Institute of Bio- and Geosciences, Agrosphere (IBG-3), Forschungszentrum Jülich
22	GmbH, Wilhelm-Johnen-Straße, 52428 Jülich, Germany)

23 Abstract:

24 Straw residue amendment is a key global management strategy to achieve more 25 sustainable agriculture. Straw residue returns affect the labile soil organic carbon (SOC) fractions and microbial community in soils. We examined temporal changes in 26 27 dissolved organic carbon (DOC), microbial biomass carbon (MBC), particulate organic carbon (POC) and microbial community structure in relation to the overall straw-28 29 derived residue decomposition. The topsoil (0-20 cm) from three fertilizer management strategies (no fertilization control, CK; inorganic fertilizer, IF; inorganic fertilizer plus 30 31 manure, IFM) was collected from a unique 29-year long-term field experiment 32 (Mollisols) in Northeast China. An *in-situ* micro-plot incubation experiment with ¹³Clabeled maize straw residue ($\delta^{13}C = 246.9\%$) (i.e., no fertilization control + straw, CKS; 33 34 inorganic fertilizer + straw, IFS; inorganic fertilizer plus manure + straw, IFMS) and without straw residue was conducted. We used the ¹³C isotope technique to measure the 35 36 content of residue-derived labile SOC fractions and used high-throughput sequencing-37 based amplification of bacterial 16S rRNA and fungal internal transcribed spacer (ITS) rRNA to evaluate the dynamic changes of bacterial and fungal community structure and 38 diversity at different times (the 1st day, 60th day and 150th day) under all six treatments. 39 40 We found that residue-derived POC was significantly increased, but residue-derived

41	DOC was significantly decreased during straw residue decomposition. The residue-
42	derived MBC content was higher in the fertilized (IFS and IFMS) compared to the
43	unfertilized (CKS) treatment. The Linear discriminant analysis Effect Size (LEfSe)
44	revealed changes after adding straw of soil microbes in CK and IF were significantly
45	higher than in IFM soil. Network analysis showed that straw residue addition decreased
46	the bacterial microbial network complexity in all treatments, but increased fungal
47	network complexity in IFS and IFMS. The random forest model predicted that during
48	straw decomposition, Chloroflexi (CKS), Actinobacteria (IFS), Proteobacteria (IFMS)
49	were keystone taxa in the bacterial community, and Ascomycota was the keystone taxon
50	in the fungal community. The straw residue was retained as POC in labile SOC fractions
51	and was further enhanced in fertilizer management with manure addition. Less
52	pronounced effects on the microbial community following straw addition were found
53	for the combined inorganic and organic fertilizer treatment compared to those without
54	or with only inorganic fertilizer addition. The observed temporal changes in the
55	microbial community suggested that in these Mollisols, independent of agricultural
56	fertilizer management, straw residue-derived POC and DOC promoted fungal C
57	processing, whereas for bacterial this was facilitated only via residue-derived MBC.

59	Keywords: Labile soil organic carbon; Soil microbial community; Soil microbial
60	network; Key species; High-throughput sequencing; ¹³ C-labelling technique

61

62 **1 Introduction**

Soil organic carbon (SOC) plays a key role in chemical, physical, and biological 63 64 properties and processes in soils, thus it is considered to be a crucial factor to affect soil 65 quality (Wiesmeier et al., 2019). Fertilization is an important agricultural practice that 66 aims to improve soil quality and plant nutrition for increasing crop yield (Chen et al., 67 2014; Xiang et al., 2020). Application of inorganic fertilizer could temporarily slow 68 down food shortage but would lead to soil degradation, soil acidification, and SOC 69 depletion (Stockmann et al., 2013; Zamanian and Kuzyakov, 2019). Organic fertilizer 70 application promoted SOC accumulation, alleviated soil acidification, increased crop 71 yield, therefore would be an alternative option to applicate inorganic fertilizer (Afreh 72 et al., 2018; Han et al., 2016). Moreover, SOC is highly heterogeneous, according to 73 the turnover rates of various SOC fractions, it could be divided into stable and labile 74 fractions (Bol et al., 2009; Six et al., 2002). Labile SOC fractions generally include 75 dissolved organic carbon (DOC), microbial biomass carbon (MBC), and particulate 76 organic carbon (POC) (Franzluebbers et al., 2000; Jardine et al., 1989; Moore et al.,

77	2000). Labile SOC fractions had relatively short turnover time (weeks to months), they
78	are sensitively responding to agricultural practice (Yan et al., 2007), and are an easily
79	available and important source of energy for soil microbes (Kaye and Hart, 1997;
80	McLauchlan and Hobbie, 2004). Labile SOC pools have been considered as early
81	sensitive indicators of soil quality which affect soil function in special ways due to their
82	different fractions (Blanco-Moure et al., 2016; Rudrappa et al., 2006). The stable
83	isotope technique provided a good way to quantitatively investigate the changes of
84	labile SOC fractions, that results would bring more direct and clear theoretical support
85	for the management of SOC (Amelung et al., 2008). Different fertilizer management
86	strategies also affected SOC pools. For example, fertilization strategies which include
87	organic manure can increase the pool of stable C in the surface soil layer and increase
88	concentrations and proportions of labile C (Li et al., 2018), but apply the different rates
89	of inorganic fertilizer would not affect the concentrations of soil labile fractions (Naylor
90	et al., 2020). Furthermore, previous studies reported that the application of fertilizer
91	affected the diversity and composition of the soil microbial community (Xiang et al.,
92	2020; Zhou et al., 2015). Microbial communities play a key role in soil organic matter
93	(SOM) transformation and nutrient cycling in agricultural soils (Gattinger et al., 2007).
94	Long-term applicated inorganic fertilizer reduced the biodiversity and abundance of

95	bacteria, but manure application increased bacterial abundance and diversity (Cui et al.,
96	2018; Zhou et al., 2015). Long-term application of inorganic fertilizer plus pig manure
97	altered fungal community composition by increasing the abundance of Pezizales. (Ye
98	et al., 2020). The number of differential populations in bacterial communities in applied
99	manure soils was significantly higher than that in applied inorganic fertilizer soils,
100	whereas those of fungal communities showed the opposite trend, therefore, the
101	responses of different fertilizers to soil bacterial and fungal communities were various
102	(Pan et al., 2020).

103 Crop residues were assumed to be the dominant carbon (C) source to the arable soil and primary control of the accumulation of SOC (Clemmensen et al., 2013). Therefore 104 105 crop residue return is believed to be an effective promising approach to improve soil quality, promote agricultural SOC storage, and mitigate climate change (Dikgwatlhe et 106 107 al., 2014; Lal, 2004; Liu et al., 2014). Straw returns also could affect the SOC pool, 108 especially the labile SOC fractions (Lei et al., 2010; Zhao et al., 2016). Different 109 fertilizer management strategies would influence the decomposition of residue-derived C. Our previous studies showed apply nitrogen fertilization increased the content of 110 111 residue-derived DOC and decreased the content of residue-derived MBC, apply manure increased the content of residue-derived MBC, apply manure combined with inorganic 112

113	fertilizer significantly increased residue-derived POC in Alfisols (Jin et al., 2020; Jin et
114	al., 2018; Wang et al., 2020). Meanwhile, labile SOC is also an essential source in which
115	microorganisms can receive nutrients and energy from it, and it affects the life activities
116	of soil microorganisms. Soil microorganisms simultaneously participate in the
117	processes of labile SOC fractions formation, transformation, and decomposition
118	(Chantigny, 2003). Furthermore, the decomposition of crop residues in the soil is a
119	complex biogeochemical process, and microorganisms are the major drivers of crop
120	residue decomposition and turnover in soils (Marschner et al., 2011). The straw residue
121	is metabolized by soil microbes, transformed into microbial biomass, and becomes
122	stabilized as SOM (Cotrufo et al., 2015). Added crop residue to the soils may greatly
123	influence soil microbial diversity, abundance, and composition thus altering SOC
124	dynamics (Goldstein et al., 2020). Soil bacterial and fungal communities play important
125	roles in the decomposition of SOM and provide available nutrients for plant growth
126	(Allison et al., 2007). Crop residue decomposition usually proceeds through a series of
127	well-characterized stages involving a succession of soil bacterial and fungal
128	communities (Guo et al., 2018). Due to soil bacterial and fungal communities exhibit
129	different dynamic patterns, the microbial succession would lead to the changes of
130	specific microbial taxa during the straw residue decomposition process (Gao et al., 2016;
	7

131	Prewitt et al., 2014). Therefore, it is essential to simultaneously detect the dynamic
132	changes of differential populations in soil bacterial and fungal communities in each
133	stage in the process of straw residue decomposition under different fertilizer regimes
134	(Baldrian et al., 2012; Banerjee et al., 2016). In addition, the quantitative relationship
135	between the different residue-derived labile carbon fractions and soil microbes during
136	the process of straw decomposition are still unclear.
137	Interactions between microbes are also important aspects in maintaining a diverse
138	microbial community, at the same time, network analyses provide a very useful tool to
139	explore the co-occurrence patterns and reflecting the microbial community in
140	ecosystems (Faust and Raes, 2012; Rottjers and Faust, 2018). Exploring changes in soil
141	microbial community networks can increase our knowledge of the complexity and
142	diversity of microbial communities (Mora-Montes et al., 2010). The application of the
143	organic amendment significantly changed the bacterial and fungal community's
144	network than inorganic fertilizer and therefore fertilization could influence the
145	synergistic interactions between species. (Ling et al., 2016; Xue et al., 2017). However,
146	it remains unclear on exploring the effect of straw residue decomposition to soil
147	bacterial and fungal co-occurrence networks under different fertilizer regimes. Besides,
148	the information on the key role of specific microbial species of soil bacterial and fungal 8

149 in the process of straw residue decomposition is still limited.

150	The North China Plain is one of the most important food production regions in
151	China (Liu et al., 2019). However, in the past several decades, long-term unreasonable
152	field management degraded the soil and deteriorated the soil quality (Liu et al., 2010).
153	Return of straw residue within overall fertilizer management strategies is widely used
154	to mitigate these problems and enhance the storage of SOC in this region (Qiu et al.,
155	2016). Most of the studies only reported the effect of long-term fertilizer or straw
156	incorporation on soil microbial communities (Guo et al., 2020; Pan et al., 2020; Zhao
157	et al., 2019). The microbial community composition and metabolic activity correspond
158	to different fertilizer regimes, and soil fertility affected the residue decomposition
159	process (An et al., 2015b; Zhang et al., 2019). Therefore, deeply revealing the dynamic
160	response mechanisms of the key role in the microbial community during the process of
161	straw residue decomposition in integrated fertility management is crucial for enhancing
162	C sequestration. In this study, we collected arable soils of three fertilizer management
163	strategies from a 29-year fertilization experiment in the North China Plain. We
164	conducted an <i>in-situ</i> micro-plot incubation experiment with and without maize straw
165	residue in these soils to investigate the changes of residue-derived labile SOC fractions
166	content and complex temporal responses of soil microbial communities to straw residue

167	decomposition, as well as their quantitative relationships under various fertilization
168	regimes. We hypothesized that under the varying fertilizer regimes in Mollisols: (i)
169	Straw residue-derived content of labile SOC fractions would differ between specific
170	management strategies, (ii) Temporal shifts in soil bacterial and fungal population
171	abundances and keystone decomposer species occur during straw decomposition, (iii)
172	Soil bacterial and fungal network complexity would change following straw
173	incorporation, (iv) Different residue-derived labile carbon fractions have varying
174	effects on soil fungal and bacterial population.

175 **2 Materials and methods**

176 **2.1 Study site description**

A long-term field experiment site used in this study was located at Jilin Academy 177 of Agricultural Sciences at Gongzhuling County, Jilin Province, Northeast China (43° 178 30'N, 124° 48'E, and 200 m above sea level). The experiment was established in 1990, 179 180 has a typical continental monsoon climate with mean annual temperatures of 4-5 °C and 181 mean annual precipitation of 400-600 mm (Song et al., 2015). The soil is a Mollisol (classified as a Luvic Phaeozem, FAO) with 39% sand, 30% silt, and 31% clay at the 182 beginning of the experiment (Xie et al., 2014). The three application were selected in 183 this study: (1) unfertilized control (CK), (2) balanced inorganic fertilizers at 165 kg N 184

ha⁻¹,82.5 kg P₂O₅ ha⁻¹, and 82.5 kg K₂O ha⁻¹ (IF), (3) balanced inorganic fertilizers at
50 kg N ha⁻¹, 82.5 kg P₂O₅ ha⁻¹, and 82.5 kg K₂O ha⁻¹ plus manure at 115 kg N ha⁻¹
(IFM) (Dou et al., 2016). The manure was pig manure and applied in autumn after corn
harvesting in the IFM plots each year (Song et al., 2015). The basic soil properties of
each treatment are provided in Table S1.

190 2.2 *In-situ* field experiment design

191 The micro-plot experiment was not undertaken at the main long-term field-site itself, but in a nearby field to avoid any presence of straw influencing future soil. Two 192 soil pits of the following dimensions (length \times width \times height = 1.0 m \times 0.6 m \times 0.3 m) 193 were therefore dug in a nearby field for the micro-plot experiment. Two polyvinyl 194 195 chloride (PVC) material boxes (length \times width \times height = 1.0 m \times 0.6 m \times 0.6 m) of 196 similar dimensions to the pit were then inserted vertically into field pits on May 5, 2018, i.e., the boxes were 0.3 m above the ground level to avoid any impacts by other soil in 197 198 the field. The boxes were not closed at the bottom to allow for drainage. Each box 199 consisted of 9 equal sections, allowing 3 random replicates of the 3 treatments under consideration (CK, IF and IFM). The topsoil layer (0-20 cm) was taken from each 200 201 fertilization treatment of the long-term field experiment and individually passed in the 202 field through a 7 mm sieve to remove crop roots and rocks. More details of soil

219 220	2.3 Measurements of dissolved organic carbon (DOC) and microbial biomass carbon (MBC)
218	kept at -80 °C for microbial analysis.
217	stored in the 4°C incubator and transported to the laboratory, then the samples were
216	and October 2, 2018 (the 150 th day). Soil samples were sealed in plastic bags, then
215	depth of 0-20cm three times: on May 6, 2018 (the 1 st day), July 4, 2018 (the 60 th day),
214	the other box only contained soil (CK, IF, IFM). Soil samples were collected at the
213	In one box all 9 compartments consisted of the soil mixed with straw (CKS, IFS, IFMS),
212	the soils from the three treatments were mixed homogenously with ¹³ C-labelled straw.
211	depth) in each section box. Subsequently, the upper 20 cm was filled with soil, part of
210	sections, we first added only soil from three treatments to the bottom 10 cm (20-30 cm
209	grown in all the boxes during the experimental period. Before completely filling the
208	straw to 15.84 kg of soil per section (equivalent to 2.3 g straw kg ⁻¹ soil). No plants were
207	i.e., where all straw residue after harvest is plowed back into the soil. We added 36 g
206	The method of straw incorporation was based on the concept of full straw incorporation,
205	procedure (An et al., 2015a). The maize straw residue was cut in the size of 0.5-1.0 cm.
204	plants pulse-labeled using ¹³ CO ₂ four times over a growing season according to the
203	properties can be found in Table S1. The ¹³ C-labelled maize straw was mature maize

MBC was determined by the chloroform-fumigation extraction method (Vance et $^{12}\,$ 221

222	al., 1987). Fresh soil equivalent to 10 g of oven-dried soil was fumigated for 24 h at
223	25 °C and subsequently extracted with 100 ml of 0.5 mol L ⁻¹ K ₂ SO ₄ . The same amount
224	of un-fumigated soil was extracted with 100 ml of 0.5 mol L ⁻¹ K ₂ SO ₄ . The non-
225	fumigated extract was used to determine DOC. The contents of organic C of soil
226	extracts were determined by the Total Organic Carbon Analyzer (Multi N/C 3100 TOC,
227	Germany). MBC was calculated as the difference in organic C content between
228	fumigated and un-fumigated soil extracts with a correction factor ($k_{\rm EC}$) of 0.45 (Wu et
229	al., 1990). All K ₂ SO ₄ extracts were freeze-dried before further analysis of ¹³ C
230	abundance. Soil and K ₂ SO ₄ extracts samples were analyzed for total C and δ^{13} C values
231	with an elemental analyzer (Elementar Vario PYRO cube, Germany) coupled to an
232	isotope ratio mass spectrometer (IsoPrime 100 Isotope Ratio Mass Spectrometer,
233	Germany).

234 **2.4 Measurements of particle organic carbon (POC)**

POC was isolated from bulk soil using the procedure (Cambardella and Elliott, 1992). A 10 g subsample of bulk soil was passed through a 2 mm sieve, dispersed into 30 ml of 5 g L⁻¹ sodium hexametaphosphate, and shaken for 5 h. Next, the suspension was filtered through a 53 μ m sieve. The material remaining on the sieve (POC) was rinsed thoroughly with deionized water, dried at 50 °C for 24 h weighed, and stored 240 before measurement. Organic C and δ^{13} C in POC were determined using the EA-IRMS.

The Content of MBC was calculated with the following equation (Vance et al.,1987):

244
$$C_{MBC} = (C_f - C_{nf}) / k_{EC}$$
 (1)

where C_f and C_{nf} refer to the amount of dissolved organic C (mg kg⁻¹ soil) from the fumigated and the nonfumigated K₂SO₄ extracts, respectively. The k_{EC} value was used to convert measured data into biomass C, in this study we used it as 0.45 (Joergensen, 1996).

249 The δ^{13} C of MBC (‰) was calculated with the following equation (Engelking et 250 al., 2008):

251
$$\delta^{13}C_{MBC} = \left(\delta^{13}C_f \times C_f - \delta^{13}C_{nf} \times C_f\right) / \left(C_f - C_{nf}\right)$$
(2)

where C_f and C_{nf} refer to the amount of dissolved organic C (mg kg⁻¹ soil) from the fumigated and the nonfumigated K₂SO₄ extracts, respectively, and $\delta^{13}C_f$ and $\delta^{13}C_{nf}$ refer to the $\delta^{13}C$ values (‰) of the fumigated and the nonfumigated K₂SO₄ extracts, respectively.

256 The proportion of MBC derived from maize straw C in total MBC (f_{MBC}) was

257 calculated according to the following equation (De Troyer et al., 2011):

258
$$f_{MBC} = (\delta^{13}C_{MBC} - \delta^{13}C_{MBC-WS}) / (\delta^{13}C_{straw} - \delta^{13}C_{MBC-WS})$$
 (3)259where $\delta^{13}C_{MBC}$, $\delta^{13}C_{MBC-WS}$, and $\delta^{13}C_{straw}$ are the $\delta^{13}C$ values of MBC from whole soil260samples with straw, MBC from whole soil samples without straw (WS), and ^{13}C -261labelled maize straw itself, respectively.262The content of straw-derived MBC ($^{13}C_{MBC}$) was calculated with the following equation263(Blaud et al., 2012):264 $^{13}C_{MBC} = C_{MBC} \times f_{MBC}$ (4)265where CMBC denotes the content of MBC, and f_{MBC denotes the proportion of MBC266derived from maize straw C in total MBC.267The proportion of DOC derived from maize straw C in total DOC (fboc) was268calculated according to the following equation (De Troyer et al., 2011):269 $f_{DOC} = (\delta^{13}C_{DOC} - \delta^{13}C_{DOC-WS}) / (\delta^{13}C_{straw} - \delta^{13}C_{DOC-WS})$ (4)270Where $\delta^{13}C_{DOC}$, $\delta^{13}C_{DOC-WS}$, and $\delta^{13}C_{straw}$ are the $\delta^{13}C$ values of DOC from soil samples

itself, respectively.

273 The content of straw-derived DOC (C_{DOC}) was calculated using the following 274 equation (Blaud et al., 2012):

$$^{13}C_{DOC} = C_{DOC} \times f_{DOC}$$
(5)

where C_{DOC} denotes the content of DOC, and f_{DOC} denotes the proportion of DOC

- 277 derived from maize straw C in total DOC.
- 278 The proportion of POC derived from maize straw C in total POC (fPOC) was
- calculated according to the following equation (De Troyer et al., 2011):

280
$$f_{POC} = (\delta^{13}C_{POC} - \delta^{13}C_{POC-WS}) / (\delta^{13}C_{straw} - \delta^{13}C_{POC-WS})$$
 (6)

281 Where $\delta^{13}C_{POC}$, $\delta^{13}C_{POC-WS}$, and $\delta^{13}C_{straw}$ are the $\delta^{13}C$ values of POC from soil samples 282 with straw, POC from soil samples without straw (WS), and ¹³C-labelled maize straw 283 itself, respectively.

284 The content of straw-derived POC (CPOC) was calculated using the following 285 equation (Blaud et al., 2012)

$$^{13}C_{POC} = C_{POC} \times f_{POC}$$
(7)

where C_{POC} denotes the content of POC, and f_{POC} denotes the proportion of POC
derived from maize straw C in total POC.

All results are shown as the mean of the three plot replicates with standard error. Significant differences between treatments (P < 0.05) were calculated by one-way analysis of variance (ANOVA) in combination with Duncan's test using SPSS 19.0 and OriginPro 2019.

293 **2.6 High-throughput sequencing**

294 Soil total DNA was extracted from 0.5 g of soil using the MP FastDNA SPIN Kit for Soil (MP Biomedicals) according to the manufacturer's instructions. The final DNA 295 concentration and purification were quantified by NanoDrop 2000 UV-vis 296 297 spectrophotometer (Thermo Scientific, Wilmington, USA), and DNA quality was checked by 1% agarose gel electrophoresis. We performed Polymerase chain reaction 298 299 (PCR) amplification of the prokaryotic 16S rRNA V3-V4 hypervariable region with 300 primers set 338F/806R (338F, 5'-ACTCCTACGGGAGGCAGCAG-3' and 806R, 5'-GGACTACHVGGGTWTCTAAT-3') (Lee et al., 2012) and fungal ITS1 region with 301 302 primers set ITS1F/ITS2R (ITS1F, 5'-CTTGGTCATTTAGAGGAAGTAA-3', ITS2R, 303 5'-GCTGCGTTCTTCATCGATGC-3') (Adams et al., 2013) primer pairs. The PCR reactions were conducted using the following program: 3 min at 95°C; 27 cycles of 30 304 305 s at 95°C, 30 s at 55°C, and 45 s at 72°C; and 72°C for 10 min. The bacterial PCR 306 amplification was performed in triplicate 20- μ L mixtures containing 4 μ L of 5 × FastPfu

307	Buffer, 2 μ L of 2.5 mM dNTPs, 0.8 μ L of each primer (5 μ M), 0.4 μ L of FastPfu
308	polymerase, 0.2 μ L of BSA, and 10 ng of template DNA. The fungal PCR amplification
309	was performed in triplicate 20-µL mixtures containing 2 µL of 10 × Buffer, 2 µL of 2.5
310	mM dNTPs, 0.8 μ L of each primer (5 μ M), 0.2 μ L of rTaq Polymerase, 0.2 μ L of BSA,
311	and 10 ng of template DNA. The products were extracted from 2% agarose gels,
312	purified using the AxyPrep DNA Gel Extraction Kit (Axygen Biosciences, Union City,
313	CA, USA), and quantified using QuantiFluor TM -ST (Promega, Madison, WI, USA)
314	according to the manufacturer's protocol. Purified amplicons were pooled in equimolar
315	concentration and paired-end sequenced (2 \times 300) on an Illumina platform (Illumina,
316	San Diego, CA, USA) according to the standard protocols by Majorbio Bio-Pharm
317	Technology Co. Ltd. (Shanghai, China). Sequences were deposited in the Sequence
318	Read Archive (SRA) of the National Center for Biotechnology Information (NCBI;
319	Bethesda, MD, USA) under studies PRJNA644514 and PRJNA644592 for the 16S
320	rRNA and ITS rRNA genes, respectively.

321 **2.7 Bioinformatics and statistics analysis**

The sequences were processed using USEARCH v.10.0 (Edgar, 2010) and VSEARCH v.2.7.1 (Rognes et al., 2016). The paired-end Illumina reads processed in the following steps by VSEARCH: joining of paired-end reads and relabeling of

325	sequencing names; removal of barcodes and primers; filtering of low-quality reads;
326	finding non-redundancy reads. Unique reads were clustered into operational taxonomic
327	units (OTUs) with 97% similarity. The representative sequences were picked by
328	UPARSE (Edgar, 2013). The OTU table was generated by USEARCH. We used the
329	RDP database (Wang et al., 2007) for the taxonomic identity of each bacteria and the
330	UNITE database (Nilsson et al., 2018) for the taxonomic identity of fungi by
331	USEARCH. A total of 2,516,669 16S rRNA and 3,211,166 ITS paired-end high-quality
332	sequences were obtained from all of the 54 soil samples, respectively. The α -diversity
333	index of Shannon and Chao 1 was calculated by USEARCH software. Each sample had
334	a different number of reads, to even the varying read numbers, all the samples were
335	randomly re-sampled to the lowest read number. Therefore, 17229 and 33746 randomly
336	selected 16S rRNA and ITS sequences were used to calculate bacterial and fungal α -
337	diversity.

Statistical tests and graphical representations were carried out in the R environment (version 3.6.1; https://www.r-project.org/). Differences in the microbial α diversity were compared using analysis of variance (ANOVA) followed by Tukey's HSD test. *P*<0.05 was considered to reflect a statistically significant difference. The linear discriminant analysis effect size (LEfSe) method was conducted to identify different abundance values (at all taxonomic levels) that reflected the different
fertilization treatments in each incubation time. Only taxa meeting an LDA significance
threshold of 2 for bacterial and fungal communities are shown.

Co-occurrence network analysis (based on class level) was performed to examine the connections within bacterial and fungal taxa. Relative abundances of bacterial and fungal genera were used to construct networks for bacterial and fungal classes in different treatments during the whole process of straw residue decomposition. Average relative abundances of the class level higher than 0.1% were selected for Spearman's correlation analysis. Network analysis was then performed using the "igraph" package in R v. 3.6.1 (Rottjers and Faust, 2018).

353 To obtain the best discriminant performance of the taxa across the whole incubation time of straw residue decomposition in different fertilizer managements, we 354 355 regressed the relative abundances of the bacterial and fungal taxa at the order level 356 against the incubation time of straw residue decomposition in the field using the randomForest package (version 4.6-14) in R using default parameters (ntree = 1000, 357 mtry is p/3, where p is the number of taxa in the class) (Liaw and Wiener, 2002). Lists 358 359 of taxa ranked by Random Forests in order of feature importance were decided more than 100 iterations. The 10-fold cross-validation was performed by using the rfcv () 360 20

361 function in the randomForest package for selecting appropriate features with five362 repeats.

363	The structural equation model (SEM) was obtained using Amos 17.0 software
364	(Chicago, IL: Amos Development Corporation) to quantify the significant effects of
365	residue-derived labile SOC fractions on the changes of soil bacterial and fungal
366	community structure in response to different fertilizer management strategies and
367	incubation time. Three fertilizer treatments (CKS, IFS, IFMS) were categorical
368	variables with three levels: 0 (CKS), 1 (IFS), and 2 (IFMS). We used the robust
369	maximum likelihood estimation to fit the covariance matrix to the model (Wang et al.,
370	2016). The theoretical model was adjusted according to the principle of the low Chi-
371	square (χ^2 ; $0 \le \chi^2 \le 2$), nonsignificant probability (<i>P</i> ; <i>P</i> > 0.05), high goodness-of-fit-
372	index (GFI; GFI > 0.90) and root mean square error of approximation (RMSEA;
373	RMSEA < 0.05) to ensure that the final model was adequately fitted (Grace and Keeley,
374	2006).

375 **3 Results**

376 3.1 Incorporation of residue-derived labile SOC fractions under different fertilizer 377 treatments

378	The content of straw derived active SOC in three treatments occurred various trends
379	(Fig. 1). The content of straw-derived DOC was significantly decreased by 43% (CKS),
380	50% (IFS) and 77% (IFMS) from the 1^{st} day to the 60 th day, respectively (Fig. 1a). On
381	the 150 th day, that content in IFMS treatment was 0.49 mg kg ⁻¹ , significantly lower than
382	in CKS (2.82 mg kg ⁻¹) and IFS (4.30 mg kg ⁻¹) treatment. The changes in the content of
383	straw-derived MBC over time were different in three treatments (Fig. 1b). The contents
384	of straw-derived MBC in IFS treatment are 59 mg kg ⁻¹ (the 1 st day), 60 mg kg ⁻¹ (the 60 th
385	day) and 41 mg kg ⁻¹ (the 150 th day), which were the highest in three treatments during
386	the whole incubation time. At the end of incubation time, the order of the content of
387	straw-derived MBC in three treatments is IFS > IFMS > CKS. The content of straw-
388	derived POC was significantly increased during the incubation time in three treatments
389	(Fig. 1c). On the 150 th day, straw-derived POC in IFMS treatment was 198 mg kg ⁻¹ ,
390	which was twice more than in CKS and 42% more than in IFS treatment.
391	3.2 Soil bacterial and fungal community diversity and structure

392 Different fertilizer management significantly changed the α -diversity of the 393 bacterial community. During straw residue decomposition, α -diversity indices (Chao1 394 and Shannon) in IFM was the highest, the α -diversity in IF was the lowest in the 395 bacterial community (Fig. S1). However, there was no significant difference in the α - diversity indices (Chao1 and Shannon) of the fungal community among three fertilizermanagement during straw residue decomposition (Fig. S2).

398	The most abundant bacterial and fungal phyla in all soil samples were shown in
399	Fig. S3a and Fig. 2a. During the incubation time, Actinobacteria, Acidobacteria, and
400	Proteobacteria were the dominant phyla in all soil samples. There was no significant
401	difference in the relative abundance of bacterial phylum level in each treatment after
402	adding straw residue (Fig. S3a). With regards to the fungal communities, the relative
403	abundance of Ascomycota was significantly decreased from 45.0% (CK) to 25.4%
404	(CKS), but <i>Basidiomycota</i> was significantly increased from 3.4% (CK) to 9.4% (CKS)
405	on the 60 th day of the incubation. The relative abundance of <i>Basidiomycota</i> was
406	significantly increased from 1.9% (IF) to 5.7% (IFS) on the 1 st day, Zygomycota was
407	significantly decreased from 7.0% (IF) to 3.7% (IFS) on the 150^{th} day of the incubation.
408	The relative abundance of <i>Basidiomycota</i> was significantly increased from 0.7% (IFM)
409	to 8.3% (IFMS) on the 150 th day (Fig. 2a) of the incubation.
410	The top bacterial and fungal classes were detailed in Fig. S3b and Fig. 2b. The
411	relative abundances of Acidobacteria Gp6 and Acidobacteria Gp4 were significantly

412 increased from 6.2% (IF) and 4.1% (IF) to 10.5% (IFS) and 6.4% (IFS) on the 60^{th} day

413 of the incubation in the bacterial community, respectively (Fig. S3b). Regarding the

414	fungal classes, compared to CK treatment, the relative abundance of Sordariomycetes
415	was significantly decreased on the 1st day in CKS treatment. The relative abundance of
416	Agaricomycetes was significantly increased during the incubation time, and
417	Pezizomycetes was significantly increased on the 1st day in IFS than in IF treatment.
418	The relative abundance of Agaricomycetes was increased more than double in IFMS
419	than in IFM treatment on the 1 st day (Fig. 2b). Furthermore, the relative abundance of
420	Dothideomycetes in IF, IFS, IFM, IFM treatments was higher than in CK and CKS
421	treatments, the relative abundance of Sordariomycetes had the opposite trend.
422	3.3 Taxonomic biomarkers of soil microbial communities
423	The LEfSe analysis from phylum to genus levels was performed to identify high-
424	dimensional biomarker taxa with different abundances among three fertilization
425	regimes after adding straw residues in each incubation time (Fig. 3, 4, and 5). From an
426	overall perspective, after adding straw residue, more significantly different taxa
427	occurred between CK and CKS treatment, IF and IFS treatment than IFM and IFMS
428	treatment during the incubation time (Fig. 3, 4, and 5).
429	For the bacterial communities, in CKS treatment, the Nitrospirae was the most
430	abundant biomarkers on the 1 st day (Fig. 3a); the Actinomycetales and Cytophagaceae

432	enriched on the 150 th day (Fig. 3c). In IFS treatment, the Actinobacteria was enriched
433	on the 1 st day (Fig. 4a); the Nitrospirae and Deltaproteobacteria were significantly
434	enriched on the 60 th day (Fig. 4b); the Actinobacteria and Burkholderiaceae were the
435	significantly different abundance on the 150 th day (Fig. 4c). In IFMS treatment, the
436	Chloroflexi was the most differentially abundant phyla on the 60 th day (Fig. 5b).
437	Regarding the fungal communities, in CKS treatment, the Eurotiomycetes,
438	Dothideomycetes, Chytridiomycota were significantly changed on the 1 st day (Fig. 3d);
439	the Eurotiales was significantly different on the 60 th day (Fig. 3e); the Eurotiales,
440	Wallemiomycetes, and Lasiosphaeriaceae were enriched on the 150 th day (Fig. 3f). In
441	IFS treatment, the <i>Hypocreales_fam_Incertae_sedis</i> showed the highest abundance on
442	the 1st day (Fig. 4d); the Sordariomycetes and Basidiomycota were significantly
443	enriched on the 60 th day (Fig. 4e); the Sordarlales was the most abundant biomarker on
444	the 150 th day (Fig. 4f). In IFMS treatment, the Basidiomycota was the most
445	differentially abundant phylum on the 1 st day (Fig. 5d).
446	3.4 Co-occurrence network in soil bacterial and fungal community
447	To determine the effects of fertilizer treatments on the soil microbial community
448	after adding straw residue, networks were constructed for three fertilization treatments

449 with and without added straw residue (Fig. 6 and 7). Long-term different fertilizer

450	management changed bacterial and fungal co-occurrence patterns, the application of IF
451	and IFM fertilizer increased the complexity of the bacterial community but decreased
452	the complexity of the fungal community (Fig. 6a, b, and c; Fig. 7a, b, and c). After
453	adding straw residue, the num edges and average degree of bacterial communities were
454	decreased in all treatments, the most decreased in IFMS treatment particularly. In fungal
455	communities, the num edges and average degree in CKS treatment were also decreased
456	than in CK treatment but in IFS and IFMS treatments were increased than in IF and
457	IFM treatments (Table 1 and 2). These results indicated that during the process of straw
458	residue decomposition the complexity of microbial network in bacterial communities
459	was decreased, the IFMS treatment was more affected than CKS and IFS treatment. In
460	fungal communities, the complexity of microbial networks in IFS and IFMS treatment
461	were increased than in IF and IFM treatment, but in CKS treatment was decreased than
462	in CK treatment (Fig. 6 and 7).

463 **3.5 Keystone taxa of straw microbiota during straw residue decomposition**

To examine the bacterial and fungal keystone taxa of the whole process of straw residue decomposition in various fertilizer regimes, we regressed the relative abundances of bacteria and fungi at the order level using a random forest regression. We performed 10-fold cross-validation with five repeats to determine the importance

468	of bacterial and fungal classes. We showed the 11 most important bacterial and fungal
469	orders as keystone taxa. We defined these taxa as keystone taxa in the model in order
470	of discriminatory importance as shown in Fig. 8. We found bacterial and fungal
471	keystone taxa during the straw residue decomposition in different fertilizer regimes
472	were various. During the whole incubation time, in the bacterial community, the
473	Anaerolineales order of the Chloroflexi phylum was the most important taxa in CKS
474	treatment; Actinomycetales and Solirubrobacterales orders of Actinobacteria phylum
475	were the taxa in IFS treatment; Myxococcales, Rhodocycales, and Rhodobacterales
476	orders of Proteobacteria phylum were the taxa in IFMS treatment. In the fungal
477	community, the Saccharomycetales order of Ascomycota phylum was the vital taxon in
478	CKS and IFS treatment; the Hypocreales order was the essential taxon in IFMS
479	treatment during the whole straw decomposition process.
480	3.6 SEM analysis on the significant effects of residue-derived labile SOC fractions

481 on soil microbes

482 To assess the significant effects of residue-derived labile SOC fractions on the changes 483 of soil bacterial and fungal community structure in response to different fertilizer 484 management strategies and incubation time, a SEM model was conducted (Fig. 9). 485 Fertilization had positive direct impacts on both the bacterial (+0.49) and the fungal

486	community (+ 0.46), but incubation time only had a positive direct impact on the
487	bacterial community (+ 0.98). All the content of residue-derived labile SOC fractions
488	had positive impacts on the fungal community, the content of residue-derived POC
489	contributed the greatest impact (+ 0.76). The bacterial community was only affected by
490	the content of residue-derived MBC ($+ 0.54$).

491 **4 Discussion**

492 4.1 Straw residue-derived labile SOC fractions as affected by fertilizer
493 management strategies and incubation time

DOC was produced from the decomposition of SOM which is primarily driven by 494 soil microorganisms (Marschner and Bredow, 2002). The content of residue-derived 495 496 DOC decreased quickly, this result is consistent with our previous study in Alfisol (Jin et al., 2020), which indicated residue-derived DOC would be primarily utilized by 497 microorganisms after added straw residue (Kalbitz et al., 2000). Application of organic 498 499 fertilizers would maintain a steady flow of nutrients into the soil, but they released 500 nutrients more slowly (Baghdadi et al., 2018), which may be the reason that the content of residue-derived DOC in IFMS treatment is lower than CKS and IFS treatment. 501

502 MBC is the living microbial component of SOC and is considered a sensitive

503	indicator of microbial activity (Broos et al., 2007; Paul, 1984). In this study, the content
504	of residue-derived MBC in IFS and CKS treatment was higher than the IFMS treatment
505	at the beginning of the incubation time. That because microbial biomass from CKS and
506	IFS treatment responded rapidly (i.e., within 1 day) to the residue returning by
507	accumulating residue-derived C, this result indicated compared with IFM treatment,
508	CK and IF treatment lacked sufficient nutrients. After adding straw, microbes would
509	quickly participate in the decomposition of residue-derived C in a starved state (Bastida
510	et al., 2013). At the end of the incubation time, the content of residue-derived MBC in
511	IFS and IFMS treatment were higher than CKS treatment, which indicated with the
512	organic fertilizer released nutrient, the treatment included fertilizer would promote the
513	fixation of residue-derived MBC by soil microorganisms (Luan et al., 2020).
514	In this study, there was an inverse trend between the content of residue-derived
515	DOC and residue-derived POC in each treatment during the incubation time. That
516	because POC is composed of decomposing plant and microbial residues, and it is served
517	as essential sources of plant nutrients and decomposed by microbes (Feller and Beare,
518	1997; Plaza et al., 2018; Xiao et al., 2017). Thus, with the process of straw residue
519	decomposition, the content of residue-derived POC increased. At the end of incubation
520	time, IFMS treatment retained the largest content of residue-derived POC. The reason 29

is that application of manure improves the soil structure, promotes the formation of soil
aggregates, and allows part of the free SOC to be protected by the soil aggregates,
thereby promoting the increase in the content of straw residue-derived POC (Mi et al.,
2016; Verma and Sharma, 2007).

- 525 4.2 Microbial community structure responses to straw residue decomposition Straw residue returned to the soil affected the soil fungal communities more 526 strongly than the soil bacterial communities on phylum and class levels (Fig. 2), which 527 528 was consistent with the previous study (Maarastawi et al., 2018). Straw is beneficial to 529 enhance microbial processing in CK and IF treatment and has no negative effect on 530 IFM treatment (Fig. 3, 4, and 5). It is known that saprotrophic fungi contribute to the decomposition of residue-derived C and thus boost C mineralization in soil (Dini-531 532 Andreote et al., 2016). Dothideomycetes have been implicated in assimilating C derived from plants (Freedman et al., 2015). In CKS treatment, the Dothideomycetes quickly 533 responded on the 1st day. After adding straw residue, the relative abundance of 534 Basidiomycota significantly changed but these varieties initiated at different incubation 535 stages under various fertilizer regimes (Fig. 2a). Basidiomycota played particularly 536 537 important roles in degrading plant litter with high lignin contents in soils (Entwistle et
- 538 al., 2018). The response of *Basidiomycota* to add straw residue in IFS and IFMS
| 539 | treatment was much quicker than in CKS treatment. The IF and IFM treatment had |
|-----|--|
| 540 | more nutrients than in CK treatment, which could be the driver for this, and is in line |
| 541 | with previous studies that Basidiomycota is better adapted to high nutrient levels |
| 542 | (Hannula et al., 2012). From the 60 th day to the 150 th day during the incubation time, |
| 543 | the relative abundance of Sordariomycetes significantly increased in IFS and IFMS |
| 544 | treatments (Fig. 4e, 4f, 5e, and 5f). A similar result was found in another study (Ma et |
| 545 | al., 2018), probably due to IF and IFM treatment had abundant nutrients and can |
| 546 | promote the growth of microbial groups due to their diverse metabolic capacity so that |
| 547 | Sordariomycetes can play the role of straw degradation effectively (Ding et al., 2017; |
| 548 | Koranda et al., 2014). Sordariomycetes class is one of the largest classes in the phylum |
| 549 | Ascomycota and is also known to play a role in the degradation of crop residue (Qin et |
| 550 | al., 2014; Tardy et al., 2015). Fungi link the allocation of C and sequestration of |
| 551 | nutrients from organic substrates, and it is important in decomposing plant-derived |
| 552 | substrates (Hobara et al., 2014; Quirk et al., 2012). Different fertilization regimes affect |
| 553 | soil fungal community composition in various ways. The organic application directly |
| 554 | increased the amount of SOC and intensively affected soil fungal community |
| 555 | composition (Sun et al., 2016). Inorganic fertilization regulated the quality and quantity |
| 556 | of plant-derived C inputs to indirectly affect the soil fungal community (Weber et al., |

557 2013).

558 4.3 Straw residue decomposition and fertilization changed the co-occurrence 559 patterns of microbial community

In the network for the whole process of straw residue decomposition, we found the 560 bacterial and fungal communities exhibit different co-occurrence patterns (Fig. 6 and 561 7). The further addition of straw residues results in a beneficial increase of the overall 562 microbial abundance together with an enhancement of the complexity of the fungi 563 564 community in applying fertilizer field, but as observed in our study for all fertilizer strategies lead to a reduction in the bacterial complexity. These results indicated that 565 566 straw addition is always detrimental to bacterial complexity independent of fertilizer management. It showed strong competitive interactions between bacterial species for 567 composing straw residue impeded species coexistence and decreased bacterial 568 communities' stability (Ratzke et al., 2020). Although bacterial complexity was 569 570 decreased, that did not affect the role of soil microbes on straw residue decomposition, because of redundancy of functions within the microbial community, i.e., the other 571 microbes would also do the function of residue decomposition (Wagg et al., 2019). 572 573 Regarding fungal communities, the changes in the complexity of the network in IFS and IFMS treatment were different to those found within CKS treatment during the 574 32

575	whole process of residue decomposition (Fig. 7). That results suggested added organic
576	materials to the soil of a high fertility level would provide a good habitat for the growth
577	of the fungi, lead to better stability and provide stronger resistance to the disturbance
578	of the fungal community (Scheffer et al., 2012). Therefore, straw addition is beneficial
579	to increase fungal complexity in the applied fertilizer field, independent if inorganic
580	fertilizer or mixed inorganic fertilizer and manure. Fungal species however would have
581	competition for experienced the resource due to lack of nutrients in CKS treatment so
582	that here we observed a decreased stability of the network (Fuhrman, 2009). Moreover,
583	added straw residue decreased the percentage of positive links in the bacterial
584	community among three treatments, which suggested that straw residue positively
585	affected bacterial species competition and niche separation (Deng et al., 2016; Yu et al.,
586	2018). In the fungal community, the percentage of positive links also decreased in CKS
587	and IF treatment (Fig. 7). This result indicated IFMS treatment had more diverse
588	organic compounds and increased fungal community cooperation and niche overlap
589	during straw residue decomposition (Kong et al., 2020).
590 591	4.4 Temporal responses of keystone taxa dynamics during straw residue decomposition
592	A dominant species often affects ecosystem functioning or a specific process
593	exclusively by its sheer abundance, but keystone taxa can also, more subtly, utilize their

594	influence on microbial system function irrespective of abundance (Banerjee et al., 2018;
595	Fierer, 2017). We further identified 11 bacterial and fungal orders as keystone taxa of
596	microbial community dynamics during the process of straw residue decomposition by
597	random forest model (Fig. 8). These results may be useful for identifying the most
598	closely related bacterial and fungal taxa during the process. We found the keystone taxa
599	in the phylum level in the bacterial community were different in various fertilizer
600	regimes, but the keystone taxon in the fungal community was the same, i.e.,
601	Ascomycota. Chloroflexi was considered as the keystone taxon in CKS treatment (Fig.
602	8a). Chloroflexi was the oligotrophic bacteria phylum and it can grow under low
603	substrate concentrations (Pepe-Ranney et al., 2016). In this study, the relative
604	abundance of Chloroflexi was significantly changed at the early stage of straw residue
605	decomposition in CKS treatment (Fig. 3a), which indicated that CKS treatment had a
606	lower nutrient and limited growth of copiotrophic decomposers, when added straw
607	residue to CK treatment, Chloroflexi would actively participate in assimilating straw
608	residue (Tardy et al., 2015). Actinobacteria was considered as keystone taxon in IFS
609	treatment (Fig. 8b). Actinobacteria was the coptotrophic phylum and more sensitive to
610	carbon sources in the soil and maybe abundant after adding labile SOC (Goldfarb et al.,
611	2011). They are vital saprophytes capable of degrading complex plant debris (Barka et

612	al., 2015). Proteobacteria acted as keystone taxon in IFMS treatment (Fig. 8c).
613	Proteobacteria was a copiotrophic phylum and showed saprophytic lifestyles, the
614	nutrient supply in IFMS treatment was sufficient, thus Proteobacteria fit in this
615	situation and grow fast during the process of straw residue decomposition (Zhan et al.,
616	2018). Ascomycota was the keystone taxon in the fungal community among the three
617	fertilizer regimes regardless of whether the nutrients are sufficient (Fig. 8d, e, and, f).
618	That because Ascomycota is known as the largest and most diverse fungal phylum as
619	well as the key decomposers in the decomposition of the organic materials, it harbored
620	a wide scale of substrate utilization and is essential in breaking down the recalcitrant
621	organic compounds (Schoch et al., 2009; Wang et al., 2018; Wang et al., 2021).

622 4.5 Relationship of residue-derived labile SOC fractions and soil microbes in

623 response to fertilizer management and straw residue returned

The SEM model showed different residue-derived labile SOC fractions had various impacts on the bacterial and fungal communities. The residue-derived DOC, POC, and MBC significantly affected the fungal community while the bacterial community was only significantly affected by residue-derived MBC, that indicated the fungal community was more sensitive bacterial community to the residue-derived labile SOC, and thus played key roles in the process of straw residue decomposition, which was

630	consistent with other studies (Kong et al., 2020; Zhong et al., 2020). The residue-
631	derived MBC significantly affected both bacterial and fungal communities, which
632	indicated MBC is an important indicator of soil microbes and would affect microbial
633	community construction. The residue-derived POC contributed the greatest impact on
634	the fungal community, the reason might be more residue-derived C accumulated and
635	formed POC during the whole process of straw residue decomposition and then
636	significantly affect the fungal community (Goldstein et al., 2020). The influence level
637	of the fungal community from residue-derived DOC was similar to POC, the difference
638	between these two fractions is that residue-derived DOC brought impacts on the initial
639	stage, whereas residue-derived DOC worked at the end of the straw residue
640	decomposition process. This result also verified the changes in the content of residue-
641	derived DOC and POC in our present study.

642 **5 Conclusions**

643 Straw residue addition decreased the bacterial microbial network complexity in all 644 treatments, but increased fungal network complexity in IFS and IFMS. *Chloroflexi* 645 (CKS), *Actinobacteria* (IFS), *Proteobacteria* (IFMS) were keystone taxa in the 646 bacterial community, and *Ascomycota* was the keystone taxon in the fungal community. 647 The straw residue was retained as POC in labile SOC fractions and was further

656	Acknowledgments
655	only via residue-derived MBC.
654	POC and DOC promoted fungal C processing, whereas for bacterial this was facilitated
653	Mollisols, independent of agricultural fertilizer management, straw residue-derived
652	observed temporal changes in the microbial community suggested that in these
651	treatment compared to those without or with only inorganic fertilizer addition. The
650	following straw addition were found for the combined inorganic and organic fertilizer
649	higher under fertilized conditions. Less pronounced effects on the microbial community
648	enhanced in fertilizer management with manure addition. Residue-derived MBC was

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The authors declare that they have no conflict of interest.

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1057

1058 Tables

1059 Table 1 Topological properties of the co-occurrence network in the bacterial community under various

	СК	CKS	IF	IFS	IFM	IFMS
Num.edges	55	46	211	99	139	38
pos.edges (percentage)	55 (100)	43 (93)	209 (99)	91 (92)	134 (96)	36 (94)
neg.edges (percentage)	0 (0)	3 (7)	2 (1)	8 (8)	5(4)	2 (6)
num.vertices	67	67	116	87	117	55
connectance	0.02	0.02	0.03	0.03	0.02	0.03
average.degree	1.64	1.37	3.64	2.28	2.38	1.38
average.path.length	2.43	1.51	4.36	3.47	3.24	1.65
diameter	7	4	11	7	8	4
clustering.coefficient	0.53	0.29	0.52	0.41	0.50	0.24
no.clusters	22	24	11	20	24	19
centralization.degree	0.05	0.02	0.06	0.05	0.05	0.03
centralization.betweenness	0.01	0.00	0.11	0.08	0.02	0.01

1060 fertilizer regimes during the process of straw residue decomposition

1061Note: CK, IF, IFM, CKS, IFS, and IFMS denote control, inorganic fertilizer, inorganic fertilizer plus1062manure, control + straw, inorganic fertilizer + straw, and inorganic fertilizer plus manure + straw

1063 treatment, respectively.

1064 Table 2 Topological properties of the co-occurrence network in the fungal community under various

1065 fertilizer regimes during the process of straw residue decomposition

	СК	CKS	IF	IFS	IFM	IFMS
Num.edges	63	27	16	42	8	33
pos.edges (percentage)	62 (98)	19 (70)	15 (93)	38 (90)	6 (75)	30 (90)
neg.edges (percentage)	1 (2)	8 (30)	1 (7)	4 (10)	2 (25)	3 (10)
num.vertices	45	38	22	43	13	39
connectance	0.06	0.04	0.07	0.05	0.10	0.04
average.degree	2.80	1.42	1.45	1.95	1.23	1.69
average.path.length	2.91	1.44	1.35	1.34	1.00	1.96
diameter	6	4	3	3	1	5
clustering.coefficient	0.45	0.60	0.67	0.82	1.00	0.33
no.clusters	8	14	9	14	6	11
centralization.degree	0.10	0.04	0.12	0.07	0.06	0.06
centralization.betweenness	0.13	0.01	0.02	0.01	0.00	0.02

1066 Note: CK, IF, IFM, CKS, IFS, and IFMS denote control, inorganic fertilizer, inorganic fertilizer plus

1067 manure, control + straw, inorganic fertilizer + straw, and inorganic fertilizer plus manure + straw 1068 treatment, respectively.

1069 Figures Captions

1070 Figure 1 Content of residue-derived dissolved organic carbon (DOC) (a), residue-1071 derived microbial biomass carbon (MBC) (b), and residue-derived particle organic 1072 carbon (POC) (c). Different capital letters indicated significant differences (P < 0.05) 1073 among different incubation time under the same fertilization treatment. Different lowercase letters indicated significant differences (P < 0.05). The values presented in 1074 1075 the figures are given as mean \pm standard errors. CKS denotes control + straw, IFS 1076 denotes inorganic fertilizer + straw, and IFMS denotes inorganic fertilizer plus manure 1077 + straw, respectively.

Figure 2 Relative abundance of the taxonomic composition of soil fungal community at the phylum level (a) and class level (b), respectively. Treatments including CK (control), CKS (control + straw), IF (inorganic fertilizer), IFS (inorganic fertilizer + straw), IFM (inorganic fertilizer plus manure), IFMS (inorganic fertilizer plus manure + straw). Soil samplings were conducted in 1 day, 60 days, and 150 days after added straw residue. Phylum and class names were color-coded on the right listed above.

Figure 3 Cladogram plotted from LEfSe analysis revealed the bacterial taxonomic levels with different abundance values in 1 day (a), 60 days (b), and 150 days (c), and fungal taxonomic levels with different abundance values in 1 day (d), 60 days (e), and 1087 150 days (f) between CK (control) and CKS (control + straw) treatment. The circular ring from inside to outside represents phylum, class, order, family, and genus, 1089 respectively.

Figure 4 Cladogram plotted from LEfSe analysis revealed the bacterial taxonomic levels with different abundance values in 1 day (a), 60 days (b), and 150 days (c), and fungal taxonomic levels with different abundance values in 1 day (d), 60 days (e), and 1093 150 days (f) between IF (inorganic fertilizer) and IFS (inorganic fertilizer + straw) treatment. The circular ring from inside to outside represents phylum, class, order, family, and genus, respectively.

Figure 5 Cladogram plotted from LEfSe analysis revealed the bacterial taxonomic levels with different abundance values in 1 day (a), 60 days (b), and 150 days (c), and fungal taxonomic levels with different abundance values in 1 day (d), 60 days (e), and 1099 150 days (f) in IFM (inorganic fertilizer plus manure) and IFMS (inorganic fertilizer plus manure + straw) treatment. The circular ring from inside to outside representsphylum, class, order, family, and genus, respectively.

Figure 6 The network analysis showed the co-occurrence patterns of bacterial taxa at class level in different treatments including CK (control) (a), IF (inorganic fertilizer) (b), IFM (inorganic fertilizer plus manure) (c), CKS (control + straw) (d), IFS (inorganic fertilizer + straw) (e), IFMS (inorganic fertilizer plus manure + straw) (f) during the whole process of straw residue decomposition. Red lines represent significant positive (P<0.05) linear relationships and blue lines represent negative (P<0.05) linear relationships.

Figure 7 The network analysis showed the co-occurrence patterns of fungal taxa at class level in different treatments including CK (control) (a), IF (inorganic fertilizer) (b), IFM (inorganic fertilizer plus manure) (c), CKS (control + straw) (d), IFS (inorganic fertilizer + straw) (e), IFMS (inorganic fertilizer plus manure + straw) (f) during the whole process of straw residue decomposition. Red lines represent significant positive (P<0.05) linear relationships and blue lines represent negative (P<0.05) linear relationships.

1116 Figure 8 Random-forest model detects bacterial and fungal taxa that predict bacterial 1117 keystone taxa in CKS (control + straw) (a), IFS (inorganic fertilizer + straw) (b), IFMS 1118 (inorganic fertilizer plus manure + straw) (c) and fungal keystone taxa in CKS (control 1119 + straw) (d), IFS (inorganic fertilizer + straw) (e) and IFMS (inorganic fertilizer plus 1120 manure + straw) (f) during the whole process of straw residue decomposition. The top 1121 11 bacterial and fungal orders were identified by applying a random forest regression 1122 of their relative abundances in the straw residue decomposition against the incubation 1123 time of different fertilizer managements. Keystone taxa are ranked in descending order 1124 of importance to the accuracy of the model.

1125Figure 9 A structural equation model (SEM) assesses the significant effects of residue-1126derived labile SOC fractions on the changes of soil bacterial and fungal community1127structure in response to different fertilizer management strategies and incubation time.1128Numbers adjacent to arrows represent path coefficients. The width of arrows indicates1129the strength of the standardized path coefficient. The blue lines indicate positive path1130coefficients, red lines indicate negative path coefficients, and grey lines indicate non-1131coefficients, respectively. Significance levels are denoted with *P < 0.05, **P < 0.01,

- 1132 ***P < 0.001. Chi-square ($\chi^2 = 0.320$), probability level (P = 0.811), goodness-of-fit
- 1133 index (GFI = 0.990), and root-mean-square errors of approximation (RMSEA = 0.000)
- 1134 indicate that our data matches the hypothetical model.

Supplementary Materials

Table S1

Soil basic	Soil basic properties of different fertilizer management strategies in 2018.							
Treatment	SOC (g kg ⁻¹)	δ ¹³ C (‰)	TN (g kg ⁻¹)	C/N ratio	AP (mg kg ⁻¹)	AK(mg kg ⁻¹)		
CK	15.0±0.1 b	-18.8±0.0 a	1.4±0.1 c	10.6±0.4 a	15.32±0.75 c	65.04±3.18 c		
IF	15.0±0.1 b	-19.4±0.0 c	1.5±0.0 b	9.7±0.0 b	20.90±0.70 b	81.42±1.40 b		
IFM	25.2±0.2 a	-19.1±0.1 b	2.6±0.0 a	9.7±0.0 b	80.66±2.56 a	137.37±3.87 a		

Soil basic properties of different fertilizer management strategies in 2018.

Note: The CK denotes no fertilization control treatment, IF inorganic fertilizer treatment, IFM inorganic fertilizer plus manure treatment. SOC denotes soil organic carbon, TN denotes total nitrogen, AP denotes available phosphorous, AK denotes available potassium. Different lowercase letters mean significant differences (P<0.05) in various fertilizer management strategies.

Table S2

Treatment	Incubation time (day)	DOC (mg kg ⁻¹)	MBC (mg kg ⁻¹)	$POC (g kg^{-1})$
	1	828 ± 26	843±33	3.79 ± 0.64
CK	60	371 ± 32	106 ± 38	2.94 ± 0.18
	150	329 ± 52	157 ± 1.7	3.39 ± 0.18
	1	930 ± 1.8	904 ± 25	3.61 ± 0.44
IF	60	332 ± 32	89 ± 6.5	2.85 ± 0.02
	150	308 ± 2.3	111 ± 47	3.32 ± 0.05
	1	925 ± 11	904 ± 4.8	9.68 ± 0.35
IFM	60	412 ± 8.3	247 ± 31	8.44 ± 0.36
	150	268 ± 43	420 ± 12	9.07 ± 0.46
	1	946±59	857 ± 16	3.46 ± 0.1
CKS	60	408 ± 28	145 ± 24	3.83 ± 0.26
	150	369 ± 32	251 ± 81	4.04 ± 0.25
	1	930 ± 31	846 ± 64	3.61 ± 0.08
IFS	60	334 ± 8.6	232 ± 28	3.25 ± 0.29
	150	312 ± 8.1	267 ± 5.3	3.37 ± 0.37
	1	933±16	881 ± 5.2	10.37 ± 1.02
IFMS	60	424 ± 15	236 ± 32	9.04 ± 0.98
	150	326 ± 30	435 ± 24	9.33 ± 0.55

Content of soil dissolved organic carbon (DOC), microbial biomass carbon (MBC), and particle organic carbon (POC) with and without straw residue under different fertilizer management strategies during the incubation time.

Note: CK denotes control, CKS denotes control + straw, IF denotes inorganic fertilizer, IFS denotes inorganic fertilizer + straw, IFM denotes inorganic fertilizer plus manure, IFMS denotes inorganic fertilizer plus manure + straw.



(b)



Figure S1 Box plots of Chao 1 (a) and Shannon (b) diversity indices of soil bacterial community in CK (control), IF (inorganic fertilizer), IFM (inorganic fertilizer plus manure), CKS (control + straw), IFS (inorganic fertilizer + straw), and IFMS (inorganic fertilizer plus manure + straw) after adding straw residue in 1 day, 60 days and 150 days.

(a)



(b)



Figure S2 Box plots of Chao 1 (a) and Shannon (b) diversity indices of soil fungal community in CK (control), IF (inorganic fertilizer), IFM (inorganic fertilizer plus manure), CKS (control + straw), IFS (inorganic fertilizer + straw), and IFMS (inorganic fertilizer plus manure + straw) after adding straw residue in 1 day, 60 days and 150 days.

(a)





Figure S3 Relative abundance of taxonomic composition of soil bacterial community at phylum level (a) and class level (b), respectively. Treatments including CK (control), CKS (control + straw), IF (inorganic fertilizer), IFS (inorganic fertilizer + straw), IFM (inorganic fertilizer plus manure), IFMS (inorganic fertilizer plus manure + straw). Soil samplings were conducted in 1 day, 60 days and 150 days after added straw residue. Phylum and class names were color-code on the right listed aboved.





(b)









(e)









	a: Phaeosphaeriaceae
	b: Pleosporales_fam_Incertae_sedis
	c: Trichocomaceae
	d: Eurotiales
	e: Unassigned
	f: Onygenales
	g: Helotiales_fam_Incertae_sedis
	h: Helotiales
	i: Chaetosphaeriaceae
	j: Chaetosphaeriales
	k: Nectriaceae
	I: Lasiosphaeriaceae
	m: Wallemiaceae
	n: Wallemiales
	o: Unassigned
	p: Spizellomycetales

Figure 3



(a)













a: Exophiala
b: Herpotrichiellaceae
c: Chaetothyriales
d: Penicillium
e: Trichocomaceae
f: Eurotiales
g: Chaetosphaeria
h: Chaetosphaeriaceae
i: Chaetosphaeriales
j: Trichoderma
k: Hypocreaceae
I: Nectriaceae
m: Podospora
n: Lasiosphaeriaceae
o: Unassigned
p: Unassigned
q: Unassigned
r: Unassigned
s: Unassigned
t: Unassigned
u: Unassigned
v: Unassigned
w: Unassigned



a: Blastococcus
b: Geodermatophilus
c: Geodermatophilaceae
d: Arthrobacter
e: Micrococcaceae
f: Nocardioides
g: Actinomycetales
h: Chitinophagaceae
i: Sphingobacteriales
j: Burkholderia
k: Burkholderiaceae





















a: Bacillus b: Bacillaceae_1





(e)















- (d) CKS co-occurrence network
- (e) **IFS co-occurrence network**
- (f) **IFMS co-occurrence network**











Figure 8


Declaration of interests

 \boxtimes The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

□The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

None