

Use of untargeted metabolomics to analyse changes in extractable soil organic matter in response to long-term fertilisation

Tang, Sheng; Ma, Qingxu; Zhou, Jingjie; Pan, Wankun; Chadwick, David R.; Gregory, Andrew S.; Wu, Lianghuan; Jones, Davey L.

Biology and Fertility of Soils

DOI: 10.1007/s00374-023-01706-8

Published: 01/04/2023

Peer reviewed version

Cyswllt i'r cyhoeddiad / Link to publication

Dyfyniad o'r fersiwn a gyhoeddwyd / Citation for published version (APA): Tang, S., Ma, Q., Zhou, J., Pan, W., Chadwick, D. R., Gregory, A. S., Wu, L., & Jones, D. L. (2023). Use of untargeted metabolomics to analyse changes in extractable soil organic matter in response to long-term fertilisation. Biology and Fertility of Soils, 59(3), 301-316. https://doi.org/10.1007/s00374-023-01706-8

Hawliau Cyffredinol / General rights Copyright and moral rights for the publications made accessible in the public portal are retained by the authors and/or other copyright owners and it is a condition of accessing publications that users recognise and abide by the legal requirements associated with these rights.

• Users may download and print one copy of any publication from the public portal for the purpose of private study or research.

- You may not further distribute the material or use it for any profit-making activity or commercial gain
 You may freely distribute the URL identifying the publication in the public portal ?

Take down policy If you believe that this document breaches copyright please contact us providing details, and we will remove access to the work immediately and investigate your claim.

1	Use of untargeted metabolomics to analyse changes in extractable soil organic
2	matter in response to long-term fertilisation
3	
4	Sheng Tang ^{a,b} , Qingxu Ma ^{a,b*} , Jingjie Zhou ^a , Wankun Pan ^{a,b} , David R. Chadwick ^b ,
5	Andrew S. Gregory ^c , Lianghuan Wu ^a , Davey L. Jones ^{b,d}
6	
7	^a Ministry of Education Key Lab of Environmental Remediation and Ecosystem Health,
8	College of Environmental and Resource Sciences, Zhejiang University, Hangzhou,
9	310058, China
10	^b School of Natural Sciences, Bangor University, Gwynedd, LL57 2UW, UK
11	°Protecting Crops and the Environment, Rothamsted Research, Harpenden, Herts,
12	AL5 2JQ, UK
13	^d SoilsWest, Centre for Sustainable Farming Systems, Food Futures Institute, Murdoch
14	University, Murdoch, WA 6105, Australia
15	
16	Email addresses:
17	Qingxu Ma: qxma@zju.edu.cn; Sheng Tang: tangsheng@zju.edu.cn; Wankun Pan:
18	panwankun2017@163.com; Jingjie Zhou: 11814012@zju.edu.cn; David R. Chadwick:
19	d.chadwick@bangor.ac.uk; Lianghuan Wu: finm@zju.edu.cn; Andrew S. Gregory:
20	andy.gregory@rothamsted.ac.uk; Davey L. Jones: d.jones@bangor.ac.uk
21	*Corresponding author: Qingxu Ma
22	[contact details] qxma@zju.edu.cn
	1

23 Abstract

This study aimed to explore the soil metabolic response to long-term fertiliser 24 25 application and the effect of this response on the microbial community by taking advantage of the Woburn Organic Manuring Experiment (UK; operational since 1964). 26 Untargeted metabolomes detected by gas chromatography-time of flight mass 27 spectrometer/mass spectrometry (GC-TOFMS/MS) and ultra-high-pressure liquid 28 chromatography-quadrupole time of flight mass-spectrometer/mass spectrometry 29 (UHPLC-QTOFMS/MS) were used to explore which method better reflected soil 30 31 microbe-accessible metabolites. Microbial community abundance was detected by high-throughput sequencing. We found that long-term farmyard manure application 32 enhanced the soil's total and dissolved C and N contents. The metabolite content 33 detected by GC-TOFMS/MS (TOF detector with a cold injection unit) had a negative 34 linear correlation with soil organic matter, extractable organic nitrogen (N), and 35 microbial carbon (C). Conversely, the metabolite content detected by UHPLC-36 37 QTOFMS/MS was positively correlated with soil organic matter, indicating that metabolites detected by UHPLC-QTOFMS/MS were the main components of soluble 38 39 soil organic matter. More positive than negative correlations were observed between metabolites and bacterial (69.5%) and fungal (67.9%) taxa in the co-occurrence 40 network. Among the bacterial taxa in the network, the family Planococcaceae and 41 genus Paenibacillus showed the most correlations with metabolites. The choice of 42 extraction and detection method affects the identity and number of metabolites 43

44	detected. Therefore, careful consideration is needed when selecting which methods to
45	use. We demonstrated a strong correlation between soil metabolites and microbial
46	community abundance. However, a deeper understanding of soil microbial function
47	and metabolite formation, content, and decomposition is still needed.
48	Keywords: soil organic matter, dissolved organic matter, chemical fertiliser, farmyard
49	manure, untargeted metabolomes

52 Introduction

Most of the C in the terrestrial biosphere is retained as soil organic matter (SOM), 53 54 which originates from microbes, plants, and animals (Johnston et al. 2004). Soil microorganisms derive metabolites predominantly from SOM and its biomass 55 turnover (Brown et al. 2021; Liang et al. 2019). Dissolved organic matter (DOM) is 56 the most biologically-accessible component of SOM, playing a crucial role in C, N, 57 and sulphur (S) cycling (Ma et al. 2020b, 2021a; Swenson et al. 2015). It contains a 58 series of organic matter compounds such as carbohydrates, amino acids, hydroxyl 59 60 acids, sugar acids, nucleosides, sterols, aromatics, amines, and miscellaneous compounds (Brown et al. 2021; Ma et al. 2022); and is in a constant state of flux 61 driven by the microbial community and in situ metabolic activities (Mcleod et al. 62 63 2021; Schmidt et al. 2011). Therefore, understanding the composition and turnover of soil microbe-accessible substrates is crucial for exploring the complex dynamics of 64 microbial communities and their nutrient cycling (Ma et al. 2020c; Zhu et al. 2022). 65

Fertiliser is an important field management intervention that strongly affects soil element content, nutrient cycling, and microbial community composition and function. Globally, agricultural production produces approximately seven billion tons of farmyard manure (FYM) each year (Thangarajan et al. 2013). Manure application to arable land can increase soil structural stability and nutrient levels, thereby enhancing soil C sequestration and biological activity (Maillard and Angers 2014). Partly substituting inorganic fertiliser with FYM can sustain agricultural productivity and

reduce environmental pollution (Hoyle and Fang 2018). FYM application strongly 73 stimulates belowground biogeochemical processes: directly by adding large amounts 74 75 of organic C and nutrients and indirectly by modifying biotic activity (Ma et al. 2018; Liu et al. 2020). Subsoil differs from topsoil in nutrient content, microbial biomass, 76 community composition, bioavailability, age, and accessibility of soil C, which affect 77 the rates of SOM decomposition (Cheng et al. 2017). In contrast to chemical 78 fertilisers, which mainly affect only the topsoil, long-term FYM application generally 79 improves the total and DOM content of both the topsoil and subsoil (Ma et al. 2020b; 80 81 Yan et al. 2018). Additionally, it enhances the activities of enzymes such as β glucosidase, protease, urease, and cellulase (Chang et al. 2010; Ma et al. 2020b). 82 83 However, how the combined application of FYM and chemical fertiliser influences 84 the soil metabolite composition is unclear.

A healthy and well-functioning soil system is vital for providing ecosystem 85 services, especially food production in agricultural ecosystems (Liu et al. 2022; Wei et 86 87 al. 2021). Metabolites in DOM are intermediates or products of enzymatic reactions, including organic acids, sugars, amino acids, and fatty acids. These are involved in 88 89 microbial function, growth, and development. In addition to molecular methods of soil biological quality assessment, extracting and quantifying primary metabolites 90 91 offer an alternative approach to better understanding belowground functions. The metabolic approach has been used extensively in plant biology (Hartman et al. 2020), 92 biomedical science (Gupta et al. 2018), and research on the biochemical responses of 93 microbes (Jozefczuk et al. 2014). However, its application in soil is limited, especially 94

under field conditions, and most studies have only focused on specific metabolites
(Ma et al. 2021a; Warren, 2020). Recent studies have shown that the soil metabolome
is sensitive and can reflect the functional responses of soil microbe communities to
changes in their environment, such as fertiliser application, extreme drought, and drywet or freeze-thaw events (Brown et al. 2021; Miura et al. 2020).

100 Traditionally, soil DOM is quantified by extraction from soil samples using specific solutions (water, KCl, K₂SO₄, etc.) and subsequent analysis of its elemental 101 composition using combustion or oxidization. However, the molecular composition 102 103 cannot be detected using this method (Jones and Willet 2006). Untargeted metabolomics is rapidly gaining attention, but its results are highly dependent on the 104 extraction method and detection instrument used. Gas chromatography/mass 105 106 spectrometry (GC/MS) and liquid chromatography/mass spectrometry (LC/MS) are the most widely-used methods due to their broad analytical scope (alcohols, fatty 107 acids, sterols, carbohydrates, amino acids, etc.), and availability of the spectral 108 109 databases of various metabolites (Brailsford et al. 2019; Brown et al. 2021; Liu et al. 2021; Swenson et al. 2015). Other available methods include capillary 110 111 electrophoresis/mass spectrometry (CE/MS) (Warren 2020) and Fourier transform ion cyclotron resonance/mass spectrometry (FTICR/MS) (Hirai et al. 2004), which are not 112 113 extensively used. The compounds detected vary with the detection method used, and the method that most accurately reflects soil microbe-accessible metabolites is still 114 115 unknown.

116

Microorganisms are the most sensitive soil quality indicators and respond

quickly to changes in soil DOM under chemical and organic fertiliser application (Ma 117 et al. 2020b). A shift in microbial community composition indicates a change in the 118 119 metabolism and function of the community in a soil ecosystem (McGuire and Treseder, 2010). Moreover, the microbial community strongly drives organic C and N 120 121 utilisation and mineralisation (Ma et al. 2018). Nutrient (C, N and P) enrichment 122 induces significant changes in the soil metabolite profile, as it changes microbial activity and its metabolic processes. A recent study based on UHPLC-MS/MS found 123 that inorganic nutrient enrichment causes substantial shifts in both primary and 124 125 secondary metabolism and changes in resource flow and soil functioning, and that the microbial community composition showed significant metabolic flexibility (Brown et 126 al. 2022). C and N (together addition) generally increased peptide synthesis in soil, C 127 128 and P addition increased the fatty acids synthesis, while glucose-C addition increased the synthesis of other carbohydrates (Brown et al. 2022). The systematic coupling of 129 the microbial community and soil metabolomics can valuably improve our 130 131 understanding of microbial strategies in response to environmental stress (Swenson et al. 2018). However, this presents a challenge given the large number of metabolites 132 133 and complexity of the microbial community.

Therefore, in this field-based study, we aimed to improve our understanding of soil metabolic processes by exploring the response of soil metabolites to long-term fertiliser application, and to explore whether the metabolites extracted and detected by different methods can reflects soil organic compounds composition. We hypothesised that (1) the total DOM detected by traditional methods, GC-TOFMS/MS, and 139 UHPLC-QTOFMS/MS should be positively related to each other; (2) soil
140 metabolomics and the microbial community are systematically coupled.

141

142 Materials and methods

143 Experimental site and treatments

Soil samples were collected in June 2018 from the long-term Woburn Organic 144 Manuring experiment running since 1964 in Southeastern England 145 (www.era.Rothamsted.ac.uk/WoburnFarm) to test the effects of organic manures and 146 chemical fertilisers on soil fertility and crop production. The soil is derived from 147 Lower Greensand parent material and is classified as a sandy loam-textured brown 148 sand (10% clay, 6% silt, and 80% sand, excluding organic matter content). The soil 149 samples were collected from three typical treatments that reflected current agronomic 150 regimes: FYM applied at 25–50 t ha⁻¹ y⁻¹ for 28 y (high manure application, High-M), 151 FYM applied at 10 t ha⁻¹ y⁻¹ for 16 y supplemented with chemical fertilisers (low 152 153 manure application, Low-M), and chemical fertilisers only (No-M), with P and K inputs equivalent to 25–50 t ha⁻¹ y⁻¹ FYM. Each treatment consisted of four replicates. 154 155 Each plot was 8.83×8.00 m with a 5-year arable rotation (since 2003 this has been spring barley and mustard, winter beans, winter wheat, forage maize, and mustard, 156 and winter rye). 157

The treatment plots received chemical fertilisers or organic manures for three periods between 1964 and 2018. In the High-M treatment, FYM was applied from

160	1966–71, 1981–87, and 2003–18 (28 y in total). FYM was applied at 50 t ha ⁻¹ in the
161	first two build-up periods and 25 t ha ⁻¹ in the final period. In the Low-M treatment,
162	FYM was applied at 10 t ha ⁻¹ from 2003 onward (16 y in total). Before this, it
163	received chemical fertilisers (P & K) equivalent to 7.5 t ha-1 straw input, containing
164	approximately 30.8 kg N ha ⁻¹ . The No-M treatment received chemical fertilisers as N,
165	P, and K at rates equivalent to High-M during the same years. Since 2003, the Low-M
166	and No-M treatments received annual N (nitrochalk), P (triple superphosphate), and K
167	and S (potassium sulphate) fertilisers at 165, 20, 83, and 36 kg ha ⁻¹ , respectively
168	(equivalent annual rate for a 5-year crop rotation). All other aspects of agronomic
169	management, including harvesting, tillage regime, herbicides and fungicides were
170	consistent among the three treatments. Herbicides including spring-applied Atlantis
171	(mesosulfuron-methyl + iodosulfuron-methyl-sodium, 3:0.6% w/w, Bayer
172	CropScience Ltd, Cambridge, UK) at 400 mL ha ⁻¹ , Hiatus (thifensulfuron-methyl +
173	tribenuron-methyl, 40:15% w/w, Rotam Global AgroSciences Ltd, Hong Kong) at 50
174	g ha ⁻¹ , and Sprinter (2,4-D as the dimethylamine and the monomethylamine salts,
175	700g L ⁻¹ , Nufarm Ltd, Otahuhu, Auckland, New Zealand) at 2 L ha ⁻¹ ; fungicides
176	including spring-applied Keystone (isopyrazam + epoxiconazole, 11.6:9.2% w/w,
177	Agrichem, Yatala, Queensland, Australia) at 500 mL ha ⁻¹ , Folicur (tebuconazole,
178	25.9% w/w, Bayer CropScience Ltd, Cambridge, UK) at 800 mL ha ⁻¹ , and Cello
179	(prothioconazole + tebuconazole + spiroxamine, 10.3:10.5:26.3% w/w, Bayer
180	CropScience Ltd, Cambridge, UK) at 630 mL ha ⁻¹ . The total N, P, and S inputs during
181	the build-up phase (1964–2018) under No-M were 2.46, 1.77, and 0.96 t, respectively.
	9

The total C, N, P, and S inputs under High-M were 112.50, 5.80, 1.26, and 1.22 t,
respectively, while that under the Low-M treatment was 14.10, 2.63, 1.69, and 1.00 t.
Further details of the agronomic regime and experiment can be found in Ma et al.
(2020b).

Winter rye (Secale cereale L.) was sown in the plots, and sampling was 186 performed at the grain-filling stage in 2018. From each of four plots per treatment, the 187 188 topsoil (0–23 cm plough layer) and subsoil (23–38 cm) samples were collected using a 2.5 cm diameter corer (18 cores per plot to make up one replicate). The soil was 189 190 thoroughly mixed by hand and passed through a 5 mm sieve to remove roots, stones, 191 and earthworms. The soil samples were then portioned into three parts: the first was stored at -80 °C to analyse soil metabolites and microbial community, the second was 192 193 stored at 4 °C to assess soil microbial biomass, and the third was air-dried to 194 determine basic soil properties.

195 Determination of soil properties

Basic soil properties were determined using traditional methods. Soil pH was determined at a 1:2.5 (v/v) soil: H₂O ratio. Total C and N were measured by dry combustion of finely milled soil using a CHN-2000 Analyser (Leco Co., St. Joseph, MI, USA). To determine the K₂SO₄ extractable C and N (total, organic, NO₃⁻, and NH₄⁺), 5 g of moist soil was extracted with 25 mL of 0.5 M K₂SO₄ for 30 min at 200 rpm, and centrifuged for 10 min at 12 000 × g at 25 °C. The dissolved organic C (DOC) and total dissolved N (TDN) in the extracts were detected using a multi N/C

2100S TOC-TN Analyser (Analytic Jena AG, Jena, Germany). The NO₃⁻ and NH₄⁺ 203 content in the extracts were detected colourimetrically using a microplate 204 205 spectrophotometer (BioTek Instruments Inc., Winooski, VT, USA). Extractable organic N was calculated by subtracting the NO_3^- and NH_4^+ content from TDN. Soil 206 microbial biomass C (MB-C) and N (MB-N) were determined using the CHCl₃ 207 fumigation-extraction method (Vance et al., 1987). Organic C and N were extracted 208 and detected from the fumigated soil in the same manner as from non-fumigated soil. 209 MB-C and MB-N were calculated by a conversion factor of 2.22 for both C and N 210 211 (Vong et al. 2003). The total soluble protein in the extracts was estimated by the acid hydrolysis of proteins in solution, and amino acids were subsequently determined, as 212 described by Roberts and Jones (2008), and have been reported previously (Ma et al. 213 214 2020b). The 0.5 M K₂SO₄ extracts were passed through a 1 000 MW ultrafiltration membrane using an Amicon[®] stirred cell (Merck-Millipore, Billerica, MA, USA). To 215 quantify the fraction of peptides and free amino acids. Amino acids in the flow-216 217 through were detected using the fluorometric OPAME method before and after acid hydrolysis with 6 M HCl (105 °C, 16 h) under N₂ (Jones et al. 2002). 218

219 Untargeted metabolomics detected by GC-TOFMS/MS

The soil samples stored at -80 °C were freeze-dried using an Edwards Super Modulyo freeze-drier (SciQuip Ltd., Shropshire, UK) for 3 d. The dried soil was ground using a ball mill (Retsch MM200, GmbH, Haan, Germany) to promote metabolite recovery from the microbial biomass (Wang et al. 2015). The samples were extracted by 3:3:2

224	(v/v/v) acetonitrile-isopropanol-water (Brailsford et al. 2019; Brown et al. 2021), as
225	this extraction method can extract a broad range of metabolites. The untargeted
226	metabolome was analysed at the UC Davis West Coast Metabolomics Facility using
227	an automated linear exchange-cold injection system (ALEX-CIS) GC time of flight
228	(TOF) MS (Brailsford et al. 2019; Brown et al. 2021). Briefly, 0.5 μ L of the extracted
229	solution was injected into an Rtx-5Sil MS capillary column (0.25 μm 95%
230	dimethylsiloxane/5% diphenylpolysiloxane coating; 30 m length \times 0.25 mm i.d.;
231	Restek Corp., Bellefonte, PA, USA). This chromatography method yields excellent
232	retention and separation of primary metabolite classes (amino acids, hydroxyl acids,
233	carbohydrates, sugar acids, sterols, aromatics, nucleosides, amines, and miscellaneous
234	compounds) with narrow peak widths of 2-3 s and very good within-series retention
235	time reproducibility of better than 0.2 s absolute deviation of retention times. The GC
236	thermal program was run at 50 °C for 1 min, then increased to 330 °C at 20 °C min ⁻¹ ,
237	and finally maintained at 330 °C for 5 min, with a He mobile phase. Upon elution,
238	samples were injected into a Pegasus IV GC-TOF-MS (Leco Corp., St Joseph, MI,
239	USA), using a mass resolution of 17 spectra s ⁻¹ , from 80–500 Da, at –70 eV ionisation
240	energy and 1800 V detector voltage, with a 230 °C transfer line and 250 °C ion source
241	(Withers et al. 2020). A mixture of internal retention index markers was prepared
242	using fatty acidmethyl esters of C8, C9, C10, C12, C14, C16, C18, C20, C22, C24,
243	C26, C28, and C30 linear chain length, dissolved in chloroform at concentrations of
244	0.8 mg mL ⁻¹ (C8–C16) or 0.4 mg mL ⁻¹ (C18–C30) as detailed in Fiehn et al. (2008).
245	The raw data files were pre-processed directly after data acquisition and stored as 12

ChromaTOF-specific *.peg files. ChromaTOF v. 2.32 (Leco Corp.) was used for data 246 pre-processing without smoothing, with a 3 s peak width, baseline subtraction just 247 248 above the noise level, and automatic mass spectral deconvolution and peak detection at signal/noise levels of 5:1 throughout the chromatogram. Apex masses were reported 249 250 for use in the BinBase algorithm. The results were exported to a data server with absolute spectra intensities and further processed by a filtering algorithm implemented 251 in the metabolomics BinBase database, as shown in Withers et al. (2020). Both known 252 and unknown compounds were analysed using MetaboAnalyst v4.0 (Chong et al., 253 254 2018; Xia and Wishart 2016). Prior to analysis, the data were log_{10} transformed and scaled by Pareto scaling (Chong et al. 2018). 255

256 Untargeted metabolomics detected by UHPLC-QTOFMS/MS

Complex lipid extraction was conducted using a modified bi-phasic method (Matyash 257 et al. 2008), which is advantageous as the lipids are retained in the upper extraction 258 phase, and the methyl tertiary-butyl ether (MTBE) solvent has a density lower than 259 that of water. Compared to chloroform (CHCl₃), MTBE can be detected directly 260 without the risk of contamination from the interphase or aqueous phase. Briefly, 225 261 262 µL of MeOH with internal standards was added to a 40 mg freeze-dried and ground soil sample and vortexed for 20 s; 750 µL MTBE was subsequently added and 263 vortexed for 10 min. Samples were placed in a bead grinder for 30 s and then shaken 264 for 6 min at 4 °C; 188 µL of MS-grade water was added, and the sample was 265 centrifuged for 2 min at 14 000 \times g at 4 °C. The upper phase was transferred to two 266

tubes (350 µL/tube), and one tube was evaporated to dryness using a SpeedVac. Dried 267 extracts were re-suspended with a mixture of 1:9 toluene: MeOH (v/v) and an internal 268 269 standard. The samples were analysed using an Agilent 1290 Infinity liquid chromatography (LC) system (G4220A binary pump, G4226A autosampler, and 270 271 G1316C Column Thermostat) coupled to an Agilent 6530 MS (positive ion mode). Lipids were separated on an Acquity ultra high-pressure chromatography (UHPLC) 272 CSH C18 column (1.7 μ m; 100 \times 2.1 mm) (Brown et al. 2021). The data were 273 processed by the mass spectrometry-data independent analysis (MS-DIAL) software 274 275 (Tsugawa et al. 2015), followed by data clean-up using the mass spectral feature list optimiser (MS-FLO) (Defelice et al. 2017). Peaks were annotated, and the 276 MassHunter Quant software was applied to verify peak candidates (Brown et al. 277 278 2021). Valid and reproducible peaks were analysed using targeted MS/MS to increase overall peak annotations. In addition, nine internal standards were used to convert 279 peak heights into good estimates of absolute (micromolar) concentrations for a range 280 281 of biogenic amines typically detected in biofluids and tissues (shown in supporting materials). Notably, internal standards were included, but only for peak correction and 282 quality control. Therefore, the data presented are qualitative, and the compounds were 283 tentatively identified in line with typical untargeted analyses (Brown et al. 2021). This 284 285 UHPLC-TOFMS/MS method reportedly yields an excellent retention and separation of acylcarnitines, trimethylamine oxide, cholines, betaines, S-adenosine methionine, 286 287 S-adenosine-L-homocysteine, nucleotides and nucleosides, methylated and acetylated amines, di- and oligopeptides, while also yielding excellent retention and separation 288 14

289	of metabolite classes with narrow peak widths of 5-20s (biogenic amines, cationic
290	compounds). The internal standards were D3-Creatinine (392 ng mL ⁻¹), D9-Choline
291	(50 ng mL ⁻¹), D9-TMAO (49 ng mL ⁻¹), D3-1-Methylnicotinamide (130 ng mL ⁻¹),
292	Valine-Tyrosine-Valine (146 ng mL ⁻¹), D9-Betaine (151 ng mL ⁻¹), D3-AC(2:0) (33 ng
293	mL ⁻¹), D3-Histamine, N-methylproline (31 ng mL ⁻¹), D3-L-Carnitine (158 ng mL ⁻¹),
294	D3-Creatine (171 ng mL ⁻¹), D5-L-Glutamine (1941 ng mL ⁻¹), D3-DL-Glutamic acid
295	(2426 ng mL ⁻¹), D3-DL-Aspartic acid (9901 ng mL ⁻¹), D4-Cystine (721 ng mL ⁻¹), D4-
296	Alanine (2847 ng mL ⁻¹), D7-Arginine (743 ng mL ⁻¹), D3-Asparagine (720 ng mL ⁻¹),
297	D5-Histidine (990 ng mL ⁻¹), D10-Isoleucine (885 ng mL ⁻¹), D10-Leucine (1856 ng
298	mL ⁻¹), D8-Lysine (681 ng mL ⁻¹), D8-Methionine (495 ng mL ⁻¹), D2-Ornithine (632 ng
299	mL ⁻¹), D8-Phenylalanine (743 ng mL ⁻¹), D7-Proline (1274 ng mL ⁻¹), D3-Serine (2475
300	ng mL ⁻¹), D5-Threonine (1406 ng mL ⁻¹), D8-Tryptophan (619 ng mL ⁻¹), D8-Valine
301	$(5569 \text{ ng mL}^{-1}).$

302 Soil DNA extraction and sequencing of bacteria and fungi

303 Following the manufacturer's protocols, DNA from soil subsamples (0.5 g) was extracted using a FastDNA SPIN kit (MP Biomedicals, Irvine, CA, USA). A 304 NanoDrop ND-1000 UV-Vis spectrophotometer (NanoDrop 305 Technologies, Wilmington, DE, USA) was then used to identify the concentrations and quality of the 306 extracted DNA. Primers 515F-806R (Brown et al. 2021) for bacteria and ITS1F-ITS2 307 (Gardes and Bruns 2010) for fungi were used for amplification. The polymerase chain 308 309 reaction products were sequenced using the Illumina Novaseq platform. Bacterial and 310 fungal sequence data were processed using an in-house pipeline (Kai et al. 2017). Sequences with a length exceeding 200 bp were retained for downstream analyses. 311 312 Operational taxonomic units (OTUs) were clustered at a 97% similarity. We annotated the taxonomic data for representative sequences of bacteria and fungi using the SILVA 313 314 (Quast et al. 2012) and UNITE (Nilsson et al. 2019) databases, respectively. A total of 1 790 490 and 1 616 428 high-quality bacterial and fungal sequences were generated 315 with an average read count of 74 604 (55 781–85 255) and 67 351 (43 939–81 715) 316 per sample, respectively. 317

318 Data and statistical analysis

All statistical analyses were performed using R (version 3.4.3). The metabolomics 319 320 data were log₁₀ transformed. Agglomerative hierarchical clustering analyses were performed for the metabolite concentration data under fertiliser treatment and soil 321 depth according to Pearson correlation coefficients. The dendrograms were combined 322 with heat maps generated based on the z-scores of metabolite concentrations. 323 Principal component analysis (PCA) was performed to determine the relationship 324 between fertiliser treatment and C, N, and metabolites at two soil depths. One-way 325 326 ANOVA and Tukey *post-hoc* testing were used to assess the differences among the fertiliser treatments, and the Shapiro-Wilk test was used to check for normality; the 327 topsoils and subsoils were analysed separately (p < 0.05). A random forest analysis 328 was performed using the 'randomForest' R package of the Linear discriminant 329 analysis effect size (LEfSe) on the Galaxy platform. The interaction between 330

331 metabolite concentrations and the microbial community was visualised using the
332 'psych' package in R and Gephi (http://gephi.github.io/).

333

334 **Results**

335 Effect of long-term fertiliser on soil properties

In the collected sandy soil samples, manure application increased the total and dissolved contents of C (Total C, DOC) and N, which increased with the FYM application rate (Fig. S1). Generally, the total and dissolved C and N contents were greater in the topsoil than in the subsoil. The peptide and amino acid contents were clustered with DOC. In addition, MB-C and MB-N were clustered with total C and N, SOM, and protein content.

342 Effects of long-term fertiliser on primary metabolites detected by GC-TOFMS/MS

343 The untargeted primary metabolomics analysis using GC-TOFMS/MS tentatively identified 186 compounds, of which 71 were previously identified. Among the known 344 compounds, the concentrations of 33 compounds differed significantly between 345 treatments (p < 0.05) (supporting materials). In contrast, the dissolved SOM content 346 extracted by 3:3:2 (v/v/v) acetonitrile-isopropanol-water was generally lower in the 347 subsoil than in the topsoil. There were two distinct responses: the concentrations in 348 349 the first group decreased with long-term Low-M and High-M treatments and showed higher concentrations in the topsoil compared to those in the subsoil (n = 12); the 350

second group had higher concentrations in the subsoil than those in the topsoil (n =
59). The 50 most significant known metabolites revealed by ANOVA are presented in
Fig. 1.

354 Effects of long-term fertiliser on primary metabolites detected by UHPLC-355 QTOFMS/MS

The curated complex lipid analysis identified 2 944 individual compounds, of which 144 were known (supporting materials). Among these previously identified compounds, the 90 that appeared in the highest concentrations were clustered into three groups:

(1) Compounds that appeared at higher concentrations in the topsoil than the subsoil
and at higher concentrations under the No-M than the Low-M and High-M
treatments (n = 35).

363 (2) Compounds that appeared at higher concentrations in the topsoil than in the 364 subsoil and at the highest concentrations under the highest manure application (n 365 = 24).

366 (3) Compounds with higher concentrations in the subsoil than in the topsoil (n = 31).
367 The 50 most significant known metabolites revealed by ANOVA are presented in Fig.
368 2.

369 *PCA analysis of soil properties and soil metabolomics*

370 We observed a significant difference between the properties of the topsoil and subsoil

of the Low-M treatment and a large difference between the No-M and High-M
treatments. The PCA indicated that the No-M and High-M treatments significantly
influenced the soil metabolomes detected by GC-TOFMS/MS and UHPLCQTOFMS/MS (Fig. 3).

A linear relationship between dissolved organic matter and metabolites detected by
 GC-TOFMS/MS and UHPLC-QTOFMS/MS

The metabolite profiles detected by GC-TOFMS/MS and UHPLC-QTOFMS/MS 377 were inversely correlated (Fig. S2). Therefore, while the metabolites detected by 378 UHPLC-QTOFMS/MS were positively correlated to SOM, EON (extractable organic 379 nitrogen), and MB-C, those detected by GC-TOFMS/MS were inversely correlated 380 (Fig. 4). In addition, several compounds such as tyrosine, glucose-1-phosphate, 381 leucine, glutamine, and isoleucine were detected by both GC-TOFMS/MS and 382 UHPLC-QTOFMS/MS, but only isoleucine detected by GC-TOFMS/MS was 383 positively linked with that detected by UHPLC-QTOFMS/MS. 384

Response of bacterial and fungal communities to long-term organic and inorganic
 fertiliser application

The LEfSe analysis identified the microbial taxa that differed significantly between fertiliser regimes (Fig. 5). The High-M treatment had the most enrichment indicators (that were significant), whereas the Low-M treatment had the least. Among the bacteria, indicators belonged mainly to Proteobacteria, Actinobacteria, Firmicutes, and Acidobacteria, the predominant bacterial phyla (Fig. 5A). Particularly in the LowM treatment, Nitrospirae, which are involved in soil nitrification, were enriched. In the High-M treatment, the identified indicators included *Bacillus* and Proteobacteria, Actinobacteria, and Firmicutes. Among the fungi, the most prominent indicators were Ascomycota, Mucoromycota, and Aphelidiomycota, the predominant fungal phyla (Fig. 5B). Long-term high-rate manure application (High-M) significantly increased the abundance of Ascomycota, whereas long-term chemical fertiliser application (No-M) significantly enriched Mucoromycota.

399 Metabolites drive microbial community succession

The random forest analysis revealed the relative importance of metabolites in 400 determining microbial community succession. The 15 most important metabolites are 401 402 presented in Fig. 6. The most important driver of both bacterial and fungal community succession was 5'-methylthioadenosine (MTA). After that, N-epsilon-acetyllysine, 403 gamma-glutamylleucine, histidine, and 3-indolepropionic acid correlated the most 404 with the bacterial community (Fig. 6A). Fungal community succession correlated 405 most strongly with 2'-O-methyladenosine, 1,4-cyclohexanedione, isobutyryl-L-406 carnitine, and corticosterone after MTA (Fig. 6B). 407

We constructed a co-occurrence network based on the LEfSe and random forest analysis results to further clarify the correlation between the microbial taxa and specific metabolites (Fig. 7). The 15 most important metabolites for the two communities and the identified indicators were selected to construct the co-occurrence network. There were more positive than negative correlations between bacterial taxa

and metabolites (69.5%) and fungal taxa and metabolites (67.9%) in the network. 413 Among the bacterial taxa in the network, the family Planococcaceae and genus 414 415 Paenibacillus showed the most correlations (8) with metabolites (Fig. 7A and Table S1). In the case of metabolites, gamma-glutamylleucine had the most correlations (20) 416 417 with bacterial taxa. The fungal network was simpler, with fewer nodes and total degrees (Fig. 7B and Table S2) than the bacterial network. Aspergillus caesiellus and 418 Thermomyces lanuginosus had the most correlations (8) with metabolites among the 419 420 fungal taxa in the network, and MTA had the most links with fungal taxa.

421

422 **Discussion**

423 Effect of long-term organic and inorganic fertiliser application on soil organic matter

As expected, long-term FYM increased the stock of soil total and DOM directly by 424 adding large amounts of organic C and nutrients and indirectly by increasing the 425 microbial biomass (Liu et al. 2020; Ma et al. 2018). Microorganisms can rapidly 426 utilise organic C, and the microbial necromass contributes greatly to SOC (soil 427 organic C) sequestration, especially in soils supplemented with manure and that have 428 429 an enhanced microbial biomass (Cui et al. 2020; Ma et al. 2020a; Wang et al. 2021). Based on the evaluation of glucosamine and muramic acid from bacterial and fungal 430 431 necromasses, Wang et al. (2021) found that microbial necromass contributed to approximately half of the soil organic C in grassland and cropland soils. Therefore, 432 the increased microbial biomass after FYM application could stimulate the formation 433

434 of SOM.

Long-term high FYM application increased the EON content in the subsoil but 435 not in the topsoil, which was in direct contrast to the effect of the chemical fertilisers. 436 We ascribe this to the blockage of sorption sites by organic acids and humic 437 substances released from the manure (Haynes and Mokolobate 2001), which increases 438 soluble organic N leaching to the subsoil (similar to that of soil soluble organic P) 439 (Ma et al. 2020a). The sandy soil we studied has a lower adsorption ability compared 440 to soils with high clay content; therefore, leaching has a greater effect on dissolved 441 442 SOM content.

443 Effect of long-term organic and inorganic fertiliser application on soil metabolites 444 detected by GC-TOFMS/MS

445 Besides the basic chemical and physical soil characteristics, metabolic profiles especially sugars, amino acids, and organic acids are an important indicator of soil 446 447 quality and ecosystem function (Withers et al. 2020). Metabolites can be sensitive to changes in the soil environment directly related to the physicochemical properties and 448 microbial community. The metabolomics data detected by GC-TOFMS/MS was 449 negatively linked to dissolved organic C and N contents. Also, the total metabolite 450 content detected by GC-TOFMS/MS and UHPLC-QTOFMS/MS were negatively 451 correlated. While GC-TOFMS/MS can detect numerous primary metabolites, it is 452 453 generally limited by its poor resolving power for highly labile metabolites and several N-containing metabolites, such as coelute and other sugar compounds with the same 454

m/z (Brown et al. 2021). Additionally, the samples were only detected by MS in 455 positive ion mode; therefore, compounds only detectable in the negative mode were 456 missed. Furthermore, some compounds, such as glycine betaine, are not amenable to 457 derivatisation and hence are undetectable (Brown et al. 2021). Therefore, in this study, 458 the compounds detected by GC-TOFMS/MS were not exhaustive, and the 459 metabolomics data detected by GC-TOFMS/MS was negatively correlated to EON. 460 The extraction solution might also greatly affect the metabolites detected. Extractions 461 using 3:3:2 (v/v/v) acetonitrile-isopropanol-water reportedly cover a broad range of 462 463 metabolites, which is still lower than that when using water or other solutions (Lee et al. 2012; Swenson et al. 2015). Likewise, when focusing on sterols and fatty acids, 464 higher concentrations of organic solvent are needed, and aqueous solutions are better 465 466 at extracting polar and small compounds due to the polar nature of the compounds (Swenson et al. 2015). Our results suggest that the metabolome detected by GC-467 TOFMS/MS might not accurately reflect the state of the soil and that UHPLC-468 469 QTOFMS/MS may yield more informative results in these sandy soils. However, this 470 result is based on one study site, and the results may be different if focusing on 471 different soils.

472 Effect of long-term organic and inorganic fertiliser application on soil metabolites
473 detected by UHPLC-QTOFMS/MS

The selected compounds detected by UHPLC-QTOFMS/MS were clustered into three groups. The first group included compounds that were more concentrated in the

topsoil than the subsoil and more concentrated under chemical fertiliser application 476 (No-M) than under low and high manure application (n = 35). Their lower 477 478 concentration in the subsoil could be due to the higher absorption by soil particles, as limited compounds in the topsoil leached to the subsoil. The group comprised 479 480 predominantly large molecular compounds, such as corticosterone, phenylacetamide, coniferylaldehyde, quinolone, nicotine, and hexadecylamine, which might have been 481 derived as secondary metabolites from soil microorganisms after they utilised the 482 nutrients from chemical fertilisers. The long-term use of chemical fertiliser might 483 484 stimulate microorganisms to synthesise those compounds and assimilate the inorganic nutrients to adapt to the environmental changes caused by chemical fertiliser 485 application. The second group of compounds had the highest concentration in the 486 487 topsoil under high manure application. This group might have been derived from farmyard manure or microbial cycling. The last group had the highest concentration in 488 the subsoil, either because they leached into the subsoil because of a lower adsorption 489 490 ability, or because they were derived from microorganisms adapted to oxygen-491 deficient conditions in the subsoil (Ma et al. 2020a).

The metabolome detected by UHPLC-QTOFMS/MS was strongly correlated to total and dissolved SOM, indicating that UHPLC-QTOFMS/MS better reflected SOM content and composition, at least in this sandy bulk soil. In addition, the compounds were not strongly correlated to the dissolved organic C but were strongly correlated to extractable organic N. This might have been caused by the decoupling of C and N in

498 Correlations between soil metabolism and the bacterial community

Dissolved organic C, especially low molecular-weight compounds, including root 499 exudates, could be utilised directly as C sources by soil microbes (Swenson et al. 500 2015). Therefore, soil metabolomics could improve our understanding of the coupling 501 between organic/inorganic compounds and microbial communities in the soil (Johns 502 et al. 2017). In this study, the most correlated factor for both bacterial and fungal 503 community succession was MTA, followed by N-epsilon-acetyllysine, gamma-504 glutamylleucine, histidine, and 3-indolepropionic acid for the bacterial community 505 (Fig. 6A) and 2'-O-methyladenosine, 1,4-cyclohexanedione, isobutyryl-L-carnitine, 506 and corticosterone for the fungal community (Fig. 6B). MTA is a naturally occurring 507 sulphur-containing nucleoside, indicating that S metabolism is important for the 508 formation of microbial communities. Recently, S might have become a limiting 509 element for microbial growth as a result of considerably decreased sulphur dioxide 510 511 emissions following strict air-quality regulations, application of fertilisers with a limited S content, and a reduced S return via farmyard manure (Piotrowska-Długosz 512 513 et al. 2017).

N-epsilon-acetyllysine is a derivative of the amino acid lysine, and A glutamyl-L-amino acid is obtained through formal condensation of the gamma-carboxy group of glutamic acid with the amino group of leucine. Indole-3-propionic acid is a bacterial metabolite that exerts antioxidant and neuroprotective activities. Most of these metabolites are amino acid derivatives, which can be utilised by soil microorganisms, hence regulating microbial activity and/or changing microbial diversity (Ma et al. 2021b). Maltose and sucrose are low molecular compounds directly utilised as energy sources by microbes in the soil (Vives-Peris et al. 2020). In particular, organic acids and sugars are the main drivers of shifts in soil microbial communities in the rhizosphere and are positively or negatively correlated with the relative abundances of bacteria (Song et al. 2020; Swenson et al. 2015).

Our results showed that metabolite profiling and high-throughput sequencing 525 could be successfully integrated. We found more positive correlations between 526 bacterial taxa and metabolites (69.5%) and fungal taxa and metabolites (67.9%) than 527 negative correlations in the co-occurrence network. The family Planococcaceae and 528 genus Paenibacillus showed the most correlations with metabolites among the 529 bacterial taxa in the network (Fig. 7A and Table S1). Paenibacillus is an important 530 bacterium in bulk soil that plays an important role in N fixation, hormone production, 531 siderophore secretion, and mineral nutrient activation (Li et al. 2021; Timmusk et al. 532 2005). In the rhizosphere, Proteobacteria are reported to be the main utilisers of plant 533 root exudates (Haichar et al. 2008) and respond positively to low molecular-weight 534 substances (Goldfarb et al. 2011). However, Bacteroidetes is not a dominant bacterial 535 phylum in bulk soil but is found in high abundance in the rhizosphere (Aleklett et al. 536 2015). Therefore, it was not the dominant bacterial phylum in the tested bulk soil. In 537 the case of metabolites, gamma-glutamylleucine had the most links (20) with bacterial 538

539 taxa.

Unlike the bacterial network, the fungal network was simpler, with fewer nodes 540 541 and lower total degrees (Fig. 7B and Table S2). Previous studies have demonstrated that fungi tend to decompose recalcitrant SOC, such as lignin and cellulose, and 542 bacteria then utilise the fungal-derived products (de Boer et al. 2005). Among the 543 fungal taxa in the network, Aspergillus caesiellus and Thermomyces lanuginosus had 544 the most correlations (8) with metabolites. In addition, MTA was found with the most 545 degrees with the fungal taxa. Soil microbial community composition can be achieved 546 547 by high-throughput sequencing. However, the actual microbial functions, such as their metabolism, are difficult to obtain with soil metagenome or amplicon sequencing 548 (Jansson and Hofmockel, 2018). 549

The soil metabolome was formed mainly of organic acids, sugars, and sugar 550 derivatives, which were largely negatively correlated with bacterial alpha-diversity. 551 Compared to sugars, organic acids accounted for more bacterial community 552 compositions at high taxonomic ranks, but this was reversed at the species and genus 553 levels. Keystone species in the co-occurrence network, such as Microvirga, 554 555 Bryobacter, and Bradyrhizobium were significantly correlated with organic acids and sugars (Liu et al. 2020). We anticipate that these substrate-genome linkages could be 556 557 further evaluated and refined using other approaches. Stable isotope probing coupled with labelled DNA sequencing (Orsi et al. 2016; Pepe-Ranney et al. 2016) and 558 559 integrated NanoSIMS and FISH imaging (Woebken et al. 2015; Fike et al. 2008) may

be used to examine the spatial localisation of microbes and their activities (Swenson 560 et al. 2018). Complementary analyses of metabolic flux through real-time MS or 561 562 NMR combined with stable isotopes may also offer a deeper understanding of metabolic network dynamics (Ina and David 2016; Jeong et al. 2017). A metabolomic 563 profile alone cannot provide a complete understanding of interacting molecular 564 pathways and their modes of regulation; the variation of metabolite levels cannot 565 definitively infer functional change. Combining genomic and proteomic or 566 transcriptomic results with metabolites may contribute toward a more holistic 567 568 understanding of soil microbial function and regulation (Trauger et al. 2008).

569 Conclusions

We found that long-term farmyard manure application enhanced the total and 570 dissolved soil contents of C and N. The metabolome detected by UHPLC-571 QTOFMS/MS was positively linearly correlated to SOM, EON, and MB-C, indicating 572 that the metabolites detected by UHPLC-QTOFMS/MS reflect the soil organic matter 573 content and composition. There were more positive correlations between bacterial and 574 fungal taxa and metabolites than negative correlations in the network. The family 575 576 Planococcaceae and genus Paenibacillus showed the most correlations with metabolites among the bacterial taxa in the network. Combining genomic and 577 proteomic or transcriptomic results with metabolites may contribute toward a more 578 holistic understanding of soil microbial function and regulation. It is impossible to 579 extract all metabolites from soil, and the detected metabolites depend on the 580

581	extracting solution; therefore, a more detailed exploration of both extraction and
582	detection methods that more accurately reflect the composition of soil compounds and
583	their turnover is needed.

584 Acknowledgements

This work was supported by the National Natural Science Foundation of China 585 (32102488, 32172674); Zhejiang Provincial Natural Science Foundation of China 586 (LZ23C150002); Zhejiang Key Research and Development Program (2022C02018, 587 2023C02016); the UK-China Virtual Joint Centre for Agricultural Nitrogen [grant 588 589 number CINAg, BB/N013468/1], which is jointly supported by the Newton Fund, via 590 UK BBSRC and NERC, and the Chinese Ministry of Science and Technology. The field experiment is maintained as part of the Rothamsted Long-Term Experiments 591 592 National Capability, funded by BBSRC (BBS/E/C/000J0300). We thank the curators of the Electronic Rothamsted Archive (e-RA) for access to data from the Rothamsted 593 Long-Term Experiments. 594

595

596 **Conflict of interest**

597 The authors declare no conflict of interes
--

598

599

601 **References**

Aleklett K, Leff JW, Fierer N, Miranda H (2015) Wild plant species growing closely
 connected in a subalpine meadow host distinct root-associated bacterial communities.
 PeerJ, 3: e804.

Brailsford FL, Glanville HC, Golyshin PN, Marshall MR, Lloyd CE, Johnes PJ, Jones
DL (2019) Nutrient enrichment induces a shift in dissolved organic carbon (DOC)

metabolism in oligotrophic freshwater sediments. Sci Total Environ 690: 1131–1139.

Brown RW, Chadwick DR, Zang H, Jones DL (2021) Use of metabolomics to
quantify changes in soil microbial function in response to fertiliser nitrogen supply
and extreme drought. Soil Biol Biochem 160: 108351.

611 Brown RW, Chadwick DR, Bending GD, Collins CD, Whelton HL, Daulton E, 612 Covington JA, Bull ID, Jones DL (2022) Nutrient (C, N and P) enrichment induces 613 significant changes in the soil metabolite profile and microbial carbon partitioning.

614 Soil Biol Biochem 172, 108779.

Chang EH, Chung RS, Tsai YH (2010) Effect of different application rates of organic
fertilizer on soil enzyme activity and microbial population. Soil Sci Plant Nutr 53:
132–140.

Cheng L, Zhang N, Yuan M, Xiao J, Qin Y, Deng Y, Tu Q, Xue K, Van Nostrand JD,
Wu L (2017) Warming enhances old organic carbon decomposition through altering
functional microbial communities. ISME J: 11, 1825–1835.

Chong J, Soufan O, Li C, Caraus I, Li S, Bourque G, Wishart DS, Xia J (2018)
MetaboAnalyst 4.0: towards more transparent and integrative metabolomics analysis.
Nucleic Acids Res 46: 486–494.

Cui J, Zhu Z, Xu X, Liu S, Jones D, Kuzyakov Y, Shibistova O, Wu J, Ge T (2020)
Carbon and nitrogen recycling from microbial necromass to cope with C:N
stoichiometric imbalance by priming. Soil Biol Biochem 142: 107720.

de Boer W, Folman LB, Summerbell RC, Lynne B (2005) Living in a fungal world:
impact of fungi on soil bacterial niche development. FEMS Microbiol Rev 29: 795811.

630 Defelice BC, Mehta SS, Samra S, Ajka T, Wancewicz B, Fahrmann JF, Fiehn O (2017)

631 Mass Spectral Feature List Optimizer (MS-FLO): a tool to minimize false positive

peak reports in untargeted LC-MS data processing. Anal Chem 89: 3250–3255.

Fiehn O, Wohlgemuth G, Scholz M, Kind T, Lee DY, Lu Y, Moon S, Nikolau B (2008)

634 Quality control for plant metabolomics: reporting MSI-compliant studies. Plant J 53:

635 691**-**704.

636 Fike DA, Gammon CL, Ziebis W, Orphan VJ (2008) Micron-scale mapping of sulfur

- cycling across the oxycline of a cyanobacterial mat: a paired nanoSIMS and CARDFISH approach. ISME J 2: 749–759.
- Gardes M, Bruns TD (2010) ITS primers with enhanced specificity for
 basidiomycetes--application to the identification of mycorrhizae and rusts. Mol Ecol 2:
 113–118.
- 642 Goldfarb KC, Karaoz U, Hanson CA, Santee CA, Bradford MA (2011) Differential
- 643 Growth Responses of Soil Bacterial Taxa to Carbon Substrates of Varying Chemical
- 644 Recalcitrance. Front Microbiol 2: 94.
- Gupta L, Ahmed S, Jain A, Misra R (2018) Emerging role of metabolomics in
 rheumatology. Int J Rheum Dis 21: 1468–1477.
- Haichar FE, Marol C, Berge O, Rangel-Castro JI, Prosser JI, Balesdent J, Heulin T,
 Achouak W (2008) Plant host habitat and root exudates shape soil bacterial
 community structure. ISME J 2(12): 1221–1230.
- Hartman S, Sasidharan R, Voesenek L (2020) The role of ethylene in metabolic
 acclimations to low oxygen. New Phytol 229: 5–7.
- Haynes RJ, Mokolobate MS (2001) Amelioration of Al toxicity and P deficiency in
 acid soils by additions of organic residues: a critical review of the phenomenon and
 the mechanisms involved. Nutr Cycl Agroecosystems 59: 47–63.
- Hirai MY, Yano M, Goodenowe DB, Kanaya S, Kimura T, Awazuhara M, Arita M,
 Fujiwara T, Saito K (2004) Integration of transcriptomics and metabolomics for
 understanding of global responses to nutritional stresses in Arabidopsis thaliana. P
 Natl Acad Sci USA 101: 10205–10210.
- Hoyle FC, Fang Y (2018) Impact of agricultural management practices on the nutrient
 supply potential of soil organic matter under long-term farming systems. Soil Till Res
 175: 71–81.
- Ina A, David M (2016) Advantages and pitfalls of mass spectrometry basedmetabolome profiling in systems biology. Int J Mol Sci 17: 632.
- Jansson JK, Hofmockel KS (2018) The soil microbiome-from metagenomics tometaphenomics. Curr Opinion Microbiol 43: 162–168.
- Jeong S, Eskandari R, Sun MP, Alvarez J, Keshari KR (2017) Real-time quantitative
 analysis of metabolic flux in live cells using a hyperpolarized micromagnetic
 resonance spectrometer. Sci Adv 3: e170034.
- Johns CW, Lee AB, Springer TI, Rosskopf EN, Hong JC, Turechek W, KokalisBurelle N, Finley NL (2017) Using NMR-based metabolomics to monitor the
 biochemical composition of agricultural soils: A pilot study. Eur J Soil Biol 83: 98–
 105.
- Johnston CA, Groffman P, Breshears DD, Cardon ZG, Currie W, Emanuel W, 31

- Gaudinski J, Jackson RB, Lajtha K, Nadelhoffer K (2004) Carbon cycling in soil.
 Front Ecol Environ 2: 522–528.
- Jones DL, Willett VB (2006) Experimental evaluation of methods to quantify
 dissolved organic nitrogen (DON) and dissolved organic carbon (DOC) in soil. Soil
 Biol Biochem 38(5): 991–999.
- Jones DL, Owen AG, Farrar JF (2002) Simple method to enable the high resolution
 determination of total free amino acids in soil solutions and soil extracts. Soil Biol
 Biochem 34: 1893–1902.
- Jozefczuk S, Klie S, Catchpole G, Szymanski J, Willmitzer L (2014) Metabolomic
 and transcriptomic stress response of Escherichia coli. Mol Syst Biol 6: 364.
- Kai F, Zhang Z, Cai W, Liu W, Xu M, Yin H, Wang A, He Z, Ye D (2017)
 Biodiversity and species competition regulate the resilience of microbial biofilm
 community. Mol Ecol 26 (21): 6170–6182
- Lee DY, Bowen BP, Nguyen DH, Parsa S, Huang Y, Mao JH, Northen TR (2012)
 Low-dose ionizing radiation-induced blood plasma metabolic response in a diverse
 genetic mouse population. Radiat Res 178: 551–555.
- Li Q, He X, Liu P, Zhang H, Chen S (2021) Synthesis of nitrogenase by Paenibacillus
 sabinae T27 in presence of high levels of ammonia during anaerobic fermentation.
 Appl Microbiol Biot 105 (7): 2889–2899
- Liang C, Amelung W, Lehmann J, Matthias K (2019) Quantitative assessment of
 microbial necromass contribution to soil organic matter. Global Change Biol 25:
 3578–3590
- Liu K, Ding X, Wang J (2020) Soil metabolome correlates with bacterial diversity and
 co-occurrence patterns in root-associated soils on the Tibetan Plateau. Sci Total
 Environ 735: 139572.
- Liu L, Wang T, Li S, Hao R, Li Q (2021) Combined analysis of microbial community
 and microbial metabolites based on untargeted metabolomics during pig manure
 composting. Biodegradation 32: 217–228.
- Liu Q, Cornelius TA, Zhu Z, Muhammad S, Wei X, Johanna P, Wu J, Ge T (2022)
 Vertical and horizontal shifts in the microbial community structure of paddy soil
 under long-term fertilization regimes. Appl Soil Ecol 169: 104248
- Liu S, Wang J, Pu S, Blagodatskaya E, Kuzyakov Y, Razavi BS (2020) Impact of
 manure on soil biochemical properties: A global synthesis. Sci Total Environ 745:
 141003.
- Ma Q, Xu M, Liu M, Cao X, Hill PW, Chadwick DR, Wu L, Jones DL (2022) Organic
 and inorganic sulphur and nitrogen uptake by co-existing grassland plant species
 competing with soil microorganisms. Soil Biol Biochem 168: 108627.

Ma Q, Kuzyakov Y, Pan W, Tang S, Chadwick DR, Wen Y, Hill PW, Macdonald A,
Ge T, Si L, Wu L, Jones DL (2021a) Substrate control of sulphur utilisation and
microbial stoichiometry in soil: Results of 13C, 15N, 14C, and 35S quad labelling.
ISME J 15 (11): 3148–3158.

Ma Q, Tang S, Pan W, Zhou J, Chadwick DR, Hill PW, Wu L, Jones DL (2021b)
Effects of farmyard manure on soil S cycling: Substrate level exploration of high- and
low-molecular weight organic S decomposition. Soil Biol Biochem 160: 108359.

- 718 Ma Q, Wen Y, Ma J, Macdonald A, Hill PW, Chadwick DR, Wu L, Jones DL (2020a)
- 719 Long-term farmyard manure application affects soil organic phosphorus cycling: A
- combined metagenomic and 33P/14C labelling study. Soil Biol Biochem 149: 107959.
- Ma Q, Wen Y, Wang D, Sun X, Hill PW, Macdonald A, Chadwick DR, Wu L, Jones
 DL (2020b) Farmyard manure applications stimulate soil carbon and nitrogen cycling
 by boosting microbial biomass rather than changing its community composition. Soil
 Biol Biochem 144: 107760.
- Ma Q, Wen Y, Pan W, Macdonald A, Hill PW, Chadwick DR, Wu L, Jones DL (2020c)
 Soil carbon, nitrogen, and sulphur status affects the metabolism of organic S but not
- its uptake by microorganisms. Soil Biol Biochem, 149, 107943.
- Ma Q, Wu L, Wang J, Ma J, Zheng N, Hill PW, Chadwick DR, Jones DL (2018)
 Fertilizer regime changes the competitive uptake of organic nitrogen by wheat and
 soil microorganisms: An in-situ uptake test using 13C, 15N labelling, and 13C-PLFA
 analysis. Soil Biol Biochem 125: 319–327.
- Maillard E, Angers DA (2014) Animal manure application and soil organic carbon
 stocks: a meta-analysis. Global Chang Biol 20: 666–679.
- Matyash V, Liebisch G, Kurzchalia TV, Shevchenko A, Schwudke D (2008) Lipid
 extraction by methyl-tert-butyl ether for high-throughput lipidomics. J Lipid Res 49:
 1137–1146.
- Mcguire KL, Treseder KK (2010) Microbial communities and their relevance for
 ecosystem models: Decomposition as a case study. Soil Biol Biochem 42(4): 529–535.
- Mcleod ML, Bullington L, Cleveland CC, Rousk J, Lekberg Y (2021) Invasive plantderived dissolved organic matter alters microbial communities and carbon cycling in
 soils. Soil Biol Biochem 156: 108191.
- Miura M, Hill PW, Jones DL (2020) Impact of a single freeze-thaw and dry-wet event
 on soil solutes and microbial metabolites. Appl Soil Ecol 153: 103636.
- Nilsson RH, Larsson K, Taylor AFS, Bengtsson-Palme J, Jeppesen TS, Schigel D,
- 745 Kennedy P, Picard K, Glöckner FO, Tedersoo L, Saar I, Kõljalg U, Abarenkov K
- 746 (2019) The UNITE database for molecular identification of fungi: handling dark taxa
- and parallel taxonomic classifications. Nucleic Acids Res 47: 259–264.

748 Orsi WD, Smith JM, Liu S, Liu Z, Sakamoto CM, Wilken S, Poirier C, Richards TA,

Keeling PJ, Worden AZ, Santoro AE (2016) Diverse, uncultivated bacteria and
archaea underlying the cycling of dissolved protein in the ocean. ISME J 10 (9):
2158–2173.

Pepe-Ranney C, Koechli C, Potrafka R, Andam C, Eggleston E, Garcia-Pichel F,
Buckley DH (2016) Non-cyanobacterial diazotrophs mediate dinitrogen fixation in
biological soil crusts during early crust formation. ISME J 10 (2): 287–298.

Piotrowska-Długosz A, Siwik-Ziomek A, D Ugosz J, Gozdowski D (2017) Spatiotemporal variability of soil sulfur content and arylsulfatase activity at a conventionally
managed arable field. Geoderma 295, 107–118.

- 758 Quast C, Pruesse E, Yilmaz P, Gerken J, Glckner FO (2012) The SILVA ribosomal
- RNA gene database project: Improved data processing and web-based tools. NucleicAcids Res 41: 590–596.
- Roberts P, Jones DL (2008) Critical evaluation of methods for determining total
 protein in soil solution. Soil Biol Biochem 40: 1485–1495.

Schmidt M, Torn MS, Abiven S, Dittmar T, Guggenberger G, Janssens IA, Kleber M,
K Gel-Knabner I, Lehmann J, Manning D (2011) Persistence of soil organic matter as
an ecosystem property. Nature 478, 49–56.

- Song Y, Li X, Yao S, Yang X, Jiang X (2020) Correlations between soil metabolomics
 and bacterial community structures in the pepper rhizosphere under plastic
 greenhouse cultivation. Sci Total Environ 728: 138439.
- Swenson TL, Jenkins S, Bowen BP, Northen TR (2015) Untargeted soil metabolomics
 methods for analysis of extractable organic matter. Soil Biol Biochem 80: 189–198.

Swenson TL, Karaoz U, Swenson JM, Bowen BP, Northen TR (2018) Linking soil
biology and chemistry in biological soil crust using isolate exometabolomics. Nat
Commun 9: 19.

- Thangarajan R, Bolan NS, Guanglong T, Naidu R, Kunhikrishnan A (2013) Role of
- organic amendment application on greenhouse gas emission from soil. Sci TotalEnviron 465: 72–96.
- 777 Timmusk S, Grantcharova N, Wagner E, Gerhart H (2005) *Paenibacillus polymyxa*
- invades plant roots and forms bioflims. Appl Environ Microb 71: 7292–7300.

Trauger SA, Kalisak E, Kalisiak J, Morita H, Weinberg MV, Menon AL, Poole FL 2nd,
Adams MW, Siuzdak G (2008) Correlating the transcriptome, proteome, and
metabolome in the environmental adaptation of a hyperthermophile. J Proteome Res 7:
1027–1035.

Tsugawa H, Cajka T, Kind T, Ma Y, Higgins B, Ikeda K, Kanazawa M,
VanderGheynst J, Fiehn O, Arita M (2015) MS-DIAL: data-independent MS/MS

- deconvolution for comprehensive metabolome analysis. Nat Methods 12: 523–526.
- Vance ED, Brookes PC, Jenkinson DS (1987) An extraction method for measuring
 soil microbial biomass C. Soil Biol Biochem 19: 703–707.
- Vives-Peris V, De Ollas C, Gómez-Cadenas A, Pérez-Clemente RM (2020) Root
 exudates: from plant to rhizosphere and beyond. Plant Cell Rep 39: 3–17.
- Vong PC, Dedourge O, Lasserre-Joulin F, Guckert A (2003) Immobilized-S, microbial
 biomass-S and soil arylsulfatase activity in the rhizosphere soil of rape and barley as
 affected by labile substrate C and N additions. Soil Biol Biochem 35: 1651–1661.
- Wang B, An S, Liang C, Liu Y, Kuzyakov Y (2021) Microbial necromass as the source of soil organic carbon in global ecosystems. Soil Biol Biochem 162: 108422.
- Wang X, Yang F, Zhang Y, Xu G, Liu Y, Tian J, Peng G (2015) Evaluation and
 optimization of sample preparation methods for metabolic profiling analysis of *Escherichia coli*. Electrophoresis 36: 2140–2147.
- Warren CR (2020) Pools and fluxes of osmolytes in moist soil and dry soil that hasbeen re-wet. Soil Biol Biochem 150, 108012.
- Woebken D, Burow LC, Behnam F, Mayali X, Schintlmeister A, Fleming ED, PrufertBebout L, Singer SW, López Cortés A, Hoehler TM, Pett-Ridge J, Spormann AM,
 Wagner M, Weber PK, Bebout BM (2015) Revisiting N₂ fixation in Guerrero Negro
- 803 intertidal microbial mats with a functional single-cell approach. ISME J 9: 485–496.
- Wei L, Ge T, Zhu Z, Luo Y, Yang Y, Xiao M, Yan Z, Li Y, Wu J, Kuzyakov Y (2021)
 Comparing carbon and nitrogen stocks in paddy and upland soils: Accumulation,
 stabilization mechanisms, and environmental drivers. Geoderma 398: 115121.
- Withers E, Hill PW, Chadwick DR, Jones DL (2020) Use of untargeted metabolomics
 for assessing soil quality and microbial function. Soil Biol Biochem 143: 107758.
- Xia J, Wishart DS (2016) Using MetaboAnalyst 3.0 for Comprehensive Metabolomics
 Data Analysis. Curr Protoc Bioinform 55, 1–91.
- Yan Z, Chen S, Dari B, Sihi D, Chen Q (2018) Phosphorus transformation response to
 soil properties changes induced by manure application in a calcareous soil. Geoderma
 322: 163–171.
- Zhu Z, Fang Y, Liang Y, Li Y, Liu S, Li Y, Li B, Gao W, Yuan H, Kuzyakov Y, Wu J,
 Richter A, Ge T (2022) Stoichiometric regulation of priming effects and soil carbon
- balance by microbial life strategies. Soil Biol Biochem 169: 108669
- 817
- 818
- 819

821 Figure captions

Fig. 1. Heat map of the 50 most significant known metabolites (detected by GC-822 TOFMS/MS) identified by ANOVA. Metabolites were clustered by Pearson 823 correlation. The colour of squares linking metabolites to samples ranges from blue to 824 red, indicating the number of standard deviations from the mean. No-M, chemical 825 fertilisers without application of manure; Low-M, medium application rate of manure 826 with chemical fertilisers; High-M, high application rate of only manure. 4-amino acid: 827 4-amino butyric acid; 4-hydro acid: 4-hydroxybenzoic acid; N-acety.: N-828 acetylmannosamine; UDP-N-acety.: UDP-N-acetylglucosamine; gly. alf. phos.: 829 glycerol-alpha-phosphate; glucose-1-phos: glucose-1-phosphate; beta-mann.: beta-830 831 mannosylglycerate.

Fig. 2. Heat map of the 50 most significant known metabolites (detected by UHPLC-832 QTOFMS/MS) identified by ANOVA. Metabolites were clustered by Pearson 833 correlation. The colour of squares linking metabolites to samples ranges from blue to 834 red, indicating the number of standard deviations from the mean. No-M, chemical 835 fertilisers without application of manure; Low-M, medium application rate of manure 836 with chemical fertilisers; High-M, high application rate of only manure. Butyl: 837 butylisopropylamine, N-N,N-dipro: N-(4-piperidinyl)-N,N-dipropylamine; 4-hydro: 838 4-hydroxy-1-(2-hydroxyethyl)-2,2,6,6-tetramethylpiperidine; 839 3-indol. Acid: 3indoleacetic acid, arach. dopam.: arachidonyl dopamine; N,N-diethyl: N,N-diethyl-2-840

indole-3-carbox.: indole-3-carboxaldehyde; 841 aminoethanol; guanid. acid: 4guanidinobutyric acid; isobu. carni.: isobutyryl-L-carnitine; 1,1-dimet.: 1,1-dimethyl-842 843 4-phenylpiperazinium; 4-amino. Acid: 4-aminobenzoic acid; 5-methv.: 5'methylthioadenosine; glycer.: glycerophosphocholine; atrazine-desis.: atrazine-844 desisopropyl-2-hydroxy; 8-oxo-2-deoxy.: 8-oxo-2-deoxyadenosine; N-epsilon-acety.: 845 N-epsilon-acetyllysine; gamma-gluta.: gamma-glutamylleucine. 846

Fig. 3. Principal component analysis (PCA) of soil carbon and nitrogen content

detected by traditional methods (A), and metabolites detected by GC-TOFMS/MS (B)

and UHPLC-QTOFMS/MS (C) under long-term (1964–2018) manure and chemical

850 fertiliser applications. Prior to analysis, the data were log₁₀ transformed. No-M,

chemical fertilisers without manure application; Low-M, medium application rate of

852 manure with chemical fertilisers; High-M, high application rate of only manure; T,

topsoil; S, subsoil.

Fig. 4. Linear correlations of metabolites detected by GC-TOFMS/MS (A) and

855 UHPLC-QTOFMS/MS (B) with soil carbon and nitrogen content detected by

traditional methods under long-term (1964–2018) manure and chemical fertiliser

applications. DOC, dissolved organic carbon; SOM, soil organic matter; EON,

858 extractable organic N; MB-C, microbial biomass carbon; MB-N, microbial biomass

859 nitrogen.

Fig. 5. The response of bacterial (A) and fungal (B) communities at phylum to genus
levels to long-term organic and inorganic fertiliser application based on a linear

discriminant effect size analysis. Only taxa meeting a linear discriminant analysis significance threshold of LDA > 3 are shown and colour-coded. The six rings of the cladogram indicate the domain (d), phylum (p), class (c), order (o), family (f), and genus (g), from inside to outside.

Fig. 6. Random forest analysis to determine factors affecting bacterial (A) and fungal
(B) community succession. The metabolites detected by UHPLC-QTOFMS/MS were
used in this analysis.

Fig. 7. Co-occurrence network of the metabolites and bacterial (A) and fungal taxa (B). The node size represented the degree in the network. Only significant Pearson correlation coefficients (r > 0.8 or r < -0.8 and p < 0.05) are shown. The metabolites detected by UHPLC-QTOFMS/MS were used in this analysis. Light purple and red lines indicate positive and negative correlations, respectively. Pink circles represent microorganisms, and green circles represent metabolites.

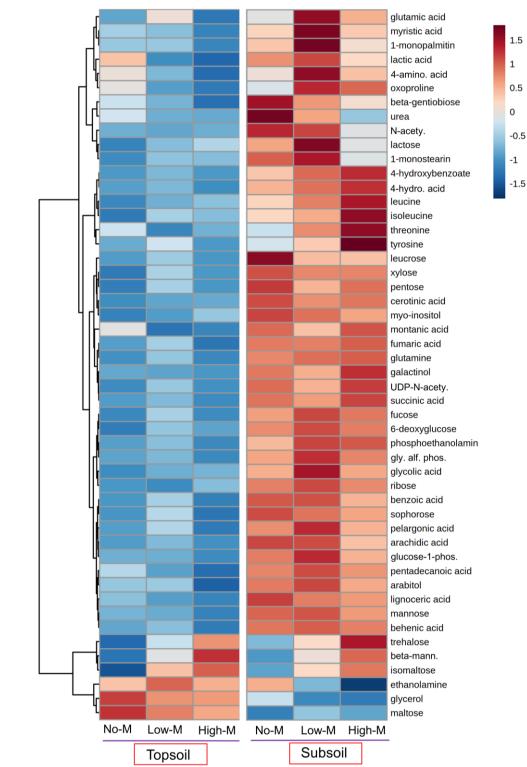
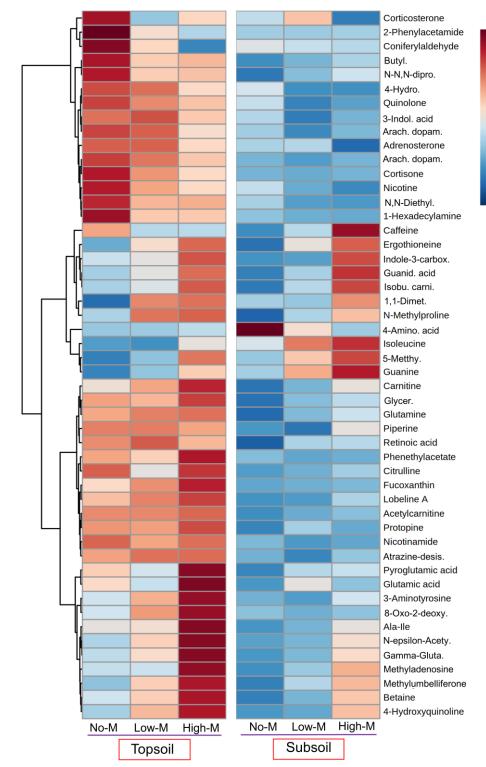






Figure 1

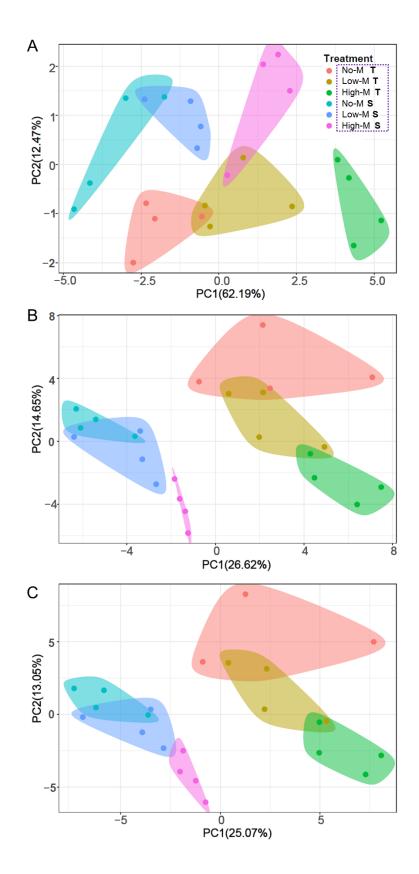




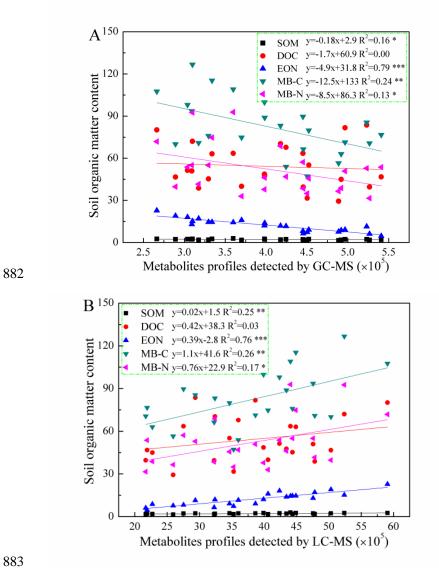
-1





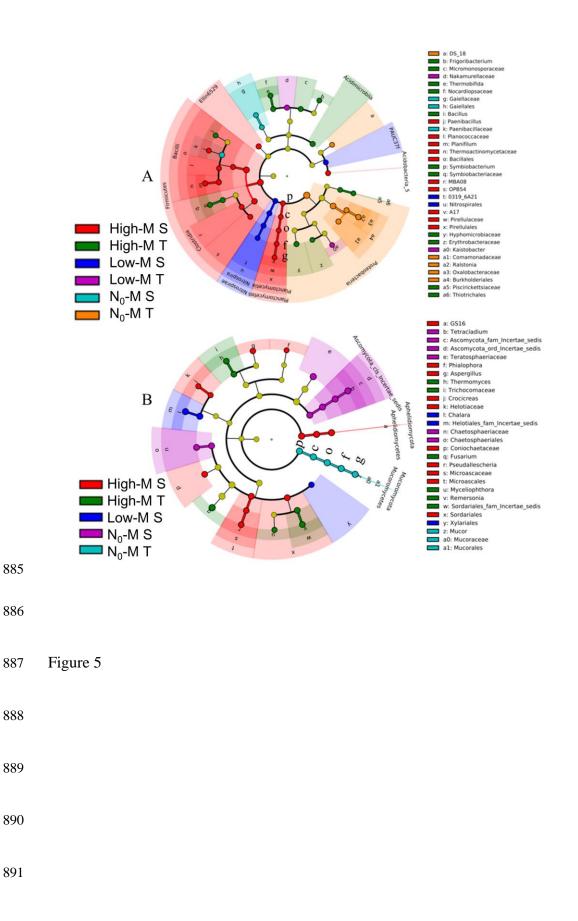


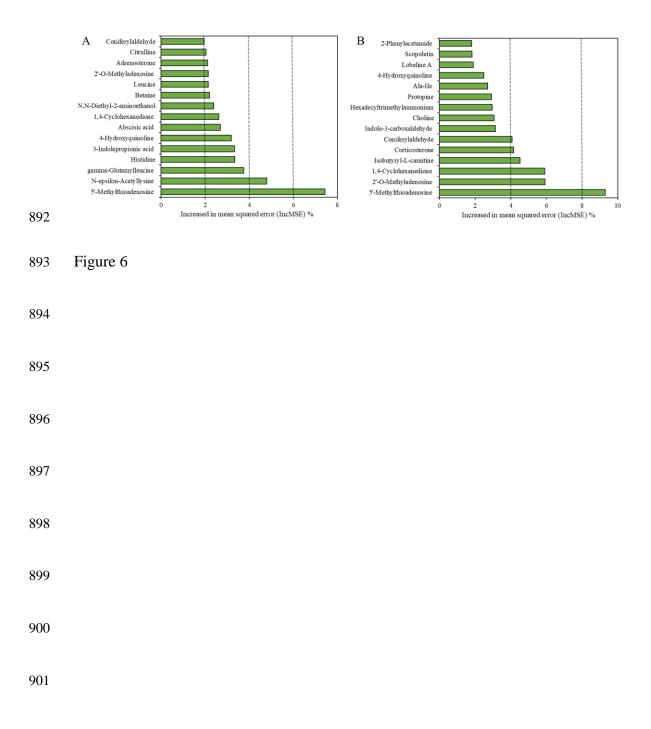
881 Figure 3

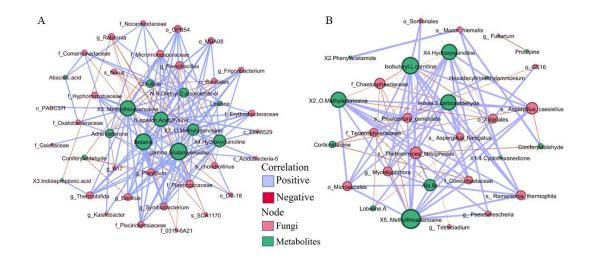




884 Figure 4







903 Figure 7