

Effect of fertilizer type on antibiotic resistance genes by reshaping the bacterial community and soil properties

Wu, Jie; Guo, Shumin; Li, Kejie; Li, Zhutao; Xu, Pinshang; Jones, Davey L; Wang, Jinyang; Zou, Jianwen

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5 Bio-organic fertilizer containing *Trichoderma* reduces antibiotic resistance

6 genes by reshaping the bacterial community and soil properties

- 7 Jie Wu^a, Shumin Guo^a, Kejie Li^a, Zhutao Li^a, Pinshang Xu^a, Davey L. Jones^{b,c}, Jinyang Wang^{a,d,*},
- 8 Jianwen Zou^{a,d}
- 9 ^a Key Laboratory of Green and Low-carbon Agriculture in Southeastern China, Ministry of Agriculture
- 10 and Rural Affairs, College of Resources and Environmental Sciences, Nanjing Agricultural University,
- 11 Nanjing 210095, China
- 12 ^b School of Natural Sciences, Bangor University, Bangor, Gwynedd LL57 2UW, UK
- 13 ^c SoilsWest, Centre for Sustainable Farming Systems, Food Futures Institute, Murdoch University,
- 14 Murdoch WA 6105, Australia
- 15 ^d Jiangsu Key Laboratory of Low Carbon Agriculture and GHGs Mitigation, Jiangsu Collaborative

- 16 Innovation Center for Solid Organic Waste Resource Utilization, Nanjing 210095, China
- 17 *Corresponding author:
- 18 Nanjing Agricultural University, Nanjing 210095, Jiangsu, China
- 19 tel.: +86 25 8439 6286; fax: +86 25 8439 5210; e-mail: jywang@njau.edu.cn
- 20

21 Abstract

22 Bio-organic fertilizers containing Trichoderma spp. play an essential role in promoting plant growth 23 and defense responses. However, the effect of bio-organic fertilizer on the prevalence of antibiotic 24 resistance genes (ARGs) in the vegetable cropping systems has been largely overlooked. Here, we 25 investigated the effects of soil properties and biotic factors on ARG profiles by analyzing ARG and 26 bacterial communities in vegetable copping soils with a long-term history of manure and bio-organic 27 fertilizer applications. The abundance of ARG in the soil was significantly increased by 116% with 28 manure application compared to synthetic NPK fertilizer application. This finding was corroborated by 29 our meta-analysis that the longer the duration of manure application, the greater the response of 30 increased soil ARG abundance. However, bio-organic fertilizer containing Trichoderma spp. 31 significantly reduced ARG contamination by 31% compared to manure application. About half of the 32 ARG variation was explained by changes in bacterial abundance and structure, followed by soil 33 properties. The mitigation of ARG by Trichoderma spp. is achieved by altering loosening the bacterial 34 community structure and weakening the close relationship between bacteria and ARG prevalence. 35 Taken together, these findings shed light on the role of bio-organic fertilizers in mitigating ARG 36 contamination in agricultural soils, which can help manage the ecological risk posed by ARG inputs 37 associated with manure applicationsources.

38 Keywords: manure; Trichoderma; antibiotic resistance genes; food chain; ecological risk

40 1 Introduction

41 The continued increase in antibiotic resistance genes (ARGs) is becoming recognized as an emerging 42 global crisis for ecological safety and human health due to the heavy use of antibiotics in agricultural 43 and healthcare settings (Gothwal and Shashidhar, 2015; Qiao et al., 2018; Zhu et al., 2013). Vegetable 44 soils are important reservoirs of ARGs in the presence of long-term manure application (Marti et al., 45 2013; Pu et al., 2020). Compared to chemical fertilizers, the application of livestock manure introduces large amounts of ARGs to soil and leads to changes in the soil bacterial communityies in the soil, 46 47 especially for topsoil and within the rhizosphere root systems (Li et al., 2022; Xie et al., 2018). However, 48 the effect of the Further, the long-term application of organic and inorganic fertilizers increases on ARG 49 abundance and diversity in greenhouse vegetable soils showed that both increased the diversity and 50 abundance of ARG, although each fertilizer regime results in but with different types of dominant ARGs 51 (Sun et al., 2019). This is of particular interest considering that the abundance of ARGs and associated 52 integrase genes and bacterial communities can be greatly affected by different fertilizer applications in 53 the soil (Liu et al., 2017; Nõlvak et al., 2016). Although many studies have revealed the distribution of 54 ARGs in fertilized soils, the variability of ARG profiles in soils with different fertilizer applications 55 remains poorly understood.

56 The bacterial community and soil properties are the major drivers associated with the variance of 57 ARGs in soil. As hosts of ARGs, the composition and structure of the bacterial community are primary contributors to the shaping of the ARG profiles in the rhizosphere-soil (Chen et al., 2020b; Guo et al., 58 59 2021). Organic fertilizer application significantly enhanced the abundance of ARGs in the soil-60 vegetable system, probably attributed to the transfer of <u>commontypical</u> ARGs among special bacterial 61 species (Huang et al., 2021; Wei et al., 2022). Furthermore, the long-term application of swine manure 62 toon the land can increase microbial diversity and reshape antibiotic resistance in the receiving soilenvironment (He et al., 2019). Notably, it has been shown that environmental heterogeneity 63 64 determines the distribution and diversity of microbial communities (Ramette and Tiedje, 2007). This implies that soil physicochemical variables shaped by different fertilizer application conditions may 65 66 indirectly affect the distribution of ARGs by influencing the bacterial community. Previous studies have

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67 identified that target ARGs such as tetracycline ARGs were closely related to soil organic carbonmatter 68 (SOC) and total nitrogen (TN) levels, factors that directly support and regulate the size of the microbial community owing to the microbiological utilization of earbon and nitrogen in agricultural soils (Zhang 69 70 et al., 2020; Zhou et al., 2017). Acidification promoted the reduction of sulfonamide genes in manure 71 by inhibiting the proliferation of sulfonamide-resistant bacteria and suppressing the accumulation of 72 int/1 (Lin et al., 2020). Other soil quality indicators, such as Soil-texture, also drives the persistence of 73 ARGs, and in particular, the clay fraction also affects the sorption ability and longevity of antibiotics in 74 soil (Hu et al., 2019; Macedo et al., 2020). Additionally, farmland application of manure is frequently 75 associated every with a higher -abundance of heavy metals (e.g., Zn, Cu). Heavy metals, as a 76 component of antibiotic resistance co-selection and transmission, promote the proliferation of ARGs 77 (Baker-Austin et al., 2006; Li et al., 2017). Hence, it is necessary to elucidate the complexity and 78 relevance of the influence of bacterial community and soil physicochemical factors on distributions of ARGs in relation to key soil quality indicators and considering the soil applied withunder different long-79 80 term-different fertilizer regimess.

81 Adding Trichoderma spp. may influence the distribution pattern of ARGs by promoting plant 82 growth and altering the physicochemical properties and bacterial community in soils. Trichoderma spp. 83 belong to a class of free-living fungi and are widely used as a safe, eco-friendly, and effective biological 84 control agent due to their ability to antagonize plant pathogenic microorganisms and stimulate plant 85 growth and defense responses (Contreras-Cornejo et al., 2009; Druzhinina et al., 2011; Sood et al., 86 2020). Inoculation with Trichoderma can also promote be considered an effective means of using 87 fertilizer use efficiencys in vegetable growing systems, and it resultings in increased the accumulation of above-and below-ground biomass-in both roots and above-ground parts of the plant as well as 88 improved crop quality (Ye et al., 2020; Yedidia et al., 2001). This is likely due to Trichoderma affecting 89 90 plant growth by promoting root development through numerous biologically active compounds or by 91 inducing an increase in auxin in the early stages of plant development (Martínez-Medina et al., 2014; 92 Vinale et al., 2008). The promotion of vegetable yields by Trichoderma maintains stable vegetable 93 vields-may also be due to enhanced soil nutrient cyclingeffectiveness (Cai et al., 2015). Evidence

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94 suggests that inoculation with *Trichoderma* altered the chemical composition of the rhizosphere soil 95 and regulated the soil microbial community composition (Zhang et al., 2019). Field trials have also 96 confirmed that the application of bio-organic fertilizer with *Trichoderma* significantly changed the 97 community structure of the dominant bacteria (Qiu et al., 2012).

98 In this study, we aimed to investigate whether the long-term application of bio-organic fertilizers 99 with or without Trichoderma spp. would alter ARG profiles in soils of vegetable cropping systems and 100 to determine the underlying mechanisms. To accomplish this, we collected soil from a vegetable field 101 with different fertilizer treatments, quantified 285 ARG subtypes using a high-throughput quantitative 102 PCR platform, and analyzed the soil bacterial community using Illumina sequencing technology. The 103 specific objectives of this study were (1) to investigate the ARG profile characteristics of vegetable soils 104 with long-term organic fertilizer application, (2) to combine meta-analysis further to reveal the patterns 105 and controlling factors of the effect of organic fertilizer application on ARGs, and (3) to elucidate the 106 effect of bio-organic fertilizer with Trichoderma spp. on ARG distribution in soil compared to manure.

107 2 Materials and Methods

108 2.1 Experimental design and sample collection

109 The field experiment was established in 2015 at the teaching and research site of Nanjing Agricultural 110 University in Nanjing, eastern China (31°43'N, 118°46'E). The climate is characterized by a subtropical 111 monsoon, with hot-rainy summers and mindless rainy winters. The annual mean temperature is 15.4°C, 112 and the average rainfall is 1106 mm. The soil was classified as a Eutric Planosol (FAO, 1981) with 32.6% 113 sand, 15.9% silt, 51.5% clay, and a bulk density of 1.33 g cm⁻³. The surface soil (0-15 cm) possessed 114 contained an organic C content of 13.548 g C kg⁻¹, a total N content of 2.08 g N kg⁻¹, and a pH (1:2.5, 115 soil/water) of 5.63. The experimental site has been in long-term continuous vegetable cultivation (e.g., 116 Brassica rapa subsp. chinensis L. and Brassica oleracea var. capitata L. Chinese cabbage and cabbage). A randomized block design with three replicates (each plot with a size of 2.7 m \times 2 m) was carried 117 118 out for with four treatments in a plastic greenhouse (30 m \times 6 m). There was a 0.6 m wide buffer row

119 between adjacent plots. The four treatments referred to the unfertilized control (control), chemical

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120	fertilizer (NPK), organic fertilizer (OF), and bio-organic fertilizer (OF+T). The organic manure used is
121	was swine manure with an containing organic matter content of 457.1 g kg ⁻¹ , total N of 14.8 g N kg ⁻¹ ,
122	phosphorus content of 25.2 g P2Os kg ⁻¹ , and potassium content of 20.1 g K2O kg ⁻¹ . The bio-organic
123	fertilizer constituted was organic manure plus 5% (mass fraction) solid Trichoderma guizhouense
124	NJAU 4742 (10 ⁹ CFU g ⁻¹ dry weight (DW), an antagonist of <i>Fusarium oxysporum</i> ; Alabouvette et al.,
125	2001). This treatment eontaining had an organic matter content of 445 g kg ⁻¹ , total N of 14.5 g N kg ⁻¹ ,
126	phosphorus <u>of</u> $10.5 \text{ g P}_2\text{O}_5 \text{ kg}^{-1}$ and potassium <u>of</u> $14.5 \text{ g K}_2\text{O} \text{ kg}^{-1}$. In each growing stage, each treatment
127	received equal amounts of <u>NPK at rates of 222 kg N ha⁻¹</u> , 150 kg P ₂ O ₅ ha ⁻¹ , and 180 kg K ₂ O ha ⁻¹ . <u>The</u>
128	OF and OF+T treatments were applied at a rate of with 6 t-ha-1 organic fertilizer, and the missing N, P,
129	and K were supplemented with urea, superphosphate, and potassium sulfate, respectively. The NPK
130	fertilizer treatment wasis applied directly to as a combination of urea, superphosphate, and potassium
131	sulfate-of the appropriate nutrient. According to the Following local farmer practice, fertilizer was
132	broadcasted on the soil surface with two-thirds of the total nutrients applied as $\underline{-\underline{a}}$ basal fertilizers before
133	planting and the remaining applied as topdressing. The vegetable planting density wasis 8 plants per
134	plot <u>(Brassica rapa</u> subsp. chinensis L).

Soil samples were collected from surface layers (0-15 cm) at three points on the diagonal of each plot at the harvest of Chinese cabbage in October 2020 and mixed and passed through a 2 mm sieve to remove stones and <u>other organic debrisimpurities (e.g. roots, crop residues)</u>. The homogenized soil samples were stored at 4°°C or -80°°C for further analyses.

139 2.2 Measurements of soil properties

- 140 Soil pH was measured using the a_pH probe (PHS-3C, Shanghai, China) at the a_soil-to-water ratio of
- 141 1:2.5 (w/v). Soil mineral N (NH₄⁺, NO₃⁻, and NO₂⁻) was determined by extracting the soil with 1 M KCl
- 142 at <u>athe</u> soil-to-water ratio of 1:5 (w/v) after shaking for <u>1 h one hour</u> on a rotary shaker. The NH₄⁺, NO₃⁻,
- 143 and NO₂⁻ concentrations in the extracts was were measured using the a Skalar San Plus segmented flow
- 144 analyzer (Skalar Analytical, Breda, Netherlands). Dissolved organic carbon (DOC) was extracted from
- 145 <u>soil using a 1:5 (w/v) using ultrapure water extract and DOC and determined using the a TOC-L</u> analyzer
- 146 (TOC L, Shimadzu, Kyoto, Japan). Soil total C and N were measured with the <u>a</u> Multi N/C 3100

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analyzer (<u>Analytik</u>Jena, <u>Jena TOC Analyzer</u>, Germany). All the extractions and digestion solutions of
 <u>containing</u> heavy metals were measured by inductively coupled plasma mass spectrometry (ICP-MS)
 (Thermo Fisher Scientific, USA).

150 2.3 DNA extraction, *Illumina* sequencing, and prediction of bacterial community function

Soil DNA was extracted from 0.25 g of fresh soil using the MoBio PowerSoil[™] DNA Isolation kit (Mo
Bio Laboratories, Carlsbad, CA, USA) according to the manufacturer's protocols. The concentration
and quality of extracted DNA were measured with a Nanodrop ND-2000 spectrophotometer (Thermo
Scientific, USA).

155 To analyze the bacterial community composition, specific V3-V4 regions of the 16S rRNA gene 156 were amplified with the primers 515F (5'-CCTACGGGNGGCWGCAG-3') and 907R (5'-157 GGACTACNVGGGTATCTAAT-3'). The resulting samples were subjected to paired-end sequencing 158 on the Illumina MiSeq300 platform (Illumina Inc., San Diego, USA). Three biological replicates were 159 performed per set. PCR was performed using diluted genomic DNA as a template using Taq DNA 160 Polymerase to ensure accurate and efficient amplification. Library quality control of PCR products was 161 performed using a Fragment Analyzer (make). After the libraries passed quality control, the results of 162 the Smear analysis (500-750 bp) were examined using the Fragment Analyzer. The mixed libraries were 163 purified by gel cutting using the QIA quick gel recovery kit (Qiagen, Venio, The Netherlands IAGEN 164 brand). The libraries were quality-checked and quantified using the Applied Biosystems-Quant Studio6 165 real-time fluorescence quantitative PCR instrument (Applied Biosystems, Waltham, MA). The raw 166 sequences were filtered and optimized using FLASH (Magoč and Salzberg, 2011). The data were 167 analyzed using the QIIME pipeline (version 1.9.1) (Caporaso et al., 2010; Kemp and Aller, 2004). The sequences were clustered into operational taxonomic units (OTUs) for taxonomic classification at the 168 169 97% similarity level using UCLUST clustering (Edgar, 2013). The taxonomy of each OTU 170 representative sequence was analyzed using the RDP Classifier (version 2.2) (Wang et al., 2007) against 171 the 16S rRNA database (Silva database version 138). Chaol index, Shannon index, and observed 172 species were used to evaluate alpha diversity for each sample (Schloss et al., 2009). Illumina raw 173 sequences for bacterial communities were deposited in the NCBI SRA under bio-project number Commented [DJ12]: Needs details here of the extracts used

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175 To better understand the potential functional contributions of the observed shifts in microbial 176 composition, we used the PICRUSt2 software to predict the functional potential of the bacterial 177 community (Langille et al., 2013). PICRUSt2 uses the resulting data for predictive analysis through the 178 Kyoto Encyclopedia of Genes and Genomes (KEGG). OTUs were normalized by dividing each known 179 16S rRNA gene copy number before functional prediction. The software for storing information on 180 Clusters of Orthologous Groups of proteins (COG) and the KEGG Ortholog (KO) family was obtained 181 by the green gene id associated with each OTU. COG and KO pathway information from the KEGG database was used to predict functional classification at three levels based on OTU abundance (Malik 182 183 et al., 2018; Morrow et al., 2015).

184 2.4 High-throughput quantitative PCR (HT-qPCR) of ARGs

185 HT-qPCR was conducted to determine the composition and abundance of ARGs and mobile genetic 186 elements (MGEs) in samples using a Wafergen SmartChip Real-time PCR system (Wafergen Inc., CA, 187 USA). A total of 296 primer sets (Table S1) were selected, including 285 ARGs to major classes of 188 antibiotics, 10 MGEs, and the 16S rRNA gene (Su et al., 2015; Wolters et al., 2018; Zhu et al., 2013). 189 The SmartChip was loaded into the Wafergen SmartChip Real-Time PCR Cycler using a PCR protocol 190 of 95 °C for 10 min, followed by 40 cycles of 95 °C for 30 s and 60 °C for 30 s. Wells with multiple 191 melting peaks or amplification efficiencies beyond the acceptable range (90-110%) were discarded. A threshold cycle (CT) value of 31 was used as the detection limit. Three technical replicates were 192 193 included for each sample, and ARGs with amplification in all three technical replicates were regarded 194 as positive quantification.

195 2.5 Meta-analysis of the response of ARGs to manure application and soil properties

To test the generality and add additional evidence for elucidating the underlying response of manure and soil physicochemical properties affecting the variance of ARGs in vegetable soil, we performed a systematic search using the ISI Web of Science (Thomson Reuters, New York, NY, USA), Google Scholar (Google, Mountain View, CA, USA) and China National Knowledge Infrastructure (CNKI, Beijing, China). Our search terms included ("soil*" OR "vegetable soil*") AND ("antibiotic resistance gene*" OR "ARG*" OR "resistome*"). Studies included met the following criteria: (1) both field and pot studies were selected; (2) studies had to be replicated; (3) the means, standard errors, and replication of the variables could be extracted directly from the text, tables, or digitized graphs. The following information was documented for each study: the type and abundance of ARGs, the duration of the experiment, the type of fertilizer, and soil physicochemical properties. A total of 294 observations were extracted from 15 papers (Text S1).

207 The data-set was analyzed using a meta-analytic technique described previously (Hedges et al., 208 1999). The effects of manure and soil physicochemical properties on the variance of ARGs were 209 quantified using the natural log of response ratio (ln*R*), which was calculated as $\ln R = \ln(Xc/Xt)$, where 210 Xc and Xt represent the abundance of ARGs from the control and manure-added treatments, 211 respectively. We performed a meta-analysis using the Metawin (version 2.1) (Rosenberg et al., 2000). 212 Similar to the previous research (van Groenigen et al., 2011), we used the number of replications for 213 weighting. The effects of manure addition were considered significant if the confidence intervals (CIs) 214 did not overlap with zero. Means of categorical variables were considered significantly different if their 215 95% CIs did not overlap. To facilitate ease-interpretation, the results from our analyses were back-216 transformed and reported as the percentage change under manure addition ([lnR-1]×100) in the main 217 text.

218 2.6 Statistical analysis

219 All analyses were performed in R version 4.0.3 (R Core Team, 2020). The CT measured by the WaferGen qPCR was used to calculate the copy number of genes via Copy Number = $10^{(30-CT)/(10/3)}$ (Fu 220 221 et al., 2021; Stedtfeld et al., 2008). Circos and heatmap diagrams were generated to show the ARG 222 classification in the different treatments. The differences in bacterial alpha diversity were determined 223 using a one-way analysis of variance ANOVA followed by Tukey's test was used to determine 224 differences between individual treatments. Non-metric multidimensional scaling (NMDS) analyses 225 based on the Bray-Curtis distance were performed to visualize the overall pattern of bacterial 226 communities, and the permutational multivariate analysis of variance (PERMANOVA) was used to

227	analyze the dissimilarity in soil microbial diversity based on the OTU in different samples (Legendre
228	and Gallagher, 2001). The Mantel test and Procrustes analysis were used to determine the correlation
229	between ARGs and the bacterial community (Delgado-Baquerizo et al., 2018; Peres-Neto and Jackson,
230	2001). The Hmisc package within R was used to calculate Spearman's correlation matrix between
231	bacterial OTUs and ARGs, and the correspondences of the bacteria and ARGs were shown by network
232	analysis on the Gephi platform (version 0.9.2) (Li et al., 2015). Redundancy analysis (RDA) and
233	variance partitioning analysis (VPA) were applied to investigate how ARGs were related to the
234	components of bacterial community and soil physicochemical properties.

235 3 Results

236 **3.1 Distribution of ARGs in vegetable soil under long-term fertilization**

Diverse ARGs and MGEs were detected in all samples, covering major antibiotics classes, including aminoglycoside, beta-lactamase, chloramphenicol, macrolide-lincosamide-streptogramin B (MLSB), multidrug, sulfonamide, tetracycline, and vancomycin resistance genes (Fig. 1a). Aminoglycoside resistance genes were the most abundant ARG type in OF, while multidrug resistance genes were the most abundant in the NPK and OF+T treatments. The number of the major classes of ARGs ranged from 3 to 42 (Fig. 1b). Across all treatments, the number of ARGs detected was around 60 and independent of the type of fertilizer application.

244 Overall, the abundance of ARGs in the soil receiving applied with organic fertilizer was universally 245 higher than in the NPK and control treatments. Interestingly, we found that the addition of Trichoderma 246 effectively reduced prompted the alleviation of the prevalence of ARGs contamination in in the soil. For 247 example, the OF+T treatment significantly reduced the abundance of ARGs by 30.9% overall compared 248 to the OF treatment without *Trichoderma* (P < 0.05). This reduction was , and in pp articularly evident 249 for aminoglycoside, tetracycline, sulfonamide, and beta-lactamase resistance genes which reduced by 250 50.5, 40.7, 35.9, and 17.2%, respectively (Fig. 1c). In contrast, OF+T increased the-ARG abundances 251 by 49.4 and 14039.9% relative to compared to the NPK and the control treatments, respectively. In addition, the OF treatment contained the highest average ARG abundances which were elevated and 252 253 increased by 116.2 and 247.1% compared to the NPK and the control treatments, respectively, ... This 10

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was particularly <u>the case</u> for MLSB, tetracycline, sulfonamide, aminoglycoside, and beta-lactamase resistance genes. NPK <u>fertilizer application</u> increased ARG abundances by 60.5% compared to <u>the</u> control. Hierarchical clustering of their abundances demonstrated a clear division between <u>the</u> different fertilize<u>r regimesd soils</u>, and phylogenetic analysis revealed the clustering of soils with and without organic fertilizer <u>application</u> was distinguished (Fig. 1c).







270 3.2 Meta-analysis

271 Based on 294 paired measurements of 15 studies, our meta-analysis showed that manure application 272 significantly increased ARG abundances by an average of 51.9% in vegetable cropping soils (95% CIs: 273 31.0% to 77.9%; Fig. 2a). Specifically, with respect toon the time-scale, ARG contamination was more 274 severe in vegetable soil where manure had been applied for more than three years (1610.8%, 95% CIs: 275 140.4% to 187.4%). Manure application showed a significant and positive effect on aminoglycoside 276 and sulfonamide resistance gene<u>abundances</u> (375-2%, 95% CI: 238-0 to 6832-7%; 1165-9%, 95% CI: 277 53.1 to 222.1%), whereas a slightly but nonsignificant increment of ARG abundances was found in 278 other categories. Notably, fertilizer type had a significant impact on ARG abundances, especially in 279 swine manure (263-2%, 95% CIs: 156-2% to 426-87%) and mixed manure (87-8%, 95% CIs: 487-7%) 280 to 14039.8%). Regression analysis illustrated that the changes in the manure application effect on 281 abundance of ARGs were negatively correlated with soil pH (r = -0.073, P < 0.01; Fig. 2b) and soil 282 organic C (r = -0.154, P < 0.05; Fig. 2d), but <u>had</u> a positive correlation with clay content (r = 0.235, P 283 < 0.01; Fig. 2c) and soil C/N ratio-of soil (r = 0.146, P < 0.05; Fig. 2e).



Fig. 2 Effect of manure application on ARG abundances in vegetable <u>cropping</u> soils. Responses of ARG abundances to manure application in soils (a) and its relationships with soil pH (b), clay content (c), organic C content (d), and C/N ratio (e). The category is divided according to the duration (year), type of fertilizers, and type of ARGs. Values are means ± 95% confidence intervals. The number of

289 observations is given beside each category. The vertical dashed line is drawn at zero. The sizes of the

symbols grouped by ARG type are drawn proportional to the weights in the meta-regression analysis.

291 **3.3 Diversity and composition of the bacterial community in soils**

A total of 1,803,910 high-quality sequences were detected across all samples, with sequences per sample ranging from 131,456 to 199,430. A total of 5518 OTUs were obtained for 16S rRNA based on a 97% similarity cut-off. Evaluation of the Specie richness, Chao1 index, and Shannon index demonstrated that the OF treatment showed the highest alpha diversity of the bacterial community, followed by the control, OF+T, and NPK (Fig. 3a-c).

297 The bacterial communities were classified into 18 phyla-from the soil (Fig. 3d). Proteobacteria 298 (43.3-53.9%), Firmicutes (6.4-15.2%), Bacteroidetes (7.4-10.1%), Actinobacteria (5.9-9.3%), and 299 Acidobacteria (2.8-9.2%) were the main phylum in all treatments. Proteobacteria was the dominant 300 bacterial phylum, and adding Trichoderma significantly increased the relative abundance of 301 Proteobacteria in compared comparison to the OF treatment, whereas a decrease occurred in the relative abundance of Acidobacteria and Planctomycetes (all P < 0.05; Figs. 3d and S1a). Compared with the 302 303 NPK treatment, OF+T or OF significantly increased the relative abundance of Proteobacteria, 304 Actinobacteria, Acidobacteria, and Planctomycetes but significantly reduced the relative abundance of 305 Firmicutes, Gemmatimonadetes, and Candidatus Saccharibacteria (P < 0.05-0.01). The application of 306 different fertilizers did not greatlyremarkably change the relative abundances of another bacterial phylum in soil (P > 0.05). 307

The NMDS ordinations based on the Bray-Curtis dissimilarity matrices revealed that the overall patterns of the bacterial community were clustered and distinctly separated from different treatments (Fig. 3e), which was further supported by the Adonis test ($R^2 = 0.64$, P < 0.01). Meanwhile, the clusters of bacteria in soils applied with manure were closer to the than-NPK treatment than tound the control. Commented [DJ16]: The response in panel c looks nonlinear

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Others could be 'other ARGs' ?



Fig. **3** Bacterial diversity and community composition in vegetable cropping soils with different fertilizer regimes. The α -diversity (Specie richness, Chao1 index, and Shannon index) (a, b, and c) and percentage of dominant phylum (d) levels of soilthe bacteria among the different fertilizer treatments in soils. Non-metric multidimensional scaling (NMDS) ordination plot of the Bray-Curtis dissimilarity matrices between different treatments and the <u>An</u>donis PERMANOVA analyses of bacterial distribution among treatments (e). The treatments included a no-fertilizer control (control), chemical fertilizer (NPK), organic fertilizer (OF), and bio-organic (*Trichoderma*) fertilizer (OF+T).

Bacterial community functions in vegetable-soil systems were predicted by PICRUSt2 using the 16S rRNA gene sequencing data to understand the differences in community potential functions between treatments. In total, 4020 KEGG orthologs with known and unknown functions involved in 407 level 3 KEGG ortholog pathways were identified in this study. These pathways were found to be Commented [DJ17]: Adonis not spelt right in panel e.

Specie in panel a should be Species

involved in the seven most dominant KEGG functional categories consisting of cellular processes,
environmental information processing, genetic information processing, human diseases, metabolism,
and organismal systems, and not included in pathway or <u>BRITE hierarchiesbrite</u>, and the metabolism
was the main pathways accounting for 37% in soils (Fig. S2).

328 We screened out the relevant functions involved in carbon, nitrogen metabolism, and resistance 329 pathways (Fig. S3). Notably, the relative abundance of predicted functional genes involved in nitrogen metabolism was significantly reduced in the OF+T treatment (P < 0.05). Similarly, the abundance of 330 331 pathways associated with carbon metabolism was the lowest but did not reach significance in the 332 presence of Trichoderma. Furthermore, these results demonstrated that although the manure application 333 maintained the resistance-related bacterial functional mostly stable, the relative abundance of functional 334 genes assigned to beta-lactamase resistance, cationic antimicrobial peptide resistance, vancomycin 335 resistance, antifolate resistance, and insulin resistance was found to decline in OF+T.

336 3.4 Driving forces for the distribution of ARGs in vegetable soil

Procrustes analysis and Mantel test were conducted to investigate the relationship between bacterial communities and ARG profiles (Fig. 4a). Our study indicated that the structure of bacterial composition (Bray-Curtis dissimilarity of OTU tables) fitted well with ARG profiles (Bray-Curtis dissimilarity) relative according to the different treatments. There was a significant association between bacterial communities and ARG profiles ($M^2 = 0.23765$, P < 0.05, permutations = 999). The Mantel test further verified this finding based on Spearman's rank (r = 0.1693, P < 0.05, permutations = 999), indicating the bacterial structures and composition influenced the ARG profiles in the soil.

Network analysis was performed to illustrate the detailed relationships between individual ARG subtypes and the potential host bacteria at the phylum levels (Fig. 4b). There were 50 nodes and 46 edges in general, and the average degree and modularity index were 1.84 and 0.76, respectively. The density of each node indicated the co-occurrence patterns between these bacteria and ARGs in soil. The data suggests that the bacterial taxa *Proteobacteria*, *Bacteroidetes*, *Actinobacteria*, *Acidobacteria*, and *Planctomycetes* may be the ARG hosts a<u>snd they</u> were <u>positively</u> intensively-correlated with multiple ARGs. Specifically, as the most pervasive taxa, *Proteobacteria* had the most edges with ARGs and were **Commented [DJ18]:** This needs splitting into 2 sentences

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significantly related to *tet*Q (tetracycline resistance genes). <u>It should be noted that o</u>One ARG might be carried by <u>a</u> diverse <u>range of potential bacterial</u> hosts; the *vgb*B-02 (MLSB resistance genes) were significantly related to *Proteobacteria* and *Bacteroidetes*. Meanwhile, *Actinobacteria* and *Acidobacteria* were significantly related to *lnu*B-01 (MLSB resistance genes) and *mph*A-02 (MLSB resistance genes), respectively. *Planctomycetes* were significantly related to *mph*A-02 and *bla*_{VEB} (betalactamase resistance genes). In contrast, the relationship between other bacterial phylum and ARG types was generally weak.



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359 Fig. 4 Association of ARGs distribution with the bacterial community in vegetable soils with 360 contrasting fertilizer regimes. Procrustes analysis, Mantel test, and Network analysis were used to 361 revealunveiled the significant associations between the ARG abundances and bacterial community at 362 the phylum levels. Procrustes test between ARG profiles (Bray-Curtis) and bacterial communities using 363 taxonomic dissimilarity metrics (Bray-Curtis), where M2 represents the sum of the squared deviations 364 (vector residuals) over the first two dimensions. Solid and hollow circles represent ARGs and bacterial 365 OTUs, respectively. The connection between ARGs and bacterial taxa represents a strong (Spearman's 366 correlation coefficient r > 0.7) and significant (P-value < 0.01) correlation. The nodes are colored according to bacterial and ARG types. The node size is proportional to the number of connections, and 367 368 the edge thickness is proportional to the correlation coefficient. The treatments included a no-fertilizer 369 control (control), chemical fertilizer (NPK), organic fertilizer (OF), and bio-organic (Trichoderma) 16

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370 <u>fertilizer (OF+T).</u>

371 The RDA and VPA analyses were performed to further explore further the linkage between ARGs 372 with the bacterial community and soil environmental variables (Fig. 5). Overall, the results showed that 373 the selected variables could explain 88.1% of the variance in ARGs with the first two axes of RDA (Fig. 374 5a). There was a positive correlation between tetracycline resistance gene abundances and MGEs and 375 pH, SOC, C/N, bacterial community structure (i.e., NMDS1, NMDS2), and multiple heavy metals (e.g., 376 Cu and Cd). Bacterial abundance and nitrogen (e.g., NH4⁺, NO2⁻, and NO3⁻) were the top two factors 377 affecting associated with changes in the abundance of MLSB, sulfonamide, beta-lactamase, and 378 vancomycin resistance genes in soil, which showeding significantly and positively correlations. To 379 differentiate the effects of the biotic and abiotic factors on the changes in the ARG profiles, VPA showed 380 that a total of 73.5% of ARG variations could be explained by them (Fig. 5b). Specifically, the soil 381 properties contributed 29.6% of ARG variances, and the bacterial communities, including abundance 382 (27.3%) and structure (16.6%) contribute nearly half of ARG profiles in soil.

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Fig. 5 Main factors driving the distribution of ARGs in vegetable soils. Redundancy analysis (RDA) (a) and variation partitioning analysis (VPA) (b) depict the contribution of biotic and abiotic factors to the variation of ARGs in soil with contrasting fertilizer regimes. Using bacterial abundance, bacterial community structure (two axes of NMDS), and soil physicochemical characteristics as explanatory variables. The blue and red arrows represent the explanatory and response variables, respectively, where

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greater lengths indicate stronger correlations between that pivotal factor and ARG distribution. The angles between arrows reveal the correlations between respective environmental parameters and individual ARGs. The factors that have a significant impact are specified in the VPA. <u>The treatments</u> included a no-fertilizer control (control), chemical fertilizer (NPK), organic fertilizer (OF), and bioorganic (*Trichoderma*) fertilizer (OF+T).

394 4 Discussion

395 4.1 Manure application increased the abundance of ARGs in vegetable soils

396 Our findings from the field experiment are in accordance with the results of the meta-analysis that 397 showed that manure application significantly increases d the ARG accumulation in soil (Figs. 1 and 2a). 398 These results are , which is also consistent with previous studies (Macedo et al., 2020; Udikovic-Kolic 399 et al., 2014). Meta-analysis revealed that ARGs were strongly enriched in soils receiving - applied with 400 manure and that this effect became stronger continued to increase over repeated applicationstime, 401 especially over three years. The prolonged misuse of antibiotics in livestock production systems and 402 subsequent ______accumulation the continued application of in manure applied to soil will inevitably 403 lead to pose a great risk to the widespread presence and accumulation of ARGs in agricultural soils (Liu 404 et al., 2021; Peng et al., 2017). Notably, most studies that met our requirements for meta-analysis were 405 short-term pot trials, so we further validated this result using a five-year field experiment with manure applications. Our field research and meta-analysis found that sulfonamide and aminoglycoside were 406 407 more easily enriched than other ARGs, which might be related to their common use in soils (Wang et 408 al., 2020; Zhao et al., 2018). For the type of manure, swine manure application had the most significant 409 increase in ARG abundances, which is consistent with the OF treatment. The higher ARG contamination 410 caused by swine manure application may be associated with the high addition of antibiotics in swine 411 feed (Martínez-Carballo et al., 2007; Qian et al., 2018; Zhao et al., 2018).

The changes in the distribution of ARGs may be closely related to the altered soil physicochemical characteristics after manure application. We observed that the abundance of ARGs was negatively correlated with pH and SOC and positively correlated with C/N and <u>Clay-clay content</u> (Fig. 2), implying that higher ARG contamination may occur in acidic environments and soils with lower nutrient levels. **Commented [DJ23]:** I think it doesn't say 5 years in the Materials and methods

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416 The lower TN confirms the reliability of this result in the OF compared to NPK (Table S2, $P_{-} \leftarrow \leq 0.05$), 417 although it was not reflected in pH and SOC. An increase in pH leads to a decrease in antibiotic sorption 418 by the soil, facilitating the removal of antibiotic residues and reducing the abundance of the 419 corresponding ARGs (Li et al., 2020; Wang et al., 2018). Soil nutrient factors (e.g., SOC and TN) may 420 also synergistically influence the composition of ARGs in soils (Sun et al., 2019). Meanwhile, soil 421 texture drives the persistence of ARGs, and clay contributes the most to ARGs distribution than other 422 properties in soils (Macedo et al., 2020; Wang et al., 2020). Animal feed additives and their subsequent 423 presence in Manure manure fertilization could introduce heavy metals into soil (Zhao et al., 2018), and 424 Here -we found evidence forgenerally higher levels of multiple heavy metals in the OF treatment 425 (Table S2). Heavy metals can act as a selection pressure forcing the evolution and spread of heavy metal 426 and antibiotic resistance, leading to the proliferation of ARGs in fertilized soil (Dickinson et al., 2019; 427 Komijani et al., 2021; Seiler and Berendonk, 2012).

428 The association between host bacteria and ARGs may contribute to this difference in the profile of 429 ARGs. Proteobacteria and Planctomycetes are widely distributed and abundant components of the 430 microbial community in soil amended with manure fertilizers (Bonanomi et al., 2016; Buckley et al., 431 2006). As the most prevalent predicted source phylum of ARGs, the Proteobacteria and Planctomycetes were the major drivers of shifting ARG profiles in soil (Forsberg et al., 2014; Pu et al., 2020). As 432 mentioned in the results above (Fig. 3e), the application of manure significantly increased the relative 433 434 abundance of multiple bacterial phyla (e.g., Proteobacteria, Actinobacteria, Acidobacteria, and 435 Planctomycetes). It is evident from the network that the increased abundance of ARGs in the manured 436 soil may be due to the enhanced abundance of MLSB resistance genes carried by the above host bacteria, 437 tetracycline resistance genes carried by Proteobacteria, as well as beta-lactamase resistance genes 438 carried by Planctomycetes (Figs. 1 and 4), and this further supports this explanation.

439 4.2 Application of bio-organic fertilizers reduced ARGs by modifying biotic and abiotic factors

440 Bio-organic fertilizers reducedweakened the propagation and transmission of ARGs compared to 441 manure application alone, inferring that the change in soil physicochemical properties might be partly 442 responsible. Trichoderma is very efficient in improving soil fertility and the promoting more 19

443 effectiveness efficient cycling of soil nutrients (Bulluck et al., 2002; Caporale et al., 2019). Analysis 444 based on functional predictions objectively revealed that microorganisms associated with soil carbon 445 and nitrogen metabolism were less abundant after bio-organic fertilizer application, consistent with the 446 results of higher SOC and TN in the OF+T (Fig. S3 and Table S2). Combined with the meta-analysis, 447 it can be we concluded that more increased carbon and nitrogen made less reduced ARG pollution (Fig. 448 2). Meanwhile, the co-selection mechanisms for metals and ARGs have been demonstrated (Baker-449 Austin et al., 2006; Li et al., 2017). There was a significant effect of Cu and Cd on the distribution of ARGs in the present study (Fig. 5), as reported by previous studies (Arya et al., 2021; Seiler and 450 451 Berendonk, 2012). The concentration of microelements was significantly increased in the roots of plants 452 planted in soil inoculated with Trichoderma (Yedidia et al., 2001). Hence, the reduction in heavy metal 453 concentrations in the OF+T treatment due to plant uptake may have weakened the abundance of ARGs 454 compared to OF (Table S2). Additionally, MGEs have a lower abundance in the OF+T treatment may 455 also be an essential factor in attenuating ARG contamination by Trichoderma (Fig. 1c).

456 In general, the composition and structure of bacterial communities largely determine the profile of 457 ARGs in soils with manure application owing to the role of bacteria as the primary carriers of ARGs 458 (Huang et al., 2021; Nõlvak et al., 2016; Xie et al., 2018). Interestingly, this study demonstrated that 459 bio-organic fertilizer containing Trichoderma may affect the composition of bacterial communities and 460 thereby reduced the presence of ARGs. This may be due to the ability of bio-organic fertilizers 461 containing antagonistic microorganisms to alter bacterial abundance (Berg, 2009; Cai et al., 2015; Qiu 462 et al., 2012). Evidence of long-term field experiments suggests that bio-organic fertilizers can stimulate potentially beneficial bacteria to improve crop yield and achieve disease suppression by reshaping key 463 species in the structure and function of soil microbial communities (Qiao et al., 2019; Xiong et al., 464 465 2017). This reshapingimprovement of the microbial communities by Trichoderma may have-make a 466 significant contribution to the mitigation of ARGs. Notably, the relative abundance of Planctomycetes 467 carrying beta-lactamase resistance genes was lower in the OF+T treatment, which could be one of the 468 reasons for the low ARG content compared to manure application alone (Figs. 3d and 4b). 469 Correspondingly, the addition of Trichoderma increased the relative abundance of Proteobacteria but

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470 reduced the tetracycline resistance genes carried by them, likely owing to the replacement of its function 471 as host bacteria for ARGs by other microorganisms. Rare microbial taxa have been shown to have a high proportional role in biological processes in soil applied with Trichoderma, although Proteobacteria 472 473 is the most abundant and diverse phylum in this study (Chen et al., 2020a; Spain et al., 2009). Bacterial 474 function predictions further validated that the bacteria associated with resistance decreased in the soil 475 applied with bio-organic fertilizer (Fig. S3), which coincides with the above association of bacterial 476 abundance with the distribution of ARGs. Our findings confirm that applying bio-organic fertilizer benefits soil health by mitigating ARG pollution while ameliorating soil microorganisms. 477

478 The discrepancies in ARG profiles may also be caused by differences in bacterial structure (Guo 479 et al., 2021). Our results indicated that the contribution of bacterial structure (16.6%) to the distribution of ARGs is slightly less than that of bacterial abundance (27.3%). The two axes of the NMDS can 480 481 explain 16.6% of the ARG variance across all impact factors (Fig. 5b). Although the bacterial structural 482 composition of OF+T was tighter compared to NPK, looser than the OF treatment (Fig. 3e), this 483 phenomenon is also present in the relationship between bacterial communities and ARG (Fig. 4a). This 484 suggested that the added Trichoderma has weakened the close relationship between bacteria and ARGs. 485 The alteration of the bacterial structure by Trichoderma may derive from its competitive mechanism with other microorganisms in natural communities (Ferreira and Musumeci, 2021; Sood et al., 2020). 486 487 Additionally, the effectiveness of bio-organic fertilizer in modifying the soil may be attributed to the 488 enhancement of soil microbial activity (Ye et al., 2020). As such, we speculated the activity of 489 microorganisms associated with the host bacteria (e.g., Proteobacteria) carrying ARGs might not have 490 been enhanced in the OF+T treatment. However, further research is needed to support this hypothesis. 491 Nevertheless, a proportion of the variance in ARG profiles remains unexplained. Overall, further 492 follow-up studies characterizing the linkages in soil microbial community structure and composition 493 and ARG would be beneficial for a better understanding of the distribution and dissemination of ARGs 494 in organic fertilizer-soil-plant systems.

495 5 Conclusion

496 In summary, our field experiment and meta-analysis indicated that manure application significantly

497 increased the abundance of ARG in vegetable soils compared to synthetic fertilizer application. 498 However, bio-organic fertilizer with Trichoderma significantly reduced ARG contamination compared to manure application. The abundance and structure of the bacterial community and soil properties were 499 the main drivers of ARG contamination. The mitigation of ARG by Trichoderma was accomplished by 500 501 loosening reshaping the bacterial community structure and weakening the close relationship between 502 bacteria and ARGs. Given the importance of ARGs to soil ecological risk, these findings deepened our 503 understanding of ARG profiles in soil under long-term fertilization and highlighted the importance of 504 potentially controlling the prevalence of ARGs by-using bio-organic fertilizers.

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