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Setting evolutionary-based conservation priorities for a phylogenetically data-poor taxonomic group (Scleractinia)


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Keywords
coral; data-poor; evolutionary distinctiveness; ED score; prioritization; phylogenetic diversity; Scleractinia.

Abstract
Given the current extinction crisis coupled with the shortfall in funding, there is a pressing need to establish species conservation priorities. The prioritization of phylogenetic diversity and evolutionary distinctiveness is one approach; however, taking such an approach requires more phylogenetic data than are currently available for most taxa. Here, we investigate the effects of increased phylogenetic knowledge on the accuracy of evolutionary distinctiveness (ED) scores over time using scleractinian corals as a case study. ED scores were calculated from four molecular-based phylogenies from 2008 to 2013, each one representing a chronological step of increased phylogenetic knowledge for scleractinian corals, finally resulting in a full species-level phylogeny which is used here as the reference dataset. As expected, the most complete and up-to-date phylogenies performed well at predicting scores taken from a recent, full-coverage species-level phylogeny of scleractinian corals. Surprisingly, however, older phylogenies and scores derived from expert opinion also performed well. More unexpectedly, the expert opinion-led scores, when used as a basis for imputing scores for missing species, achieved a close second in terms of prediction accuracy compared with the most recent and largest tree, which had nearly 10 times more taxonomic coverage. We recommend, once tested further, that ED score imputation be considered for assessing the conservation priorities for other poorly studied groups.

Introduction
Current species extinction rates across all taxa are overwhelming background rates and could increase further (Pimm et al., 2006; Barnosky et al., 2011; De Vos et al., 2014; WWF 2014). With high numbers of species under threat of extinction and a global shortfall in funding (Waldron et al., 2013), there is a pressing need to establish conservation priorities among threatened species (e.g. Bibby et al., 1992; Mittermeier et al., 1998; Olson & Dinerstein, 1998; Myers et al., 2000; O’Connor, Marvier & Kareiva, 2003; Possingham & Wilson, 2005; Brooks et al., 2006; Margules & Sarkar, 2006; Isaac et al., 2007).

A large number of conservation priority-setting approaches have been documented in the literature (e.g. Brooks et al., 2006; Wilson et al., 2006; Pressey et al., 2007;
One species-based method considers the evolutionary relationships within a clade of target species (Vane-Wright, Humphries & Williams, 1991; Faith, 1992, 1996, 2007) alongside threat status to set conservation priorities (Witting & Loeschcke, 1995; Redding & Mooers, 2006; Isaac et al., 2007, 2012; Safi et al., 2013). The goal of prioritizing evolutionarily distinct species is to maximize the number of evolutionary trajectories conserved (Vane-Wright et al., 1991; Crozier, 1992). This precautionary, maximization approach is necessary as future evolutionary radiations are difficult to predict (Cooper & Fortey, 1998), which makes picking evolutionary ‘winners’ extremely difficult. Conservation actions based on measures of species-specific evolutionary isolation have been shown to conserve a greater proportion of the overall phylogenetic diversity of the target clade (Redding et al., 2008) and can preserve some unusual life-history and morphological traits, potentially helping to protect other aspects of biodiversity within a clade (Cadotte, Cardinale & Oakley, 2008; Cadotte et al., 2009; Redding, Dewolff & Mooers, 2010). A conservation programme that has applied this approach is the Evolutionarily Distinct and Globally Endangered (EDGE) of Existence programme (Isaac et al., 2007; http://www.edgeofexistence.org). The EDGE ranking system is a composite score based on two components: a measure of evolutionary distinctiveness (ED) and a measure of global endangerment (GE) or threat (Redding & Mooers, 2006; Isaac et al., 2007; Mooers, Faith & Maddison, 2008).

To date, this has been applied to mammals, amphibians and birds (Isaac et al., 2012; Safi et al., 2013; Jetz et al., 2014). While it may be the case that evolutionary-based conservation approaches offer a data-cheap method of assessing several aspects of biodiversity for the target clade concurrently (Redding et al., 2010), taking such an approach has only been attempted with data-rich species groups (e.g. mammals, birds) as it requires more phylogenetic data than are currently available for most taxa. Indeed, little work has been carried out to test how phylogenetic priority setting can be applied to relatively data-poor taxonomic groups, in particular those that have not been subjected to the same intensive regimes of genetic sequencing carried out for large-bodied vertebrates.

The threats to tropical coral reefs have been reported extensively (Sheppard, 2003; Donner et al., 2005; Hoegh-Guldberg et al., 2007; Veron et al., 2009; Dove et al., 2013; Kennedy et al., 2013). In 2006 and 2007, the threat status of the 827 reef-building scleractinian species were assessed by the world’s leading coral experts for the International Union for the Conservation of Nature (IUCN) Red List (Carpenter et al., 2008). Of the 688 species not deemed to be ‘data deficient’, 32.7% were listed in an elevated threat category (‘vulnerable’, ‘endangered’, ‘critically endangered’). In 2010, the Zoological Society of London set evolutionary-based conservation priorities for scleractinian corals, a group which, at that time, lacked detailed phylogenetic data. The assessment was based on available phylogenies (Kerr, 2005; Fukami et al., 2008) and taxonomy, combined with expert opinion (as outlined in the succeeding text; hereafter called the 2010 EDGE Coral Expert Workshop). A later iteration of EDGE scores was then published based on a more recent species-level phylogenetic supertree (Huang, 2012). This provides a unique opportunity to validate the expert-led approach, while at the same time investigate how limited phylogenetic knowledge may affect our confidence with which species are chosen as priority species for conservation attention at each stage of assessment within the EDGE framework.

Here, for the first time, we investigate setting evolutionary-based conservation priorities for a previously data-poor species group: scleractinian corals. We examine the role of increasing phylogenetic information on improving the accuracy of ED scores using five empirically derived datasets that represent chronological steps of increasing phylogenetic coverage: (1) 127 species two-gene phylogeny (Fukami et al., 2008); (2) Expert opinion added to the previous 127 species phylogeny; (3) 234 species single-gene phylogeny (Kitahara et al., 2010); (4) 466 species multi-gene phylogeny (Huang & Roy, 2013); (5) 838 species supertree (Huang, 2012). For each of these datasets, we determined the consistency of top 50 scoring species and then measured the correlation of the ED scores to the most recent ‘reference’ score set, in this case the high coverage supertree (5, above). Finally, we explore an alternative method of increasing the phylogenetic coverage of evolutionary-led conservation priorities using taxonomic imputation methods to estimate the distinctiveness scores of unsampled taxa. We show that setting conservation priorities is possible for groups with only limited sampling and that imputation offers a pragmatic method to estimate scores. Our approach has the potential to open up evolutionary-led conservation methods to the vast majority of species groups for which there is limited genetic sampling.

Materials and methods

We first surveyed the literature to determine how phylogenetic knowledge on scleractinian corals has developed since 2008 and how this expanding knowledge would have altered the EDGE-based conservation priorities within this group over time. We isolated five points in time where coral phylogenetic knowledge had notably increased and at each stage calculated ED and EDGE scores using these data. The points in time were as follows:

2008 – 127 species four-gene phylogeny

Fukami et al. (2008) analysed 127 species across 75 genera and 17 families using two mitochondrial DNA genes and two protein-coding nuclear genes in both maximum likelihood and Bayesian analyses, returning relatively strong support and high resolution for many major groups. To conform to the previous iterations of the EDGE process, we dated the root of Scleractinia at 425 million years ago (Stolarski et al., 2011) and used a strict molecular clock (MrBayes 3.2.1; Huelsenbeck & Ronquist, 2001; Ronquist et al., 2011; Selig et al., 2014). One species-based method considers the evolutionary relationships within a clade of target species (Vane-Wright, Humphries & Williams, 1991; Faith, 1992, 1996, 2007) alongside threat status to set conservation priorities (Witting & Loeschcke, 1995; Redding & Mooers, 2006; Isaac et al., 2007, 2012; Safi et al., 2013). The goal of prioritizing evolutionarily distinct species is to maximize the number of evolutionary trajectories conserved (Vane-Wright et al., 1991; Crozier, 1992). This precautionary, maximization approach is necessary as future evolutionary radiations are difficult to predict (Cooper & Fortey, 1998), which makes picking evolutionary ‘winners’ extremely difficult. Conservation actions based on measures of species-specific evolutionary isolation have been shown to conserve a greater proportion of the overall phylogenetic diversity of the target clade (Redding et al., 2008) and can preserve some unusual life-history and morphological traits, potentially helping to protect other aspects of biodiversity within a clade (Cadotte, Cardinale & Oakley, 2008; Cadotte et al., 2009; Redding, Dewolff & Mooers, 2010). A conservation programme that has applied this approach is the Evolutionarily Distinct and Globally Endangered (EDGE) of Existence programme (Isaac et al., 2007; http://www.edgeofexistence.org). The EDGE ranking system is a composite score based on two components: a measure of evolutionary distinctiveness (ED) and a measure of global endangerment (GE) or threat (Redding & Mooers, 2006; Isaac et al., 2007; Mooers, Faith & Maddison, 2008).

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& Huelsenbeck, 2003) to create a time-based ultrametric tree (10 million generations, burn-in = 25%, fixed topology).

2010a – EDGE Coral Expert Workshop

In the absence of a complete phylogeny, the EDGE programme held a participatory workshop (EDGE Coral Reefs Workshop, 4–5 March 2010, Zoological Society of London) to identify 10 priority EDGE species with which to initiate the EDGE Corals conservation programme. The workshop was attended by 22 scientists with a further 10 consulted remotely.

Prior to the workshop, 51 species were shortlisted for consideration through remote consultation with a range of coral experts (Supporting Information Appendix S1) and from ED scores derived from the Fukami et al. (2008) phylogeny. This phylogeny, although acknowledged to be incomplete, was used as the key reference during the 2010 EDGE Coral Expert Workshop as being the best available information at the time. A further five species were added during the initial workshop review process, thus totalling 56 species (Supporting Information Appendix S1). Participants were separated into two isolated working groups and asked to quantitatively score each species’ ED (arbitrary scoring of 1 to 4, with 4 being the highest ED) based on phylogenetic relationships presented in the Fukami et al. (2008) phylogeny and expert opinion against set criteria (Supporting Information Appendix S2). Note that no new phylogenies were reconstructed based on primary data for these 56 species.

2010b – 234 species single-gene barcode phylogeny

Published in late 2010 (Kitahara et al., 2010), this phylogeny had greater taxonomic coverage, but only used a single gene, one of the four utilized by Fukami et al. (2008): cytochrome c oxidase subunit I (COI), which we note is known to cause potential problems for accurate evolutionary inference (Castresana, 2007). For anthozoans in particular, the rate of COI sequence evolution is extremely low (Shearer et al., 2002; Hellberg, 2006; Huang et al., 2008), reducing the topological resolution of this tree over that of Fukami et al. (2008). Of the 21 major clades recognized in the latter, Kitahara et al. (2010) recovered 17, although most of the deeper relationships were concordant between them. One of the major clades was displaced by the conflicting placement of Oulastrea crispata (clade XI), while the remaining three were unresolved (clades XVII, XIX and XX) owing to low sequence variability of COI (Huang et al., 2009).

Here, we matched species names to those noted within the supertree and in agreement with the taxonomy used by the IUCN (2013) in the red-listing process, resulting in 234 species for which ED scores could be calculated. Again, we used a strict ultrametric clock (MrBayes 3.2.1) to create a time-based ultrametric tree (20 million generations, burn-in = 25%, fixed topology) with the root set at 425 million years.

2013 – 466 species multi-gene phylogeny

In 2013, a third phylogeny was inferred using an incomplete molecular supermatrix analysis of seven mtDNA markers (Huang & Roy, 2013). This study added three more markers over those used in Fukami et al. (2008) and nearly doubled the taxon sampling of Kitahara et al. (2010) because of the utilization of genes other than COI. All 21 major clades defined by Fukami et al. (2008) were recovered in this phylogeny, which also resolved the deepest three-clade topology (‘basal’, complex and robust groups; Stolarski et al., 2011) that were not found in the earlier studies.

We note that this product was not optimized by all authors for purposes here but is included as one possible method open to other groups where there is patchy sequencing but no published phylogeny. As for previous iterations, species names from this phylogeny were matched to those within the supertree resulting in 466 Scleractinia species that were in agreement with the IUCN taxonomy (IUCN, 2013).

2012 – 838 species supertree – reference tree

The reference tree used here was based on phylogenetic reconstructions published in Huang (2012) and Huang & Roy (2013). The topology of the tree was inferred using a supertree method (Baum, 1992; Ragan, 1992) that incorporated data from 975 reef and non-reef building coral terminals. Reef species constituted 838 of these taxa, including the 827 species evaluated for the IUCN Red List of Threatened Species (Carpenter et al., 2008). The source trees comprised a molecular tree of seven mtDNA genes from 466 species based on the above multi-gene phylogeny (Huang & Roy, 2013), 13 morphological trees and a taxonomic tree. These were coded into a bootstrap-weighted matrix representation and analysed under the maximum parsimony framework to infer a strict consensus supertree. Time-calibrated trees (chronograms) were then generated based on a mtDNA tree calibrated as per Simpson et al. (2011) and Stolarski et al. (2011). Using BEAST 1.6.2 (Drummond & Rambaut, 2007), five Markov chain Monte Carlo (MCMC) analyses of 30 million generations were carried out with a sampling interval of 1000 and a post-convergence burn-in of 10 million generations. Effective sample size values for all estimated parameters exceeded 200 when the five MCMC runs were combined, although nearly all exceeded 2000. Polytomies on the strict consensus supertree were randomly resolved using PolytomyResolver (Kuhn, Mooers & Thomas, 2011), constraining each node with a normal prior based on the 95% highest posterior density derived from the ultrametricized mtDNA tree (see Collen et al., 2011). Five MCMC runs of 3 million generations were performed and, following rejection of the first 1 million
steps, were subsampled every 10000 iterations, resulting in 1000 fully-resolved chronograms.

This dataset of 1000 chronograms is the most recent large-scale analysis of the phylogenetic relationships of corals. It contains the largest amount of information in terms of number of sources and relative quantity and also captures some measure of the variability of evolutionary relationships suggested by the data. As a result, we set this as our reference dataset. However, we acknowledge that these evolutionary trees are unlikely to represent the ‘true’ tree as there is still much work to be done to sample and analyse phylogenetic data from scleractinian corals before a phylogenetic consensus is reached (Wiens & Tiu, 2012).

**Null expectation**

We created a null expectation to which we compared the older incomplete empirically derived phylogenies by examining how ED-based information on the reference phylogeny degraded as tips were pruned from the complete trees. To create the dataset to run this analysis, we randomly chose a full tree (from the published distribution of 1000) and randomly subsampled this tree to create subtrees that contained 15 to 815 tips incrementally in steps of 50 tips (as a compromise between computational cost and resolution) to give a total of 16 subtrees of increasing size. This process was then repeated 10000 times, again this number was chosen as a balance between computational cost and examining potential variation between full trees in the distribution and the effect of randomly removing taxa on subtrees. We note this dataset represents a conservative null expectation as we would expect actual subsamples of a phylogeny to capture phylogenetic information and perform better than trees created using different data sources.

**Analysis**

For each subtree (i.e. trees in our dataset with fewer than 838 species) taken from either the empirical data or through subsampling the ‘reference’ tree, we calculated the ED metric (Isaac et al., 2007) for all species (Fig. 1). This score divides up all the branches in a target phylogeny among the tips, such that species with the fewest relatives have the largest summed branch length. We then calculated these scores on all versions of the full reference tree to provide a ‘reference’ ED score set (using ed.calc, caper package in R; Orme et al., 2012). We then calculated the correlation between these subtree scores and the reference set using only complete pairwise observations (Pearson’s rho; cor function in R package base; R Core Team, 2013).

Then, for all subtrees, we created a full set of 838 ED scores. Scores for missing species, according to taxonomy by Veron (2000), were estimated through statistical imputation (Little & Rubin, 1987; Holland et al., 2006; Swenson, 2013) using scores from related species represented in the subtrees. Specifically, missing representatives of each genus had the
mean genus score attributed to all non-sampled members. For a missing species that had no con-generic species, the mean family member score was used instead. If there were no family members represented in the subset tree, the mean ED score of all corals species in the subtree was attributed to the missing species. The development of new molecular techniques has led to new insights on coral evolution and more accurate phylogenetic inferences (e.g. Benzioni et al., 2007; Nunes et al., 2008; Huang et al., 2009; Budd et al., 2010; Gittenberger, Reijen & Hoeksema, 2011; Arrigoni et al., 2012). In fact, our reference tree was built based on data from these recent studies. Therefore, the imputation of scores for missing species based on Veron’s (2000) taxonomic classification should be considered a coarse approach and we can ask how our results align given the uncertain taxonomy.

For each subtree, we also created an EDGE species list. GE was measured using the IUCN Red List categories and criteria (IUCN, 2013), widely accepted as the best quantitative and objective framework available for evaluating extinction risk (e.g. Rodrigues et al., 2006; Hoffmann et al., 2008; Mace et al., 2008). Each category was re-coded to give a numeric value: critically endangered = 4, endangered = 3, vulnerable = 2, near threatened = 1, with lower rankings not scored. Threat data for 827 reef building corals were derived from the work of Carpenter et al. (2008). For each of the five empirical datasets, we created a list of the top 50 highest EDGE scoring species and then compared membership of the first four iterations with the top 50 EDGE species specified in our reference phylogeny.

Results

For both pairwise scores and imputed full scores, the ability for subsampled phylogenetic data to predict ED scores from the complete dataset increased in an approximately logarithmic fashion as completeness increased (Fig. 2). For example, half-sized trees (with 419 tips) produced pairwise correlation scores of 0.773 ± 0.186 and imputed full score sets that were 0.623 ± 0.368 (Pearson’s rho, n = 1000) correlated to the ‘reference’ set.

The ED scores from the largest, 2013, most recent tree were the most closely correlated (average 0.9, ± 0.01, n = 366, Spearman’s rho; 100 replications) to the reference scores, followed by the 234 species on the 2010b phylogeny (0.7, ± 0.01, n = 234, Spearman’s rho; 100 replications) (Fig. 3a). In all cases, compared with randomly selected subtrees of the same size from the reference tree, the correlations were the same as expected (Fig. 3a).

When imputing ED scores, accuracy in terms of correlation to the reference scores did not reduce as expected. Surprisingly, the 56 species for which there were expert-led scores (2010a) saw a large increase in correlation with the reference scores when they were used for estimating scores for all 838 species (from 0.63 to 0.79, n = 838, Spearman’s rho; 100 replications), although the most accurate was the 2013 phylogeny (0.85, ± 0.01, n = 838, Spearman’s rho; 100 replications) (Fig. 3b).

The number of ‘reference’ top 50 species were identified most efficiently using the workshop-derived ED scores which captured an average of 29 (± 2.24; 100 replications) of the ‘reference’ top 50, compared with imputed scores from random 53-tip subtrees which identified ~10 species successfully. With the exception of the 2013 tree, the other methods captured more of the reference top 50 species than subsamples of the full tree (Fig. 4).

Discussion

As expected, using the more complete and up-to-date phylogenies resulted in better prediction of reference scores. Surprisingly, the older phylogenies and scores derived from expert opinion also performed well. The expert opinion-led scores, when used for imputing score for missing species, was a close second in terms of prediction accuracy when compared with the most recent and largest tree, which had nearly 10 times the coverage. The key reason for this pattern is most likely a concerted attempt by the experts to consider species from every section of the tree rather than relying on those species that had already been sampled genetically (Rokas & Carroll, 2005).

The analyses also show that our approach of imputing data by simple taxonomy-based extrapolation of ED using
incomplete phylogenetic knowledge is worthwhile. Interestingly, the score accuracy remains unchanged during this process or, in the case of expert-led scores, appears to increase. The reference tree incorporates much of the recent phylogenetic data that have emerged, yet the use of a coarse classification that is in conflict with these new findings, that is Veron (2000), produces ED scores that are very similar to the reference results. Given the ability of small trees to be extrapolated to full taxonomic coverage with limited cost, we recommend more species groups to be assessed for their EDGE priorities (Isaac et al., 2007) to prompt conservation action in the absence of data. We should certainly strive for reduced phylogenetic uncertainty and even fully resolved trees based on comprehensive DNA sequence data in order to carry out these analyses, but our results suggest that a fairly accurate picture of evolutionary distinctiveness can be obtained with smaller datasets. Furthermore, it may be possible to achieve even greater accuracy in the future and we recommend research that utilizes more sophisticated methods for imputing missing data, such as a Bayesian phylogenetic approach (Hadfield & Nakagawa, 2010; Wiens & Tiu, 2012).

Our results suggest that small well-resolved trees can also accurately estimate ED scores when computing scores for just those species sampled. Although it was the least complete tree, the 2008-derived tree appears to benefit from being a well-resolved multi-gene tree with very high support values. However, our reference dataset will need to be revised as further iterations of the phylogeny are developed and a better consensus is established.

The top 50 EDGE scleractinian species lists showed significant variation from one another and from the reference list. While the GE component of the EDGE score remained consistent across the five analyses, the ED scores were sensitive to variation in the analytical methods. For example, in constructing the supertree, those taxa for which only topological information was available were initially assigned a branch length of zero at the terminal tip (Huang, 2012). During the process of creating an ultrametric time tree, these tips in particular are greatly extended by the time-calibration process so that they coincide with the present. Similarly, other processes may have caused discrepancies in the creation of our reference ED dataset. For example, Acropora echinata, Acropora speciosa and Favites abdita are each nested within clades containing numerous species. From a purely topological point of view, they are relatively indistinct and when the time-calibration process is applied, their terminal edges are
not scaled upward as much as other species, lowering their ranks.

**Value of expert opinion**

The workshop approach enables experts to contribute valuable, unpublished or field observations to complement the existing phylogenetic literature. This wealth of knowledge would be difficult for a single researcher to amass, but led to an improved accuracy of calculating EDGE scores by approximately 7% here, thus serving to validate the use of this method in setting conservation-priority species lists. However, future assessors should consider whether this improved accuracy is cost and time effective. For example, the EDGE Coral Expert Workshop was devised to establish the 'top 10' species to which conservation efforts should be directed. The two working groups within the expert workshop initially generated a list of 12 potential EDGE species, with eight being independently and consistently selected by both groups, showing a high degree of consensus. Such an approach could be easily achieved using alternative methods of opinion gathering, that is online polls such as SurveyMonkey (SurveyMonkey Inc., Palo Alto, CA, USA; Wright, 2005).

The workshop was not limited to purely setting priorities. An additional benefit of an expert-led workshop approach is that species can be assessed against secondary criteria (i.e. functional redundancy, dependant species or level of conservation attention) thus, making the resulting species list more practical. Such criteria (Supporting Information Appendix S3) were adopted at the EDGE Coral Workshop to reduce the 12 species put forward by the independent groups to the finalized top 10 priority EDGE coral species (Supporting Information Appendix S4). However, the strong differences between the top 10 species lists generated by empirical datasets suggest that in early data-poor stages, such workshops should be directed towards identifying a larger number of priority species. In data-poor situations, focusing efforts on the top 10 species appears to risk a very high turnover of high-priority species with subsequent phylogenetic inferences. This has been the case with the EDGE Mammals List. However, there is always a trade-off between conservation urgency and developing knowledge.

With the 2008 tree, as with many smaller incomplete phylogenies, species-poor groups (i.e. those likely to have high ED scores) are often not represented by genetic data (e.g. *Ctenactis* or *Lithophyllum*). It may be that these species are geographically remote, scarce and/or have unusual ecologies (e.g. *Anomastrea irregularis* and *Horastrea indica*). The EDGE Coral Expert Workshop was effective at identifying the top 50 EDGE species because all species lineages were considered for listing, whether sequence data were available or not. Indeed, the estimated ED scores produced by the EDGE Coral Expert Workshop appeared to add value to the scores from the incomplete 2008 phylogeny, allowing for more accurate identification of the top 50 species.

In addition to identifying the top 10 EDGE coral species, the workshop also added value to setting conservation priorities by identifying a list of ED species that are classified as data deficient (ED-DD species; Supporting Information Appendix S5) or not currently in a threatened category (Supporting Information Appendix S6) according to the IUCN Red List. These species should be monitored as they could become EDGE priority species in the future should their risk of extinction ever be assessed as threatened. Research into the extinction risk of these species is therefore identified as a high priority. Of these ED and ED-DD species (excluding those identified as the top 10 EDGE species), 77% were in the top 100 ED species of the reference list.

**Need for conservation action**

Increasingly, it is being recognized by the scientific community that ‘data-less and data-poor management are, under current circumstances, not just valid alternatives, but imperatives’ (Johannes, 1998). This was one of the original motivations behind the EDGE Coral Programme and in spite of a lack of phylogenetic data for some groups, active conservation projects are now underway for seven of the top EDGE corals (*Catalaphyllia jardinei, Ctenella chagius, Dendrogyra cylindrus, Heliofungia actiniformis, Horastrea indica, Parasimiplastrea sheppardi* and *Physogyra lichensteini*) through the EDGE of Existence Programme. While research is ongoing to enhance our phylogenetic knowledge of scleractinian corals and improve upon methodologies for phylogenetic-based conservation planning, the imminent and multiple threats to coral reefs necessitate urgent and decisive action. Therefore, despite the present incompleteness and uncertainty in phylogenetic inferences, an approach that uses the best science available to identify ED coral species is of great benefit to ongoing and future conservation initiatives. Here, we show that imputing data for missing species represents a promising development to allow phylogeny-led conservation programmes to assess and conserve those species groups we know little about, including the vast diversity of marine invertebrates that are even more data poor than corals. The dearth of EDGE assessments for invertebrates is indicative of our poor understanding of these animals (see Cardoso et al., 2011). Fortunately, the approach we demonstrate here, including our handling of data-deficient species and tracking of ED scores with increasing phylogenetic information, can be applied more broadly across the metazoan tree of life.

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References


**Supporting information**

Additional Supporting Information may be found in the online version of this article at the publisher’s web-site:

**Appendix S1.** Species shortlisted for consideration for 2010 EDGE Coral Workshop.

**Appendix S2.** ED scoring criteria for 2010 EDGE Coral Workshop.

**Appendix S3.** Secondary criteria for 2010 EDGE Coral Workshop.

**Appendix S4.** 2010 EDGE Coral Workshop Top 10 species.

**Appendix S5.** Summary of species identified as ED-DD species at the 2010 EDGE Coral Workshop.

**Appendix S6.** Summary of species identified as ED species at the 2010 EDGE Coral Workshop but not included in the EDGE list with IUCN priority listings.