



## Bedrock modulates the elevational patterns of soil microbial communities

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# 1 Bedrock modulates the elevational patterns of soil microbial 2 communities

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26

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34

## 35 **Abstract**

36 Elevational gradients are often used to reveal how soil microorganisms will respond to climate  
37 change. However, inconsistent microbial distribution patterns across different elevational  
38 transects have raised doubts about their practical applicability. We hypothesized that variations  
39 in bedrock, which influence soil physical and chemical properties, would explain these  
40 inconsistencies. We therefore investigated soil microbial communities (bacterial and fungal)  
41 along two adjacent elevational transects with different bedrocks (granite vs. slate) in a  
42 subtropical forest. [Our findings reveal that soil microbial communities are shaped by complex  
43 interactions between bedrock type and environmental factors along elevational gradients.  
44 Bacterial biomass was higher on slate, whereas fungal biomass was higher on granite. On  
45 granite, both bacterial and fungal biomass increased with elevation, whereas divergent patterns  
46 were observed on slate, likely due to the distinct soil properties or combinations of properties  
47 influencing microbial biomass on each bedrock. Bedrock and elevation strongly influenced  
48 microbial beta-diversity, with beta-diversity on granite driven primarily by soil total phosphorus  
49 and moisture, and on slate by soil organic carbon and pH. In contrast, alpha-diversity was  
50 impacted less by bedrock and elevation, but its relationship with environmental factors varied  
51 markedly between bedrock types. Overall, our results highlight the critical influence of bedrock  
52 in determining soil microbial community structure along elevational gradients and their potential  
53 responses to climate change.](#)

54

55 **Keywords:** altitude; climate; granite; parent material; slate; soil microbes.

56

## 57 **1. Introduction**

58 As a proxy for the impacts of climate change on microbial communities, elevational  
59 gradients provide unique insights into the regulatory mechanisms governing the spatial  
60 distribution of soil microorganisms (Sundqvist et al., 2013). Studying the distributions of soil  
61 microorganisms along elevational gradients not only reveals the mechanisms structuring soil

62 microbial communities (Nottingham et al., 2018; Peters et al., 2019), it also useful for  
63 understanding the impacts of climate change on soil biogeochemical cycles (Bahram et al.,  
64 2018; Hartmann and Six, 2022; Philippot et al., 2023). Numerous studies of the elevational  
65 patterns of soil microbial communities have emerged over the past two decades (Bryant et al.,  
66 2008; Fierer et al., 2011; He et al., 2020; Hendershot et al., 2017). However, these studies have  
67 not found consistent trends in microbial biomass or community  $\alpha$ -diversity: linear increases,  
68 linear decreases, unimodal and concave trends have all been detected (He et al., 2020;  
69 Hendershot et al., 2017; Wang et al., 2024). Environmental explanations for these disparate  
70 patterns include climatic regions (He et al., 2020), vegetation types (Li et al., 2016), or  
71 microclimate variation (Ma et al., 2022). However, the type of soil parent material -known as  
72 “bedrock”- may also explain complex elevational patterns.

73         Spatial variations in soil microbial communities are influenced strongly by soil properties  
74 (Fierer et al., 2009; Ni et al., 2022; Seaton et al., 2020). The factors shaping these communities  
75 are often determined by specific combinations of local soil characteristics. For example, soil pH  
76 is a well-established driver of bacterial community composition, particularly in acidic soils  
77 (Griffiths et al., 2011; Tripathi et al., 2018). Similarly, the availability of nutrients such as nitrogen  
78 and phosphorus plays a key role in regulating microbial dynamics, especially in nutrient-limited  
79 ecosystems (Delgado-Baquerizo et al., 2017). While climatic factors along elevational gradients  
80 tend to follow predictable trends within similar climate zones, the responses of soil microbial  
81 communities to climate change may vary depending on underlying soil conditions (e.g., acidic  
82 vs. neutral soils; nutrient-limited vs. nutrient-rich environments). Bedrock, as the parent material  
83 for soil formation, influences a wide range of soil physico-chemical properties, including pH,  
84 texture, and P levels (Augusto et al., 2017; He et al., 2021; Porder and Ramachandran, 2013;  
85 Spinola et al., 2022; Zeng et al., 2023). Consequently, soil microbial communities on different  
86 bedrock types may exhibit distinct responses to climate changes along elevational gradients,  
87 reflecting the unique soil environments created by bedrock characteristics. In the present study,  
88 we explored the extent to which bedrock can explain complex elevational patterns.

89         Bedrock varies in mountainous regions at both regional and local scales (Antonelli et al.,  
90 2018), affecting the spatial patterns of soil microorganisms (He et al., 2024; Hu et al., 2020; Li et  
91 al., 2018). However, only two studies have revealed that differences in bedrocks can affect the  
92 response of soil microbes to elevational gradients (Bhople et al., 2019; Singh et al., 2014).  
93 Singh et al. (2014) established two adjacent elevational transects on Mount Hana in South  
94 Korea, one on basalt and one on coarse-grained basalt. They observed a triple-curve in

95 bacterial species richness on the basalt and a concave pattern on the coarse-grained basalt.  
96 Bhole et al. (2019) showed a linear increase in soil microbial biomass on basaltic bedrock and  
97 acidic soils, and a unimodal pattern on limestone bedrock with pH neutral soils.

98         Studies focusing on the influence of bedrock on elevational patterns of soil microbes are  
99 remarkably scarce, and these studies often concentrate on a single feature of soil microbial  
100 communities, such as biomass, or community  $\alpha$ -diversity. Biomass,  $\alpha$ -diversity, and  $\beta$ -diversity  
101 are crucial characteristics of soil microbial communities, usually regulated by different factors.  
102 Soil microbes are often carbon (C) limited, which is why microbial biomass is predominantly  
103 driven by the availability of labile C (He et al., 2020).  $\alpha$ -diversity is more responsive to variations  
104 in soil pH (Fierer, 2017; Looby and Martin, 2020), particularly within acidic environments  
105 (Calderón-Sanou et al., 2022). The  $\beta$ -diversity of soil microbial communities, which describes  
106 the compositional variation among microbial communities across different environments, is  
107 influenced by a complex interplay of factors (Chen and Lewis, 2023). With such diverse  
108 characteristics of microbial communities being regulated by an assortment of environmental  
109 factors, it is no surprise that the interactions between bedrock, elevation, and soil microbial  
110 communities are extremely intricate.

111         In accordance with our hypothesis that bedrock modulates the environmental factors  
112 regulating soil microbial communities at different elevations, we anticipate significant differences  
113 in microbial community composition between the two bedrock types. Specifically, we expect to  
114 find that: (1) Soil microbial biomass,  $\alpha$ -diversity, and  $\beta$ -diversity differ between bedrock types.  
115 We expect microbial biomass and  $\alpha$ -diversity to increase on the slate transect due to its higher  
116 SOC, phosphorus levels, and pH (He et al., 2021). (2) Bedrock type will govern the relationship  
117 between elevation, microbial biomass and  $\alpha$ -diversity. Given that key environmental conditions  
118 structuring microbial communities differ on different bedrocks, we also predict that (3) factors  
119 driving  $\beta$ -diversity along the elevational gradient will vary between the two bedrocks, and similar  
120 environmental conditions may shape microbial community patterns differently on different  
121 bedrock types.

122

## 123 **2. Materials and methods**

### 124 **2.1 Study sites**

125 We worked in the Chebaling National Nature Reserve in the Guangdong Province of  
126 southern China (114°09'–114°16'E, 24°40'–24°46'N). The climate is a typical subtropical  
127 monsoon (He et al., 2021). The geological structure of the Reserve belongs to the South China  
128 fold system. Elevation ranges from 330 meters above sea level to 1,256 m.a.s.l. Cambrian and  
129 Ordovician strata are present in the northwest section. Northeast-southwest slate was formed  
130 after fold-fracture. The middle and south are Cambrian strata, forming slate mountains. The  
131 northern parts experienced intrusion of Jurassic plutonic rocks, forming acid plutonic rock  
132 mountains. Soils are classified in the Ultisol order and the Udult suborder based on the USDA  
133 soil classification system (Zhou et al., 2013).

134 We identified two adjacent mountains with different bedrocks (granite and slate) in the  
135 Chebaling National Nature Reserve (He et al. 2021). The geographic distance between the two  
136 mountains does not exceed 10 km. The vegetation on both mountains is well-preserved  
137 subtropical evergreen broad-leaved forest. The forest on the granite bedrock is dominated by  
138 *Schima superba*, *Machilus chinensis*, and *Eurya nitida*, while the forest on the slate bedrock is  
139 dominated by *Machilus chinensis*, *Eurya nitida* and *Rhododendron simsii*. A total of 18 sites  
140 were established along two elevational transects (Fig. S1), with nine sites on each bedrock.  
141 Plots were distributed at about 100-m intervals in elevation (determined by GPS) within each  
142 transect, with elevations ranging from 410 to 1,080 m.a.s.l. on the granite bedrock and 350 to  
143 1,120 m.a.s.l. on the slate bedrock. To reduce the influence of aspect, sampling plots were  
144 located on the south side of any microtopography at each site.

## 145 **2.2. Sampling and analytical methods**

146 All plots (40 m x 40 m) were sampled in October 2018. All trees with a diameter at breast  
147 height above 1 cm were recorded in each plot. We estimated the forest above-ground biomass  
148 (AGB) using diameter at breast height of each tree and allometric relationships (Réjou-Méchain  
149 et al., 2017). We installed a Micro Station Data Logger (USA, HOBO, H21-002) in each plot,  
150 with two probes inserted into the soil (at a depth of approximately 10 cm) which monitored soil  
151 temperature and moisture. Recordings were taken hourly from July 13, 2018, to July 13, 2019.  
152 Here, we use the data collected over the entire year to calculate the soil mean annual  
153 temperature (MAT) and moisture, which we use to explain the spatial variation of soil microbial  
154 community characteristics.

155 Volumetric soil samples were taken to determine soil bulk density. Soil depth was more  
156 than 100 cm in all but two of the high elevation sites in the slate transect. In these two plots, soil

157 depth was roughly 60 cm. These shallow soil depths were likely due to severe erosion on the  
158 steeper slopes. Five subplots (10 × 10 m) were randomly selected at each site. We removed the  
159 leaf litter from the forest floor and collected topsoil to a depth of 20 cm using a stainless soil  
160 corer (inner diameter = 3.5 cm). We collected six random soil cores and homogenized them into  
161 composite samples for each subplot. A total of 90 soil samples (i.e. 18 plots x 5 subplots) were  
162 collected and transported on ice directly to the laboratory. Each soil sample was then passed  
163 through a 2-mm sieve before being divided into two subsamples: one was stored at -80°C for  
164 phospholipid fatty acid (PLFA) analysis and high-throughput sequencing (HTS), and one was  
165 air-dried at room temperature for the measurement of soil physicochemical properties in the  
166 laboratory.

167 We measured soil pH with a PHS-3C pH acidometer (soil-water ratio of 1:5) and used  
168 dry combustion with an elemental analyser (Perkin Elmer 2400 Series II) to measure soil  
169 organic carbon (SOC) and total N (TN) concentrations. Soil total P (TP) concentration was  
170 measured using a nitric acid–perchloric acid digestion, followed by a colorimetric analysis  
171 (Murphy and Riley, 1962) using a UV-Vis spectrophotometer (UV1800; Shimadzu, Kyoto,  
172 Japan). We measured particle size distribution using a laser particle analyzer based on the laser  
173 diffraction technique operating over a range of 0.02-2000 µm (Mastersizer 2000 particle size  
174 analyzer, Malvern Instruments, Ltd., UK).

175 We used a modified PLFA analysis (Frostegård and Bååth, 1996) to determine bacterial  
176 and fungal biomass. The abundance of individual fatty acids was expressed as µg per g of dry  
177 soil. Concentrations of each PLFA were calculated based on the 19:0 internal standard  
178 concentrations and microbial biomass was expressed as the sum of identifiable PLFAs. We  
179 chose a set of fatty acids to represent bacterial PLFAs. Bacterial PLFAs were obtained by  
180 summing the phospholipid fatty acid 14:00, 15:00, 16:00, 18:00, 13:0 anteiso, 13:0 iso, 14:0 iso,  
181 14:1 ω5c, 15:0 anteiso, 15:0 iso, 15:1 ω6c, 16:0 iso, 16:1 ω5c, 16:1 ω7c, 17:0 anteiso, 17:0  
182 cyclo ω7c, 17:0 iso, 18:1 ω7c, 18:1 ω9c, 19:0 cyclo ω7c, and 19:0 cyclo ω9c contents. [Gram-](#)  
183 [positive bacteria were identified by branched-chain fatty acids, including 13:0 anteiso, 13:0 iso,](#)  
184 [14:0 iso, 15:0 anteiso, 15:0 iso, 16:0 iso, and 17:0 anteiso and iso. Gram-negative bacteria were](#)  
185 [distinguished by monounsaturated and cyclopropyl fatty acids, specifically 14:1 ω5c, 15:1 ω6c,](#)  
186 [16:1 ω5c, 16:1 ω7c, 17:0 cyclo ω7c, 18:1 ω7c, 18:1 ω9c, 19:0 cyclo ω7c, and 19:0 cyclo ω9c.](#)  
187 The sum of 18:2ω6c and 18:3 ω6c represented fungal PLFAs.

188 Soil DNA was extracted from composite soil samples using the FastDNA SPIN Kit for  
189 Soil (MP Biomedicals, Heidelberg, Germany) and purified by agarose gel electrophoresis. The  
190 quality of the DNA samples was checked on a spectrophotometer (NanoDrop, ND2000,  
191 ThermoScientific, USA). Total DNA was used for high-throughput sequencing on an Illumina  
192 MiSeq platform (San Diego, CA, USA). The bacterial V4 hypervariable region of the 16S rRNA  
193 gene and fungal internal transcribed spacer (ITS) region was amplified using the primer pair  
194 505F/816R (5'-GTGCCAGCMGCCGCGG-3'/5'-GGACTACHVGGGTWTCTA AT-3') (Caporaso  
195 et al., 2011) and ITS1F/ITS2 (5'-GGAAGTAAAAGTCGTAACAAGG-3'/5'-  
196 GCTGCGTTCTTCATCGATGC-3') (Shen et al., 2020) along with the Illumina adaptor sequence  
197 and barcode sequences, respectively.

198 The raw sequence data were processed and analyzed using QIIME Pipeline (Caporaso  
199 et al., 2011). To improve sequence quality we removed average quality (value  $\leq 20$ ) sequencing  
200 reads with ambiguous nucleotides in barcodes, and homopolymer reads between 8 bp and 150  
201 bp in length. Paired ends were joined with FLASH (Magoc and Salzberg, 2011). Chimeric  
202 sequences were detected and eliminated using the Uchime algorithm (Edgar, 2013). All  
203 sequences were clustered into operational taxonomic units (OTUs) at a 97% identity threshold.  
204 Finally, the representative sequences of each OTU were classified against the RDP 16S rRNA  
205 database for bacteria and UNITE Fungal ITS database for fungi with an 80% confidence  
206 threshold. The resultant OTU abundance tables from these analyses were rarefied to an even  
207 number of sequences per sample to ensure equal sampling depth (26,160 and 26,760 for 16S  
208 rDNA and ITS, respectively). [To minimize the influence of potentially spurious OTUs, we](#)  
209 [excluded those with a total read count below 50 or present in fewer than five samples after](#)  
210 [rarefaction](#). All subsequent analyses of  $\alpha$ - and  $\beta$ -diversity were conducted based on this filtered  
211 OTU table. The raw reads have been deposited into the National Centre for Biotechnology  
212 Information (NCBI) Sequence Read Archive database (PRJNA1177672).

### 213 **2.3 Statistical analyses**

214 We used Wilcoxon tests to assess differences in microclimate, plant traits, and soil  
215 properties between granite and slate bedrocks. To evaluate elevational trends, we applied  
216 univariate linear regression models, while multivariate linear regression models were used to  
217 examine soil microbial community responses across different elevations and bedrock types, as  
218 well as to identify interactive effects. Model fit was evaluated using Akaike's Information  
219 Criterion (AIC), with the model having the lowest AIC score selected as the best fit.

220 Spearman correlation analyses were conducted to determine whether bedrock type  
221 influenced relationships between environmental variables and soil microbial communities.  
222 Additionally, we applied multiple regression models to investigate associations between  
223 microbial variables (bacterial and fungal biomass, biomass ratios, and alpha diversity indices)  
224 and a range of environmental predictors, including soil properties, i.e., pH, moisture, clay  
225 content, soil organic C (SOC), soil P, soil C-to-N ratio (C:N), soil C-to-P ratio (C:P), soil N-to-P  
226 ratio (N:P), plant traits (above-ground biomass and plant Shannon diversity), and climatic  
227 factors (mean annual temperature). Multicollinearity among predictors was assessed using  
228 Variance Inflation Factor (VIF) values calculated with the *vif* function from the *car* package.  
229 Initial VIF analysis revealed high collinearity among certain soil nutrient ratios (soil C:P and soil  
230 N:P), with VIF values exceeding 100; thus, these variables were excluded, reducing the VIF of  
231 all remaining predictors to below 5. To examine interactions between environmental predictors  
232 and bedrock type, we incorporated selected interaction terms (bedrock:TP, bedrock:moisture,  
233 bedrock:pH, and bedrock:MAT) aligned with our research questions. Due to the limited sample  
234 size, we focused on these specific interactions rather than including all possible terms. Stepwise  
235 model selection using AIC was performed with the *dredge* function from the *MuMIn* package to  
236 identify best-fit models for each microbial variable, allowing for retention of the most informative  
237 predictors while optimizing model performance.

238 We calculated the Chao1 index, Shannon, and Inverse Simpson diversity index as  $\alpha$ -  
239 diversity indices of soil microbial communities. Shannon index is defined as  $H = -\sum_1^i P_i \log P_i$ ,  
240 where  $P_i$  is the proportional abundance of species  $i$ . Inverse Simpson index is defined as  $1/D$ ,  
241 where  $D = \sum P_i^2$ . We used the Bray-Curtis-dissimilarities-based principal components analysis  
242 (PCoA) to assess differences ( $\beta$ -diversity) in microbial communities in different sites and  
243 bedrocks. We performed square root transformations of the OTU relative abundances before  
244 the PCoA. We performed distance-based Redundancy Analysis (db-RDA) of the correlation  
245 between predictor variables and microbial composition. We calculated these diversity indices  
246 and conducted these ordination analyses using the *vegan* R package (Oksanen et al., 2020).  
247 We performed a Principal Component Analysis (PCA) to visualize the variation in environmental  
248 variables across elevational gradients on two bedrock types. The analysis was conducted using  
249 the *PCA* function from the *FactoMineR* package. A biplot was created using *fviz\_pca\_biplot*  
250 from the *factoextra* package. We used a neutral community model (NCM) (Sloan et al., 2006) to  
251 test whether deterministic or stochastic processes were structuring the microbial communities.

252 We used *Hmisc*, *minpack.lm* and *stats4* packages for the NCM, with default parameters for  
253 model fitting.

254 All statistical analyses were performed using R (R Core Team, 2023) and graphs were  
255 generated with the *ggplot2* package (Wickham, 2016).

256

## 257 **3. Results**

### 258 **3.1 Effects of elevation and bedrock on soil characteristics**

259 Wilcoxon tests showed no significant differences in soil MAT, moisture, and SOC  
260 concentration between the granite and slate transects (Table S1). AGB, soil C:N, C:P, N:P  
261 ratios, and soil silt and sand contents were higher on the granite transect, whereas plant  
262 diversity, soil pH, bulk density, TN, TP, and clay content were lower on the granite than on the  
263 slate transect. Univariate linear regression models revealed a consistent pattern of significant  
264 declines in MAT, plant diversity and soil pH with elevation, and an increase in SOC, TN, TP and  
265 silt content across both bedrock types (Fig. S2 and S3). Soil clay content and moisture showed  
266 no significant elevational trend along either transect (Fig. S3). AGB decreased with elevation on  
267 the granite but showed no significant trend on the slate (Fig. S2). [PCA results revealed a clear  
268 separation of sampling sites along PC1 \(Figure S4\). Key variables associated with PC1, such as  
269 soil P, C:N, C:P, N:P, clay, and moisture, appear to be major environmental drivers of microbial  
270 community differences between the two bedrocks. Variables closely aligned with PC2, including  
271 elevation, MAT, SOC, and pH, are likely primary drivers of microbial community changes along  
272 the elevational gradient within each transect.](#)

### 273 **3.2 Effects of elevation and bedrock on soil microbial biomass**

274 [Bacterial biomass was slightly higher on slate, whereas fungal biomass was higher on  
275 granite \(Table 1\). Consequently, the bacteria-to-fungi biomass ratio \(B/F\) was significantly  
276 higher on slate than on granite. Given that Gram-positive bacteria are ecologically and  
277 functionally more similar to fungi, our findings align with this pattern: the Gram-positive to Gram-  
278 negative bacteria ratio \(G+/G- ratio\) was significantly higher on granite than on slate. Notably,  
279 both bacterial \(B\) and fungal \(F\) biomass, along with the G+/G- ratio increased significantly with  
280 elevation on the granite but not on the slate transect; B/F ratios showed no significant linear  
281 trends on either bedrock \(Fig. 1\). Multivariate linear regression models confirmed significant](#)

282 impacts of elevation and bedrock on microbial biomass characteristics, including significant  
283 interactive effects for bacterial and fungal biomass but not for B/F ratios or G+/G- ratios (Table  
284 2).

285 The best models selected through stepwise regression provided strong explanatory  
286 power for variations in microbial biomass, with adjusted  $R^2$  values ranging from 0.45 to 0.71  
287 (Table 3). Soil microbial biomass was shaped significantly by multiple environmental factors and  
288 their interactions with bedrock type. For bacterial biomass, SOC, TP, and the soil C:N ratio were  
289 key factors, with positive associations observed for soil C and P, and a negative association  
290 with the C:N ratio. Fungal biomass was also influenced by SOC, C:N ratio, and TP, though the  
291 impacts of moisture and MAT varied depending on bedrock type. The B/F biomass ratio was  
292 shaped by soil moisture, plant Shannon diversity, and TP, with an interaction between moisture  
293 and bedrock type. Additionally, the G+/G- ratio was driven by soil clay content, MAT, and pH,  
294 with a significant interaction between TP and bedrock type.

295 Spearman correlations corroborated the multiple regression results (Tables S2 & S3).  
296 On the granite transect, bacterial biomass was strongly associated with SOC and TP, whereas  
297 on slate, no significant correlations were observed. Fungal biomass on granite correlated  
298 positively with SOC and the soil N:P ratio, while being negatively associated with AGB and  
299 MAT. On slate, fungal biomass correlated only with moisture. The B/F ratios on granite were  
300 positively linked with soil TP and negatively to moisture, whereas on slate, they correlated  
301 negatively with moisture. Additionally, the G+/G- ratios showed a strong positive correlation with  
302 the soil N:P ratio on granite and a negative correlation with MAT on slate.

### 303 **3.3 Effects of elevation and bedrock on soil microbial community diversity and** 304 **composition**

305 Microbial richness, as indicated by Chao1 indices, was consistently and significantly  
306 higher on slate for both bacterial and fungal communities (Table 1). In contrast, Shannon  
307 indices showed no significant differences between bedrock types for either bacterial or fungal  
308 communities, indicating comparable overall diversity. The inverse Simpson index, however, was  
309 significantly lower for bacterial communities on slate than on granite, suggesting reduced  
310 evenness and potential dominance by a few species on slate. Interestingly, soil microbial  $\alpha$ -  
311 diversity, encompassing Shannon and inverse Simpson indices, showed no clear elevational  
312 trends (Fig. 2). Multivariate linear regression analyses confirmed these observations, identifying  
313 significant differences between bedrock transects in the Chao1 and inverse Simpson indices for

314 bacteria, and in the Chao1 index for fungi, but not in other  $\alpha$ -diversity measures. Additionally,  
315 elevation significantly influenced only the fungal Chao1 index, with no observable effect on other  
316  $\alpha$ -diversity indices or significant interactions between elevation and  $\alpha$ -diversity (Table 2).

317 Regression models for diversity indices showed lower explanatory power compared with  
318 biomass (adjusted  $R^2 = 0.20-0.38$ ; Table 4). Bacterial Shannon index was significantly affected  
319 by clay content, MAT, moisture, and pH, while the bacterial inverse Simpson index was  
320 associated with soil C, P, and a moisture-bedrock interaction. Both the fungal Shannon and  
321 inverse Simpson indices were influenced by soil P, soil C, and pH, with strong effects from  
322 interactions between these variables and bedrock type; soil P, in particular, played a prominent  
323 role in shaping fungal community diversity.

324 Spearman correlation analysis indicated that the Shannon index of the bacterial  
325 community correlated positively with soil pH on both granite and slate transects (Table S2 and  
326 S3). The inverse Simpson index of bacteria correlated positively with soil C:P and N:P ratios on  
327 granite but showed no significant relationship on slate. The fungal community's Shannon index  
328 correlated negatively with MAT on both bedrocks, and additionally with clay content on granite.  
329 On slate, it correlated significantly with TP, C:P, and N:P. The inverse Simpson index for fungi  
330 correlated negatively with clay on granite, whereas on slate it showed significant correlations  
331 with soil pH, C:N, C:P, N:P, and other environmental factors (Table S3).

332 Principal Components Analysis results highlighted clear differences in beta-diversity,  
333 i.e., the composition of soil microbial communities (bacterial and fungal), across different  
334 bedrock types (Fig. 3 a & b). Results of db-RDA revealed that the compositions of soil bacterial  
335 and fungal communities were determined primarily by the soil C:P and N:P ratios and TP  
336 content (Fig. 3a). These findings indicate that variations in phosphorus level were key in driving  
337 the differences in community composition observed between the bedrocks. Further db-RDA on  
338 individual bedrock types revealed that on granite, soil TP content and moisture were crucial in  
339 shaping both bacterial and fungal communities (Fig. 3 c & e). Conversely, on slate, SOC and  
340 soil pH were the dominant factors influencing bacterial communities (Fig. 3d), whereas moisture  
341 and clay content significantly affected fungal communities (Fig. 3f). This analysis suggests a  
342 role for bedrock in mediating species turnover along elevational gradients, with changes in soil  
343 P concentration and moisture levels being pivotal.

344 The fit of bacterial communities on granite to the NCM was higher ( $R^2 = 0.92$ ;  $N_m =$   
345 18943) compared with slate ( $R^2 = 0.90$ ;  $N_m = 16458$ ) (Fig. S6 a & b), suggesting the importance

346 of stochastic processes in the assembly of these bacterial communities. In the case of the  
347 fungal communities, the fit to the NCM was comparable between granite ( $R^2 = 0.62$ ;  $N_m = 1237$ )  
348 and slate ( $R^2 = 0.62$ ;  $N_m = 1209$ ), indicating no discernible difference in community assembly  
349 processes (Fig. S6 c & d).

350

## 351 **4. Discussion**

352 This study reinforces findings by Bhole et al. (2019) and Singh et al. (2014), confirming  
353 that bedrock composition plays a critical role in shaping soil microbial community responses  
354 along elevational gradients in subtropical mountain ecosystems. While this and previous studies  
355 each focused on a single transect per bedrock type, together they provide robust evidence that  
356 bedrock type significantly influences microbial elevational patterns. Unique combinations and  
357 ranges of soil properties are established by different bedrock types, resulting in distinct microbial  
358 community distributions along elevation gradients. Moreover, the impact of specific  
359 environmental factors on microbial communities varies with bedrock type, highlighting the  
360 interactive effects between bedrock and environmental conditions on microbial distribution.  
361 These interactions deepen our understanding of the intricate dynamics of microbial communities  
362 and underscore the necessity to consider bedrock type when evaluating microbial responses to  
363 environmental changes.

### 364 **4.1 Bedrock modulates the elevational patterns of soil microbial biomass**

365 Bacterial and fungal biomass differed significantly on the contrasting bedrocks. Both  
366 transects were under similar climates, and we therefore conclude that differences in microbial  
367 biomass were likely caused by the variation of bedrock, which concurs with previous studies  
368 (Deng et al., 2015; Sun et al., 2016) (Deng et al., 2015; Sun et al., 2016). Deng et al. (2015),  
369 who worked in a similar subtropical monsoon climate, concluded that bedrock explained more  
370 variation in soil microbial biomass than land use, after discovering that microbial biomass in soil  
371 derived from granite was significantly higher than in soil derived from quaternary red earth and  
372 tertiary red sandstone. Sun et al. (2016) showed that agricultural soils derived from granite  
373 supported more microbial biomass than quaternary red clay soil and purple sandy shale, even  
374 after 40 years of agricultural use. These results further emphasize the fact that bedrock drives  
375 the spatial variation of soil microbial biomass.

376 Bacteria and fungi responded differently to the different bedrocks in our study, which is  
377 further evidence of the regulatory effects of bedrock on microbial communities. [Bacterial](#)  
378 [biomass, particularly the amount of Gram-positive versus Gram-negative bacteria, was higher](#)  
379 [on slate than on granite. This can be attributed to slate's higher soil TN and TP contents, closer-](#)  
380 [to-neutral soil pH, and higher soil clay content, all of which favor bacterial growth. Bacteria,](#)  
381 [especially Gram-positive types, rely heavily on nutrient availability](#) (Yu et al., 2022), [are more](#)  
382 [sensitive to pH changes](#) (Luan et al., 2023; Rousk et al., 2010), [and benefit from the simplified](#)  
383 [physical conditions of soils richer in clay](#) (Philippot et al., 2023). In contrast, fungi are better at  
384 extracting nutrients from decomposing organic matter (Koranda et al., 2014)q, and have a  
385 greater tolerance to pH changes (Rousk et al., 2010). Moreover, their multicellular, filamentous  
386 structure enables fungi to adapt to a variety of soil physical environments (Philippot et al., 2023).  
387 Furthermore, considering the competitive dynamics between fungi and bacteria (Bahram et al.,  
388 2018), the diminished bacterial biomass on granite reduces competition, potentially boosting  
389 fungal biomass. These mechanisms clarify why fungal biomass was higher on granite than on  
390 slate, providing insight into how bedrock variability distinctly influences bacterial and fungal  
391 communities.

392 As well as observing the effects of the bedrock itself on soil microbial communities, we  
393 also noticed inconsistencies in the elevational patterns of the soil microbial biomass along the  
394 two transects. On the granite, soil bacterial and fungal biomass increased with elevation,  
395 whereas on the slate, bacterial biomass showed no trend, and fungal biomass decreased  
396 slightly. Soil microbes usually need to derive energy and nutrients from soil organic matter, and  
397 as a result, their biomass is generally coupled with SOC concentration (He et al., 2020; Smith et  
398 al., 2021). However, in this study, such a relationship was observed only on granite; on slate,  
399 which had higher N and P levels, the biomass of bacteria and fungi did not show significant  
400 correlations with SOC, soil TP content, MAT, or other factors. This may be attributed to the fact  
401 that in nutrient-abundant environments, especially with sufficient P, soil microbes experience  
402 lower nutrient limitations; and their reliance on the pathway of nutrient acquisition through the  
403 decomposition of organic matter might be comparatively weaker (Lang et al., 2016). This would  
404 certainly explain why microbial biomass was not correlated with SOC and TP concentrations on  
405 slate. These findings suggest that the bedrock, by influencing the P levels in the soil and indeed  
406 throughout the entire ecosystem, can impact the responses of soil microbial biomass to  
407 elevational gradients.

#### 408 **4.2 Bedrock modulates the elevational patterns of soil microbial community diversity**

409           The higher Chao1 index for both bacterial and fungal communities on slate than on  
410 granite indicates greater microbial species richness in soils with higher nutrient content and pH  
411 (Xiao et al., 2022). However, neither the Shannon nor the inverse Simpson index on slate were  
412 significantly higher than on granite, suggesting that the increased richness on slate likely  
413 reflects a greater presence of rare or low-abundance taxa, reducing overall community  
414 evenness. In particular, the inverse Simpson index for bacterial communities was significantly  
415 higher on granite than on slate. This may be attributed to the lower P, moisture, and pH levels  
416 on granite, which may promote a broader range of microbial taxa that coexist more evenly.  
417 Conversely, the higher P, moisture, and pH levels on slate could favor a few dominant species,  
418 resulting in lower evenness despite the elevated species richness. Interestingly, unlike microbial  
419 biomass, neither bacterial nor fungal  $\alpha$ -diversity varied significantly with elevation on either  
420 bedrock type, suggesting that soil microbial biomass and community  $\alpha$ -diversity are regulated by  
421 different factors (Li et al., 2020; Ren et al., 2018). These distinct responses between microbial  
422 biomass and  $\alpha$ -diversity, with their implications for ecosystem functioning, warrant further  
423 exploration.

424           With regard to those factors influencing  $\alpha$ -diversity, our multiple linear models explained  
425 significantly less of the variation in  $\alpha$ -diversity than in microbial biomass along the elevational  
426 gradient. Alongside the high explanatory power of neutral community models (NCM) on both  
427 bedrocks—especially with over 90% for bacterial communities—our findings suggest that  
428 microbial community assembly along the elevation gradient was driven largely by stochastic  
429 processes, with environmental factors playing a lesser role. While we found significant  
430 relationships between soil pH and the Shannon indices for both bacteria and fungi, supporting  
431 the notion that microbial  $\alpha$ -diversity is sensitive to soil acidity (Luan et al., 2023; Smith et al.,  
432 2021),  $\alpha$ -diversity itself did not vary significantly with elevation. This is likely due to the relatively  
433 small pH fluctuations across the transects. Despite the limited explanatory power of our models,  
434 we observed that certain factors, such as soil phosphorus and moisture, had bedrock-specific  
435 effects on microbial  $\alpha$ -diversity. These significant interactions indicate that although  $\alpha$ -diversity  
436 did not shift noticeably with elevation, its relationship with environmental factors was still  
437 modulated by bedrock.

438           Our findings suggest that bedrock impacts the  $\beta$ -diversity of soil microbial communities.  
439 Bacteria and fungi displayed markedly distinct compositions across the two bedrocks. Our  
440 observation that bedrock influences the composition of soil microbial communities aligns with  
441 the conclusions of previous studies (Sheng et al., 2023; Tytgat et al., 2016; Weemstra et al.,

442 2020; Xiao et al., 2022). Studies of different bedrocks have proposed different mechanisms for  
443 structuring soil microbial communities. For example, Tytgat et al. (2016) found that SOC content  
444 structured bacterial communities, whereas Sheng et al. (2023) concluded that soil pH structured  
445 the bacterial community composition among different bedrocks. We identified differences in soil  
446 P as the primary mechanism structuring soil microbial communities on the granite and slate  
447 bedrocks. On granite, soil TP content and moisture govern species turnover of both bacteria  
448 and fungi, which is supported by another study in nearby subtropical forest (Chen and Lewis,  
449 2023). On slate, however, the influence of soil P on species turnover appeared to be minimal.  
450 This could be due to P not being a limiting factor, as slate and its associated soils have high P  
451 concentrations. Together, these results indicate that bedrock type not only influences the  
452 composition of soil microbial communities, but also modulates the primary drivers of microbial  
453 community structure along elevational gradients.

454 Our results should be interpreted in light of the fact that our study was based on one  
455 elevational transect per bedrock type. Nonetheless, given the inconsistency of previous studies  
456 of soil microbial elevational patterns, including those based on single transects (e.g.,  
457 Bayranvand et al., 2021; Peters et al., 2016; Zakavi et al., 2022), our findings provide relevant  
458 and valuable insights into how bedrock influences microbial community patterns along elevation  
459 gradients. Ideally, future studies should integrate multiple transects replicated within bedrock  
460 types to more thoroughly understand the responses of soil microbial communities to climate  
461 gradients.

462

## 463 **5. Conclusion**

464 We have shown that bedrock significantly influences soil microbial biomass and  $\beta$ -  
465 diversity, while having limited effects on  $\alpha$ -diversity. Moreover, bedrock modulated the impacts  
466 of the elevation gradient on soil microbial biomass and  $\beta$ -diversity. This was likely an indirect  
467 process via the alteration of soil P content, C:P, N:P ratios, soil moisture, and pH. We believe  
468 that bedrock may explain some of the inconsistencies surrounding previous studies of the  
469 elevational patterns of soil microbial communities. We also anticipate that bedrock will modulate  
470 the impacts of climate change on soil microbial communities.

471

472 **Reference:**

- 473 Antonelli, A., Kissling, W.D., Flantua, S.G.A., Bermúdez, M.A., Mulch, A., Muellner-Riehl, A.N.,  
 474 Kreft, H., Linder, H.P., Badgley, C., Fjeldså, J., Fritz, S.A., Rahbek, C., Herman, F.,  
 475 Hooghiemstra, H., Hoorn, C., 2018. Geological and climatic influences on mountain  
 476 biodiversity. *Nat. Geosci.* 11, 718–725. <https://doi.org/10.1038/s41561-018-0236-z>
- 477 Augusto, L., Achat, D.L., Jonard, M., Vidal, D., Ringeval, B., 2017. Soil parent material—A major  
 478 driver of plant nutrient limitations in terrestrial ecosystems. *Glob. Change Biol.* 23, 3808–  
 479 3824. <https://doi.org/10.1111/gcb.13691>
- 480 Bahram, M., Hildebrand, F., Forslund, S.K., Anderson, J.L., Soudzilovskaia, N.A., Bodegom,  
 481 P.M., Bengtsson-Palme, J., Anslan, S., Coelho, L.P., Harend, H., Huerta-Cepas, J.,  
 482 Medema, M.H., Maltz, M.R., Mundra, S., Olsson, P.A., Pent, M., Pölme, S., Sunagawa,  
 483 S., Ryberg, M., Tedersoo, L., Bork, P., 2018. Structure and function of the global topsoil  
 484 microbiome. *Nature* 560, 233–237. <https://doi.org/10.1038/s41586-018-0386-6>
- 485 Bayranvand, M., Akbarinia, M., Salehi Jouzani, G., Gharechahi, J., Kooch, Y., Baldrian, P.,  
 486 2021. Composition of soil bacterial and fungal communities in relation to vegetation  
 487 composition and soil characteristics along an altitudinal gradient. *FEMS Microbiol. Ecol.*  
 488 97, fiae201. <https://doi.org/10.1093/femsec/fiae201>
- 489 Bhople, P., Djukic, I., Keiblinger, K., Zehetner, F., Liu, D., Bierbaumer, M., Zechmeister-  
 490 Boltenstern, S., Joergensen, R.G., Murugan, R., 2019. Variations in soil and microbial  
 491 biomass C, N and fungal biomass ergosterol along elevation and depth gradients in  
 492 Alpine ecosystems. *Geoderma* 345, 93–103.  
 493 <https://doi.org/10.1016/j.geoderma.2019.03.022>
- 494 Bryant, J.A., Lamanna, C., Morlon, H., Kerkhoff, A.J., Enquist, B.J., Green, J.L., 2008. Microbes  
 495 on mountainsides: Contrasting elevational patterns of bacterial and plant diversity. *Proc.*  
 496 *Natl. Acad. Sci.* 105, 11505–11511. <https://doi.org/10.1073/pnas.0801920105>
- 497 Calderón-Sanou, I., Zinger, L., Hedde, M., Martinez-Almoyna, C., Saillard, A., Renaud, J.,  
 498 Gielly, L., Khedim, N., Lionnet, C., Ohlmann, M., Consortium, O., Münkemüller, T.,  
 499 Thuiller, W., 2022. Energy and physiological tolerance explain multi-trophic soil diversity  
 500 in temperate mountains. *Divers. Distrib.* 28, 2549–2564.  
 501 <https://doi.org/10.1111/ddi.13529>
- 502 Chen, J., Lewis, O.T., 2023. Limits to species distributions on tropical mountains shift from high  
 503 temperature to competition as elevation increases. *Ecol. Monogr.* e1597.  
 504 <https://doi.org/10.1002/ecm.1597>
- 505 Delgado-Baquerizo, M., Reich, P.B., Khachane, A.N., Campbell, C.D., Thomas, N., Freitag,  
 506 T.E., Abu Al-Soud, W., Sørensen, S., Bardgett, R.D., Singh, B.K., 2017. It is elemental:  
 507 soil nutrient stoichiometry drives bacterial diversity: C:N:P stoichiometry drives bacterial  
 508 diversity. *Environ. Microbiol.* 19, 1176–1188. <https://doi.org/10.1111/1462-2920.13642>
- 509 Deng, H., Yu, Y.-J., Sun, J.-E., Zhang, J.-B., Cai, Z.-C., Guo, G.-X., Zhong, W.-H., 2015. Parent  
 510 materials have stronger effects than land use types on microbial biomass, activity and  
 511 diversity in red soil in subtropical China. *Pedobiologia* 58, 73–79.  
 512 <https://doi.org/10.1016/j.pedobi.2015.02.001>
- 513 Fierer, N., 2017. Embracing the unknown: disentangling the complexities of the soil microbiome.  
 514 *Nat. Rev. Microbiol.* 15, 579–590. <https://doi.org/10.1038/nrmicro.2017.87>
- 515 Fierer, N., McCain, C.M., Meir, P., Zimmermann, M., Rapp, J.M., Silman, M.R., Knight, R.,  
 516 2011. Microbes do not follow the elevational diversity patterns of plants and animals.  
 517 *Ecology* 92, 797–804. <https://doi.org/10.1890/10-1170.1>
- 518 Fierer, N., Strickland, M.S., Liptzin, D., Bradford, M.A., Cleveland, C.C., 2009. Global patterns in  
 519 belowground communities. *Ecol. Lett.* 12, 1238–1249. <https://doi.org/10.1111/j.1461-0248.2009.01360.x>
- 520  
 521 Frostegård, A., Bååth, E., 1996. The use of phospholipid fatty acid analysis to estimate bacterial

522 and fungal biomass in soil. *Biol. Fertil. Soils* 22, 59–65.  
523 <https://doi.org/10.1007/BF00384433>

524 Griffiths, R.I., Thomson, B.C., James, P., Bell, T., Bailey, M., Whiteley, A.S., 2011. The bacterial  
525 biogeography of British soils: Mapping soil bacteria. *Environ. Microbiol.* 13, 1642–1654.  
526 <https://doi.org/10.1111/j.1462-2920.2011.02480.x>

527 Hartmann, M., Six, J., 2022. Soil structure and microbiome functions in agroecosystems. *Nat.*  
528 *Rev. Earth Environ.* 4, 4–18. <https://doi.org/10.1038/s43017-022-00366-w>

529 He, X., Chu, C., Yang, Y., Shu, Z., Li, B., Hou, E., 2021. Bedrock and climate jointly control the  
530 phosphorus status of subtropical forests along two elevational gradients. *CATENA* 206,  
531 105525. <https://doi.org/10.1016/j.catena.2021.105525>

532 He, X., Hou, E., Veen, G.F., Ellwood, M.D.F., Dijkstra, P., Sui, X., Zhang, S., Wen, D., Chu, C.,  
533 2020. Soil microbial biomass increases along elevational gradients in the tropics and  
534 subtropics but not elsewhere. *Glob. Ecol. Biogeogr.* 29, 345–354.  
535 <https://doi.org/10.1111/geb.13017>

536 He, X., Zeng, L., Zhu, G., Ellwood, M.D.F., Zhou, L., Huang, J., Wang, C., Li, W., Lin, D., Wei,  
537 P., Liu, S., Luo, M., Zhang, Y., Yang, Y., 2024. Parental material and climate jointly  
538 determine the biomass and diversity of soil microbial communities along an elevational  
539 gradient on a subtropical karst mountain. *J. Biogeogr.* <https://doi.org/10.1111/jbi.14814>

540 Hendershot, J.N., Read, Q.D., Henning, J.A., Sanders, N.J., Classen, A.T., 2017. Consistently  
541 inconsistent drivers of microbial diversity and abundance at macroecological scales.  
542 *Ecology* 98, 1757–1763. <https://doi.org/10.1002/ecy.1829>

543 Hu, A., Wang, Jianjun, Sun, H., Niu, B., Si, G., Wang, Jian, Yeh, C.-F., Zhu, X., Lu, X., Zhou, J.,  
544 Yang, Y., Ren, M., Hu, Y., Dong, H., Zhang, G., 2020. Mountain biodiversity and  
545 ecosystem functions: interplay between geology and contemporary environments. *ISME*  
546 *J.* 14, 931–944. <https://doi.org/10.1038/s41396-019-0574-x>

547 Koranda, M., Kaiser, C., Fuchslueger, L., Kitzler, B., Sessitsch, A., Zechmeister-Boltenstern, S.,  
548 Richter, A., 2014. Fungal and bacterial utilization of organic substrates depends on  
549 substrate complexity and N availability. *FEMS Microbiol. Ecol.* 87, 142–152.  
550 <https://doi.org/10.1111/1574-6941.12214>

551 Lang, F., Bauhus, J., Frossard, E., George, E., Kaiser, K., Kaupenjohann, M., Krüger, J.,  
552 Matzner, E., Polle, A., Priezel, J., Rennenberg, H., Wellbrock, N., 2016. Phosphorus in  
553 forest ecosystems: New insights from an ecosystem nutrition perspective. *J. Plant Nutr.*  
554 *Soil Sci.* 179, 129–135. <https://doi.org/10.1002/jpln.201500541>

555 Li, G., Xu, G., Shen, C., Tang, Y., Zhang, Y., Ma, K., 2016. Contrasting elevational diversity  
556 patterns for soil bacteria between two ecosystems divided by the treeline. *Sci. China Life*  
557 *Sci.* 59, 1177–1186. <https://doi.org/10.1007/s11427-016-0072-6>

558 Li, J., Shen, Z., Li, C., Kou, Y., Wang, Y., Tu, B., Zhang, S., Li, X., 2018. Stair-Step Pattern of  
559 Soil Bacterial Diversity Mainly Driven by pH and Vegetation Types Along the Elevational  
560 Gradients of Gongga Mountain, China. *Front. Microbiol.* 9, 569.  
561 <https://doi.org/10.3389/fmicb.2018.00569>

562 Li, X., Wang, T., Chang, S.X., Jiang, X., Song, Y., 2020. Biochar increases soil microbial  
563 biomass but has variable effects on microbial diversity: A meta-analysis. *Sci. Total*  
564 *Environ.* 749, 141593. <https://doi.org/10.1016/j.scitotenv.2020.141593>

565 Looby, C.I., Martin, P.H., 2020. Diversity and function of soil microbes on montane gradients:  
566 the state of knowledge in a changing world. *FEMS Microbiol. Ecol.* 96, fiae122.  
567 <https://doi.org/10.1093/femsec/fiae122>

568 Luan, L., Jiang, Y., Dini-Andreote, F., Crowther, T.W., Li, P., Bahram, M., Zheng, J., Xu, Q.,  
569 Zhang, X.-X., Sun, B., 2023. Integrating pH into the metabolic theory of ecology to  
570 predict bacterial diversity in soil. *Proc. Natl. Acad. Sci.* 120, e2207832120.  
571 <https://doi.org/10.1073/pnas.2207832120>

572 Ma, L., Liu, L., Lu, Y., Chen, L., Zhang, Z., Zhang, H., Wang, X., Shu, L., Yang, Q., Song, Q.,

573 Peng, Q., Yu, Z., Zhang, J., 2022. When microclimates meet soil microbes: Temperature  
574 controls soil microbial diversity along an elevational gradient in subtropical forests. *Soil*  
575 *Biol. Biochem.* 166, 108566. <https://doi.org/10.1016/j.soilbio.2022.108566>

576 Ni, X., Liao, S., Wu, F., Groffman, P.M., 2022. Microbial biomass in forest soils under altered  
577 moisture conditions: A review. *Soil Sci. Soc. Am. J.* 86, 358–368.  
578 <https://doi.org/10.1002/saj2.20344>

579 Nottingham, A.T., Hicks, L.C., Ccahuana, A.J.Q., Salinas, N., Bååth, E., Meir, P., 2018. Nutrient  
580 limitations to bacterial and fungal growth during cellulose decomposition in tropical forest  
581 soils. *Biol. Fertil. Soils* 54, 219–228. <https://doi.org/10.1007/s00374-017-1247-4>

582 Peters, M.K., Hemp, A., Appelhans, T., Becker, J.N., Behler, C., Classen, A., Detsch, F.,  
583 Ensslin, A., Ferger, S.W., Frederiksen, S.B., Gebert, F., Gerschlauser, F., Gütlein, A.,  
584 Helbig-Bonitz, M., Hemp, C., Kindeketa, W.J., Kühnel, A., Mayr, A.V., Mwangomo, E.,  
585 Ngereza, C., Njovu, H.K., Otte, I., Pabst, H., Renner, M., Röder, J., Rutten, G.,  
586 Schellenberger Costa, D., Sierra-Cornejo, N., Vollstädt, M.G.R., Dulle, H.I., Eardley,  
587 C.D., Howell, K.M., Keller, A., Peters, R.S., Ssymank, A., Kakengi, V., Zhang, J.,  
588 Bogner, C., Böhning-Gaese, K., Brandl, R., Hertel, D., Huwe, B., Kiese, R., Kleyer, M.,  
589 Kuzyakov, Y., Nauss, T., Schleuning, M., Tschapka, M., Fischer, M., Steffan-Dewenter,  
590 I., 2019. Climate–land-use interactions shape tropical mountain biodiversity and  
591 ecosystem functions. *Nature* 568, 88–92. <https://doi.org/10.1038/s41586-019-1048-z>

592 Peters, M.K., Hemp, A., Appelhans, T., Behler, C., Classen, A., Detsch, F., Ensslin, A., Ferger,  
593 S.W., Frederiksen, S.B., Gebert, F., Haas, M., Helbig-Bonitz, M., Hemp, C., Kindeketa,  
594 W.J., Mwangomo, E., Ngereza, C., Otte, I., Röder, J., Rutten, G., Schellenberger Costa,  
595 D., Tardanico, J., Zancolli, G., Deckert, J., Eardley, C.D., Peters, R.S., Rödel, M.-O.,  
596 Schleuning, M., Ssymank, A., Kakengi, V., Zhang, J., Böhning-Gaese, K., Brandl, R.,  
597 Kalko, E.K.V., Kleyer, M., Nauss, T., Tschapka, M., Fischer, M., Steffan-Dewenter, I.,  
598 2016. Predictors of elevational biodiversity gradients change from single taxa to the  
599 multi-taxa community level. *Nat. Commun.* 7, 13736.  
600 <https://doi.org/10.1038/ncomms13736>

601 Philippot, L., Chenu, C., Kappler, A., Rillig, M.C., Fierer, N., 2023. The interplay between  
602 microbial communities and soil properties. *Nat. Rev. Microbiol.*  
603 <https://doi.org/10.1038/s41579-023-00980-5>

604 Porder, S., Ramachandran, S., 2013. The phosphorus concentration of common rocks—a  
605 potential driver of ecosystem P status. *Plant Soil* 367, 41–55.  
606 <https://doi.org/10.1007/s11104-012-1490-2>

607 Réjou-Méchain, M., Tanguy, A., Pioniot, C., Chave, J., Hérault, B., 2017. biomass : an r  
608 package for estimating above-ground biomass and its uncertainty in tropical forests.  
609 *Methods Ecol. Evol.* 8, 1163–1167. <https://doi.org/10.1111/2041-210X.12753>

610 Ren, C., Chen, J., Lu, X., Doughty, R., Zhao, F., Zhong, Z., Han, X., Yang, G., Feng, Y., Ren,  
611 G., 2018. Responses of soil total microbial biomass and community compositions to  
612 rainfall reductions. *Soil Biol. Biochem.* 116, 4–10.  
613 <https://doi.org/10.1016/j.soilbio.2017.09.028>

614 Rousk, J., Bååth, E., Brookes, P.C., Lauber, C.L., Lozupone, C., Caporaso, J.G., Knight, R.,  
615 Fierer, N., 2010. Soil bacterial and fungal communities across a pH gradient in an arable  
616 soil. *ISME J.* 4, 1340–1351. <https://doi.org/10.1038/ismej.2010.58>

617 Seaton, F.M., George, P.B.L., Lebron, I., Jones, D.L., Creer, S., Robinson, D.A., 2020. Soil  
618 textural heterogeneity impacts bacterial but not fungal diversity. *Soil Biol. Biochem.* 144,  
619 107766. <https://doi.org/10.1016/j.soilbio.2020.107766>

620 Sheng, R., Xu, H., Xing, X., Zhang, W., Hou, H., Qin, H., Liu, Y., Zhang, L., Fang, Y., Shen, J.,  
621 Pernthaler, J., Wei, W., Zhu, B., 2023. The role of inherited characteristics from parent  
622 materials in shaping bacterial communities in agricultural soils. *Geoderma* 433, 116455.  
623 <https://doi.org/10.1016/j.geoderma.2023.116455>

- 624 Singh, D., Lee-Cruz, L., Kim, W.-S., Kerfahi, D., Chun, J.-H., Adams, J.M., 2014. Strong  
625 elevational trends in soil bacterial community composition on Mt. Halla, South Korea.  
626 *Soil Biol. Biochem.* 68, 140–149. <https://doi.org/10.1016/j.soilbio.2013.09.027>
- 627 Smith, L.C., Orgiazzi, A., Eisenhauer, N., Cesarz, S., Lochner, A., Jones, A., Bastida, F.,  
628 Patoine, G., Reitz, T., Buscot, F., Rillig, M.C., Heintz-Buschart, A., Lehmann, A., Guerra,  
629 C.A., Waring, B.G., 2021. Large-scale drivers of relationships between soil microbial  
630 properties and organic carbon across Europe. *Glob. Ecol. Biogeogr.* 30, 2070–2083.  
631 <https://doi.org/10.1111/geb.13371>
- 632 Spinola, D., Portes, R., Fedenko, J., Lybrand, R., Dere, A., Biles, F., Trainor, T., Bowden, M.E.,  
633 D'Amore, D., 2022. Lithological controls on soil geochemistry and clay mineralogy  
634 across Spodosols in the coastal temperate rainforest of southeast Alaska. *Geoderma*  
635 428, 116211. <https://doi.org/10.1016/j.geoderma.2022.116211>
- 636 Sun, L., Xun, W., Huang, T., Zhang, G., Gao, J., Ran, W., Li, D., Shen, Q., Zhang, R., 2016.  
637 Alteration of the soil bacterial community during parent material maturation driven by  
638 different fertilization treatments. *Soil Biol. Biochem.* 96, 207–215.  
639 <https://doi.org/10.1016/j.soilbio.2016.02.011>
- 640 Sundqvist, M.K., Sanders, N.J., Wardle, D.A., 2013. Community and Ecosystem Responses to  
641 Elevational Gradients: Processes, Mechanisms, and Insights for Global Change. *Annu.*  
642 *Rev. Ecol. Evol. Syst.* 44, 261–280. [https://doi.org/10.1146/annurev-ecolsys-110512-](https://doi.org/10.1146/annurev-ecolsys-110512-135750)  
643 [135750](https://doi.org/10.1146/annurev-ecolsys-110512-135750)
- 644 Tripathi, B.M., Stegen, J.C., Kim, M., Dong, K., Adams, J.M., Lee, Y.K., 2018. Soil pH mediates  
645 the balance between stochastic and deterministic assembly of bacteria. *ISME J.* 12,  
646 1072–1083. <https://doi.org/10.1038/s41396-018-0082-4>
- 647 Tytgat, B., Verleyen, E., Sweetlove, M., D'hondt, S., Clercx, P., Van Ranst, E., Peeters, K.,  
648 Roberts, S., Namsaraev, Z., Wilmotte, A., Vyverman, W., Willems, A., 2016. Bacterial  
649 community composition in relation to bedrock type and macrobiota in soils from the Sør  
650 Rondane Mountains, East Antarctica. *FEMS Microbiol. Ecol.* 92, fiw126.  
651 <https://doi.org/10.1093/femsec/fiw126>
- 652 Wang, R., He, X., Zhang, Q., Li, B., Shu, Z., Chu, C., 2024. Inconsistent elevational patterns of  
653 soil microbial biomass, diversity, and community structure on four elevational transects  
654 from subtropical forests. *Appl. Soil Ecol.* 201, 105462.  
655 <https://doi.org/10.1016/j.apsoil.2024.105462>
- 656 Weemstra, M., Peay, K.G., Davies, S.J., Mohamad, M., Itoh, A., Tan, S., Russo, S.E., 2020.  
657 Lithological constraints on resource economies shape the mycorrhizal composition of a  
658 Bornean rain forest. *New Phytol.* 228, 253–268. <https://doi.org/10.1111/nph.16672>
- 659 Xiao, D., He, X., Zhang, W., Cheng, M., Hu, P., Wang, K., 2022. Diazotroph and arbuscular  
660 mycorrhizal fungal diversity and community composition responses to karst and non-  
661 karst soils. *Appl. Soil Ecol.* 170, 104227. <https://doi.org/10.1016/j.apsoil.2021.104227>
- 662 Yu, K., Van Den Hoogen, J., Wang, Z., Averill, C., Routh, D., Smith, G.R., Drenovsky, R.E.,  
663 Scow, K.M., Mo, F., Waldrop, M.P., Yang, Y., Tang, W., De Vries, F.T., Bardgett, R.D.,  
664 Manning, P., Bastida, F., Baer, S.G., Bach, E.M., García, C., Wang, Q., Ma, L., Chen,  
665 B., He, X., Teurlincx, S., Heijboer, A., Bradley, J.A., Crowther, T.W., 2022. The  
666 biogeography of relative abundance of soil fungi versus bacteria in surface topsoil. *Earth*  
667 *Syst. Sci. Data* 14, 4339–4350. <https://doi.org/10.5194/essd-14-4339-2022>
- 668 Zakavi, M., Askari, H., Shahrooei, M., 2022. Bacterial diversity changes in response to an  
669 altitudinal gradient in arid and semi-arid regions and their effects on crops growth. *Front.*  
670 *Microbiol.* 13, 984925. <https://doi.org/10.3389/fmicb.2022.984925>
- 671 Zeng, L., He, X., Zhu, G., Zhou, L., Luo, M., Yin, X., Long, Y., Dai, J., Ouyang, X., Yang, Y.,  
672 2023. Bedrock and climate jointly control microbial necromass along a subtropical  
673 elevational gradient. *Appl. Soil Ecol.* 189, 104902.  
674 <https://doi.org/10.1016/j.apsoil.2023.104902>

675 Zhou, G., Peng, C., Li, Y., Liu, S., Zhang, Q., Tang, X., Liu, J., Yan, J., Zhang, D., Chu, G.,  
676 2013. A climate change-induced threat to the ecological resilience of a subtropical  
677 monsoon evergreen broad-leaved forest in Southern China. *Glob. Change Biol.* 19,  
678 1197–1210. <https://doi.org/10.1111/gcb.12128>  
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680 **Table 1. Results of the Wilcoxon test to compare the mean values of soil microbial**  
 681 **biomass and  $\alpha$ -diversities between two bedrock transects.** Significantly higher mean values  
 682 are in bold. Unit of biomass is  $\mu\text{g g}^{-1}$  soil.

Variable	Granite (Mean $\pm$ SD)	Slate (Mean $\pm$ SD)	<i>w</i>	<i>p</i>
Fungal biomass	<b>1.33</b> $\pm$ 0.54	1.09 $\pm$ 0.37	1267	0.040
Bacterial biomass	38.62 $\pm$ 14.2	42.95 $\pm$ 9.58	791	0.074
Bacteria to fungi ratio	30.79 $\pm$ 10.49	<b>42.28</b> $\pm$ 11.66	476	<0.001
Gram+ to Gram- ratio	<b>0.81</b> $\pm$ 0.08	0.72 $\pm$ 0.05	1634	<0.001
Bacterial Chao1	3036 $\pm$ 294	<b>3344</b> $\pm$ 349	520	<0.001
Bacterial Shannon	5.89 $\pm$ 0.2	5.95 $\pm$ 0.27	864	0.234
Bacterial inv-Simpson	<b>104</b> $\pm$ 29	82 $\pm$ 28	1413	0.001
Fungal Chao1	1237 $\pm$ 187	<b>1364</b> $\pm$ 209	695	0.010
Fungal Shannon	4.12 $\pm$ 0.48	4.03 $\pm$ 0.68	1060	0.706
Fungal inv-Simpson	18 $\pm$ 8	16 $\pm$ 9	1119	0.394

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685 **Table 2. Effects of elevation and bedrock, and their interaction on the variations of soil**  
 686 **microbial communities' characters.** Numbers in the table are the standardized linear  
 687 regression coefficients.

	Elevation	Bedrock	Elevation×Bedrock	Adjusted R <sup>2</sup>
Bacterial biomass	0.88***	0.36*	-1.01***	0.372
Fungal biomass	0.80***	-0.49*	-1.08***	0.373
B:F biomass ratio	0.19*	0.93***		0.231
G <sup>+</sup> to G <sup>-</sup> biomass ratio	0.31***	-1.06***		0.384
Bacterial Chao1 index	0.11	0.87***		0.182
Bacterial Shannon index	0.04	0.26		0.001
Bacterial inverse Simpson	0.17	-0.71***		0.140
Fungal Chao1 index	0.28**	0.62**		0.151
Fungal Shannon index	0.09	-0.12		0.001
Fungal inverse Simpson	-0.15	-0.20		0.001

688 Stars next to the numbers indicate significance in the regression model: \*, \*\*, \*\*\* indicates  
 689 significance at the 95%, 99% and 99.9% level, respectively; no star means  $p > 0.05$ .

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692 **Table 3 Summary of optimal model parameters for microbial biomass variables.** This table  
 693 presents the best-fit model results for microbial biomass variables, derived from an initial full  
 694 model that included mean annual temperature (MAT), soil organic carbon (C), phosphorus (P),  
 695 moisture, clay content, pH, carbon-to-nitrogen ratio (C:N), above-ground biomass (AGB), and  
 696 plant Shannon diversity (Plant H). Additionally, interactions between bedrock type and specific  
 697 environmental variables (soil P, moisture, pH, and MAT) were incorporated.

Microbial Variable	Predictors	Estimate	p-value	Adjusted R <sup>2</sup>
Bacterial Biomass	Soil C	9.316	<0.001	0.625
	Soil C:N	-4.936	<0.001	
	Soil P	6.011	0.008	
	Clay	-2.065	0.070	
	MAT×Bedrock	4.294	0.022	
	Soil P×Bedrock	-11.818	<0.001	
Fungal Biomass	Soil C	0.267	<0.001	0.453
	Soil C:N	-0.216	<0.001	
	Soil P	-0.351	<0.001	
	Moisture×Bedrock	0.288	0.009	
	MAT×Bedrock	0.221	0.047	
Bacterial-to-Fungal Biomass ratio	Moisture	3.715	0.002	0.502
	Plant H	3.163	0.001	
	Soil P	6.042	<0.001	
	Moisture×Bedrock	-9.585	<0.001	
Gram <sup>+</sup> to Gram <sup>-</sup> bacteria biomass ratio	Clay	-0.015	0.046	0.710
	MAT	0.014	0.018	
	pH	-0.027	0.028	
	Soil P×Bedrock	0.060	0.020	



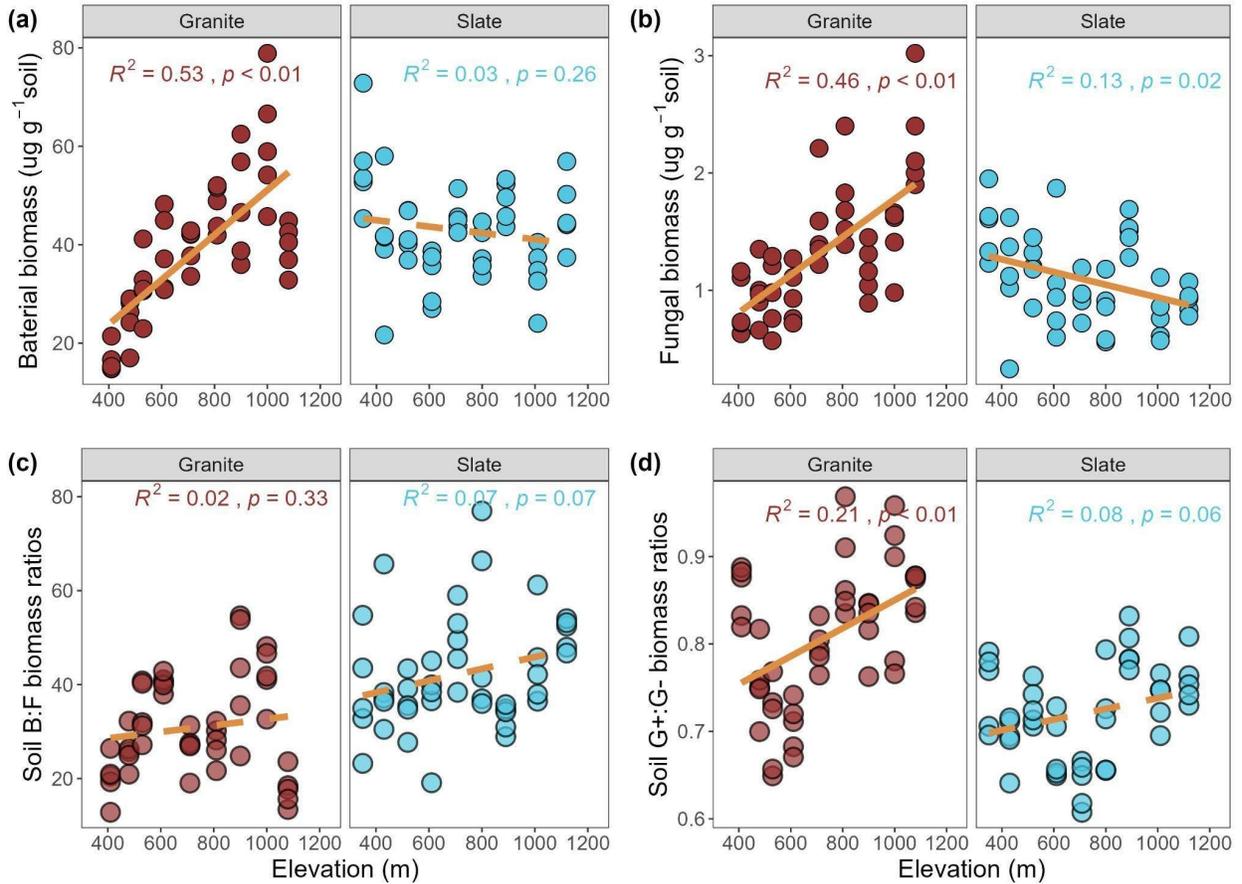
699 **Table 4 Summary of optimal model parameters for microbial community diversity index.**

700 This table presents the best-fit model results for microbial diversity index, derived from an initial  
 701 full model that included mean annual temperature (MAT), soil organic carbon (C), phosphorus  
 702 (P), moisture, clay content, pH, carbon-to-nitrogen ratio, above-ground biomass (AGB), and  
 703 plant Shannon diversity. Additionally, interactions between bedrock type and specific  
 704 environmental variables (soil P, moisture, pH, and MAT) were incorporated.

Microbial Variable	Predictors	Estimate	<i>p</i> -value	Adjusted R <sup>2</sup>
Bacterial Shannon index	Clay	0.068	0.008	0.271
	MAT	-0.098	<0.001	
	Moisture	-0.063	0.033	
	pH	0.075	0.004	
Bacterial inverse Simpson index	Soil C	9.249	0.013	0.202
	Soil P	-12.539	<0.001	
	Moisture×Bedrock	15.275	0.020	
Fungal Shannon index	Soil P	-0.608	0.000	0.383
	AGB	-0.155	0.019	
	Moisture×Bedrock	0.081	0.119	
	Soil P×Bedrock	1.145	<0.001	
Fungal inverse Simpson index	Moisture	4.086	<0.001	0.332
	Soil P	-6.853	0.001	
	Soil C	-3.039	0.045	
	pH×Bedrock	7.561	0.004	
	Soil P×Bedrock	16.332	<0.001	

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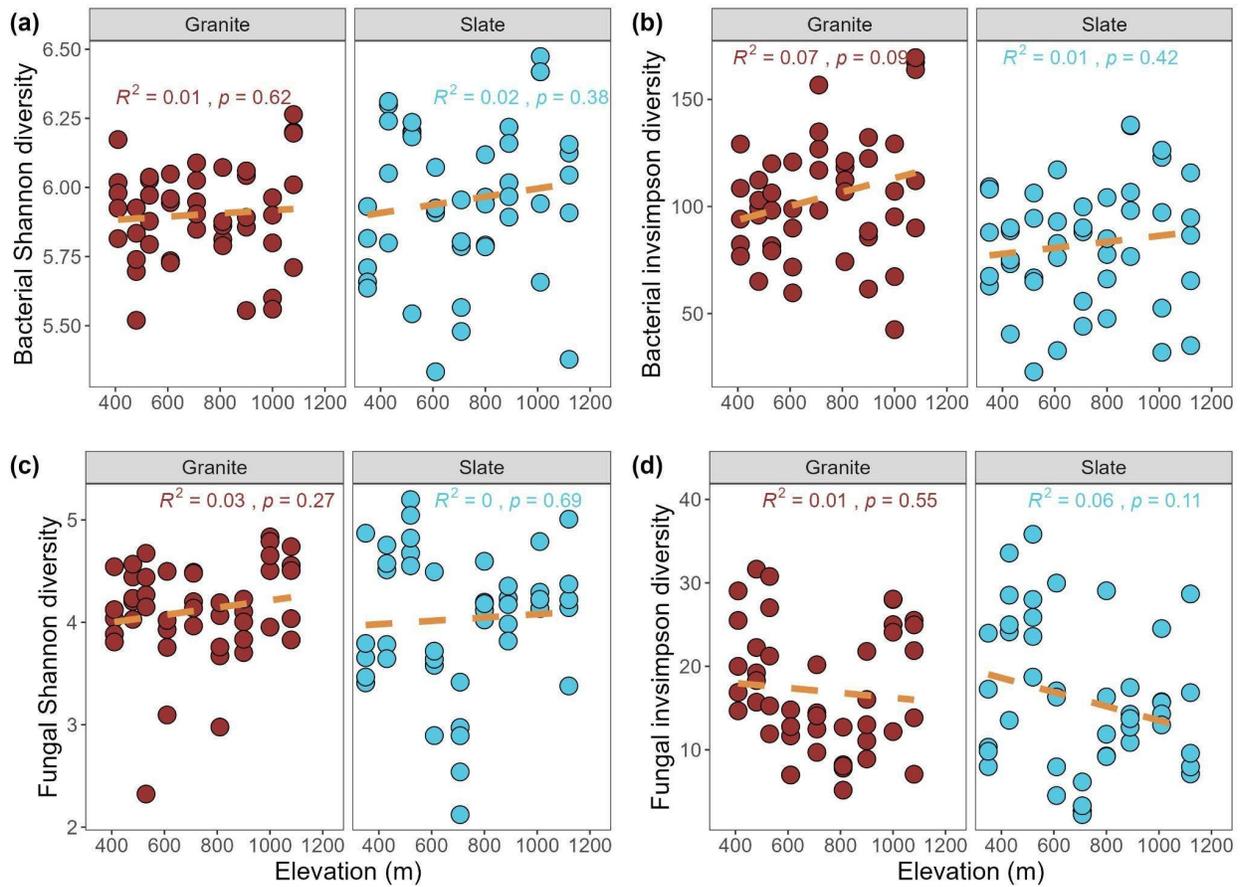
706 **Figure 1. Soil microbial biomass along elevational transects on granite and slate**  
 707 **bedrock.** (a) soil bacterial biomass; (b) soil fungal biomass; (c) bacterial biomass to fungal  
 708 biomass ratios; (d) gram-positive to gram-negative bacterial biomass ratios. Solid and dashed  
 709 lines indicate significant ( $p < 0.05$ ) and nonsignificant ( $p > 0.05$ ) linear regression relationships,  
 710 respectively.



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713 **Figure 2. Elevational patterns of soil microbial community  $\alpha$ -diversities.** (a and b) Shannon  
 714 and inverse Simpson diversity index of bacterial communities, respectively; (c and d) Shannon  
 715 and inverse Simpson diversity index of fungal communities, respectively. Dashed lines indicate  
 716 nonsignificant ( $p > 0.05$ ) linear regression relationships.



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## Supplementary information of

### Bedrock modulates the elevational patterns of soil microbial communities

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748 **Table S1. Comparisons of climate, soil, and vegetation on contrasting bedrocks.**

Variable	Granite	Slate	w	p	Significance
MAT	21 ± 2.52	21.81 ± 3.47	875	0.268	ns
AGB	<b>20.46</b> ± 7.95	14.33 ± 3.86	1550	0.000	*
Plant Shannon	2.66 ± 0.49	<b>2.89</b> ± 0.37	675	0.006	*
Soil pH	4.41 ± 0.18	<b>4.67</b> ± 0.28	373	0.000	*
Soil moisture	0.18 ± 0.05	0.17 ± 0.03	950	0.616	ns
Bulk density	0.76 ± 0.21	<b>0.83</b> ± 0.15	723	0.019	*
SOC	81.49 ± 34.39	84.87 ± 27.09	880	0.288	ns
Soil TN	2.02 ± 0.88	<b>2.42</b> ± 0.63	661	0.004	*
Soil TP	0.23 ± 0.1	<b>0.49</b> ± 0.09	41	0.000	*
Soil C:N	<b>41.93</b> ± 12.09	34.76 ± 4.88	1469	0.000	*
Soil C:P	<b>397</b> ± 166	171 ± 40	1865	0.000	*
Soil N:P	<b>9.33</b> ± 2.64	4.9 ± 0.8	1868	0.000	*
Clay	40 ± 18.48	<b>64.78</b> ± 18.82	325	0.000	*
Silt	<b>28.02</b> ± 9.26	23.36 ± 8.69	1375	0.003	*
Sand	<b>31.67</b> ± 26.14	11.86 ± 21.32	1400	0.002	*
Slope direction	72 ± 63	51 ± 29	1038	0.843	ns

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752 **Table S2.** Spearman correlation coefficients between soil microbial communities' characters  
 753 and environmental variables using all data from two transects. Significant coefficients ( $p < 0.05$ )  
 754 are in bold. B.Biomass: Bacterial biomass; F.Biomass: Fungal biomass; B.F.ratio: ratios of  
 755 bacterial to fungal biomass; G.G.ratio: ratios of gram positive to gram negative bacterial  
 756 biomass; Bacteria.H: Bacterial Shannon index; Bacteria.invs: Bacterial inverse Simpson index;  
 757 Fungi.H: Fungal Shannon index; Fungi.invs: Fungal inverse Simpson index.

	MAT	pH	Moisture	Clay	AGB	Plant.H	SOC	Soil.P	C:N	C:P	N:P
<b>All</b>											
B.Biomass	<b>-0.24</b>	-0.13	0.21	0.15	<b>-0.37</b>	-0.01	<b>0.56</b>	<b>0.38</b>	-0.10	-0.04	0.01
F.Biomass	-0.15	-0.06	0.05	-0.06	-0.20	-0.10	0.17	-0.17	0.01	<b>0.23</b>	<b>0.30</b>
B.F.ratio	-0.13	-0.03	<b>0.21</b>	<b>0.32</b>	-0.19	0.12	<b>0.33</b>	<b>0.63</b>	<b>-0.26</b>	<b>-0.41</b>	<b>-0.42</b>
G.G.ratio	<b>-0.29</b>	<b>-0.32</b>	-0.07	<b>-0.37</b>	-0.12	<b>-0.51</b>	<b>0.29</b>	<b>-0.40</b>	<b>0.38</b>	<b>0.67</b>	<b>0.69</b>
Bacteria.H	-0.16	<b>0.27</b>	-0.05	0.19	-0.15	-0.07	-0.07	0.06	0.03	-0.08	-0.11
Bacteria.invs	-0.13	<b>-0.23</b>	-0.09	-0.17	0.09	<b>-0.21</b>	0.07	<b>-0.35</b>	<b>0.34</b>	<b>0.41</b>	<b>0.38</b>
Fungi.H	<b>-0.28</b>	0.06	0.15	-0.02	-0.25	-0.03	-0.02	0.10	-0.11	-0.07	-0.09
Fungi.invs	-0.04	0.15	<b>0.27</b>	-0.05	-0.11	-0.05	-0.20	-0.04	-0.12	-0.10	-0.12

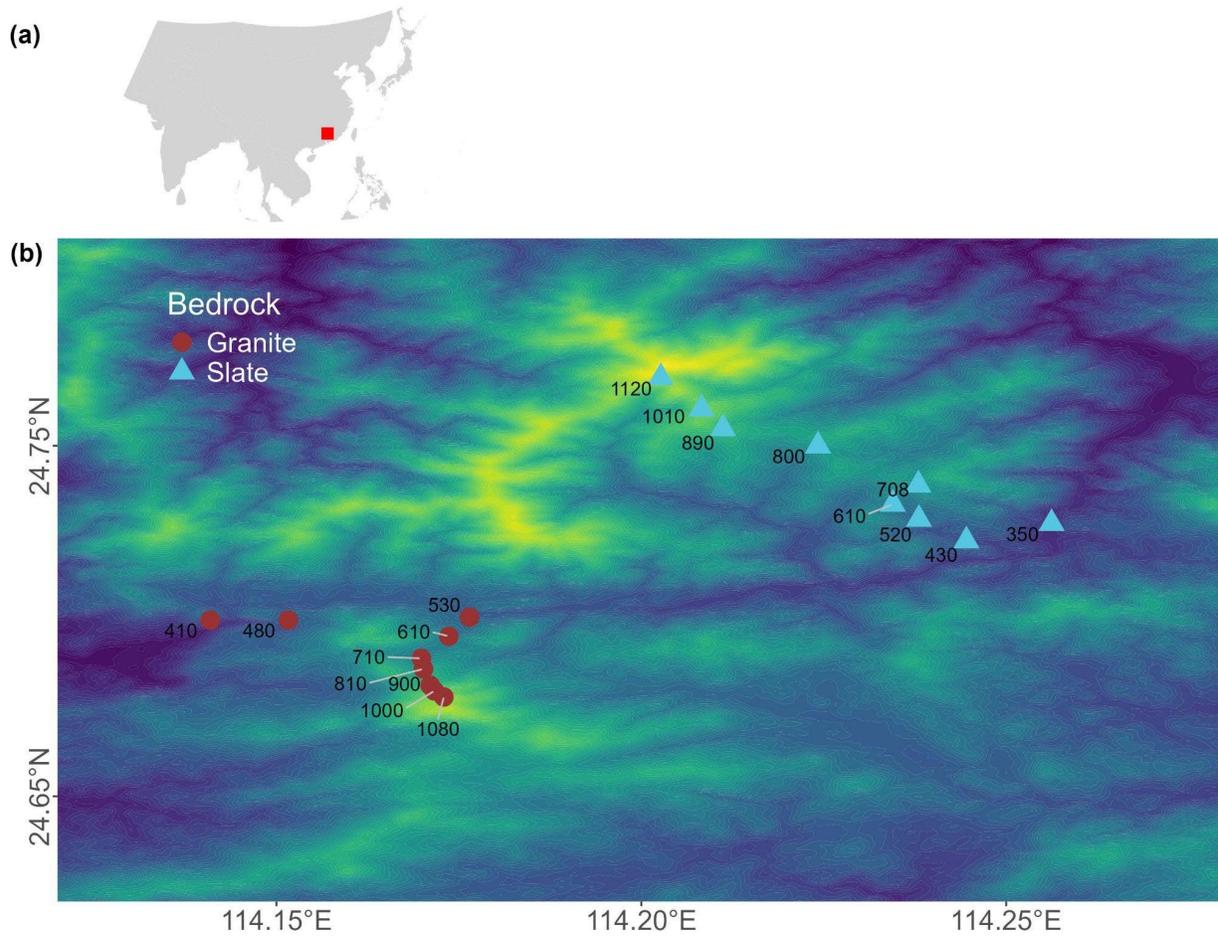
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759 **Table S3. Spearman correlation coefficients between soil microbial communities'**  
760 **characters and environmental variables on granite and slate, respectively.** Significant  
761 coefficients ( $p < 0.05$ ) are in bold. B.Biomass: Bacterial biomass; F.Biomass: Fungal biomass;  
762 B.F.ratio: ratios of bacterial to fungal biomass; G.G.ratio: ratios of gram positive to gram  
763 negative bacterial biomass; Bacteria.H: Bacterial Shannon index; Bacteria.invs: Bacterial  
764 inverse Simpson index; Funghi.H: Fungal Shannon index; Fungal.invs: Fungal inverse Simpson  
765 index.

	MAT	pH	Moisture	Clay	AGB	Plant.H	SOC	Soil.P	C:N	C:P	N:P
<b>Granite</b>											
B.Biomass	<b>-0.37</b>	<b>-0.31</b>	0.24	0.27	<b>-0.38</b>	0.09	<b>0.76</b>	<b>0.66</b>	-0.23	0.01	0.23
F.Biomass	<b>-0.36</b>	0.01	<b>-0.33</b>	0.00	<b>-0.53</b>	-0.16	<b>0.43</b>	0.13	-0.06	0.25	<b>0.51</b>
B.F.ratio	-0.07	<b>-0.41</b>	<b>0.72</b>	<b>0.40</b>	0.15	<b>0.32</b>	<b>0.38</b>	<b>0.71</b>	<b>-0.35</b>	<b>-0.40</b>	<b>-0.39</b>
G.G.ratio	-0.19	0.06	<b>-0.35</b>	<b>-0.34</b>	<b>-0.52</b>	<b>-0.53</b>	<b>0.35</b>	-0.17	0.26	<b>0.55</b>	<b>0.65</b>
Bacteria.H	-0.08	<b>0.38</b>	-0.23	0.14	0.02	-0.26	-0.04	-0.12	0.23	0.17	0.16
Bacteria.invs	-0.15	0.09	<b>-0.30</b>	-0.03	-0.20	-0.09	0.12	-0.20	0.24	<b>0.34</b>	<b>0.37</b>
Funghi.H	<b>-0.30</b>	-0.03	0.19	<b>-0.30</b>	-0.22	-0.21	0.11	0.07	-0.07	-0.08	-0.07
Funghi.invs	-0.17	-0.01	0.26	<b>-0.45</b>	-0.15	-0.44	-0.04	-0.02	0.05	-0.08	-0.21
<b>Slate</b>											
B.Biomass	-0.07	-0.21	0.05	0.01	-0.15	-0.08	0.20	0.09	0.12	0.22	0.24
F.Biomass	0.12	0.20	<b>0.42</b>	0.18	-0.07	0.25	-0.16	-0.13	-0.12	-0.10	-0.04
B.F.ratio	-0.08	<b>-0.33</b>	<b>-0.51</b>	-0.17	0.03	<b>-0.40</b>	<b>0.30</b>	0.20	0.19	0.23	0.17
G.G.ratio	<b>-0.45</b>	-0.09	0.26	0.19	<b>-0.43</b>	-0.10	<b>0.36</b>	<b>0.43</b>	0.14	0.20	0.20
Bacteria.H	-0.26	<b>0.32</b>	0.00	0.11	-0.08	0.22	-0.13	-0.02	-0.06	-0.10	-0.17
Bacteria.invs	-0.09	-0.18	0.21	0.07	0.02	-0.09	0.12	-0.05	0.21	0.23	0.11
Funghi.H	<b>-0.36</b>	<b>0.33</b>	0.02	0.10	<b>-0.41</b>	<b>0.34</b>	-0.16	<b>0.34</b>	-0.22	<b>-0.34</b>	<b>-0.37</b>
Funghi.invs	-0.05	<b>0.45</b>	0.19	0.25	-0.24	<b>0.46</b>	<b>-0.37</b>	0.12	<b>-0.32</b>	<b>-0.47</b>	<b>-0.46</b>

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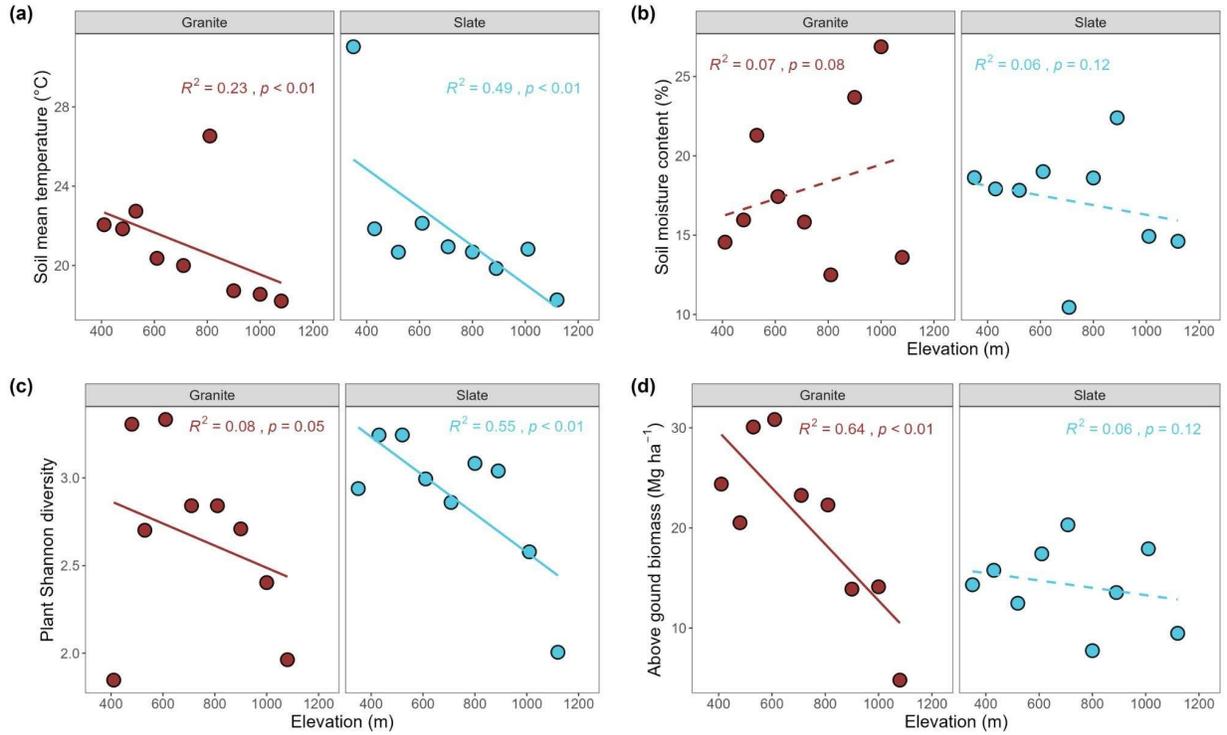
767 **Figure S1. Distribution of sampling sites along two subtropical elevational transects.** (a)  
768 Location of the sampling sites within East Asia, indicated by a red square; (b) Distribution of  
769 sites along the two transects, where red dots represent granite sites and blue dots represent  
770 slate sites. Elevations are marked by numbers near each dot. The base map is a color-coded  
771 DEM derived from SRTM 90m data.



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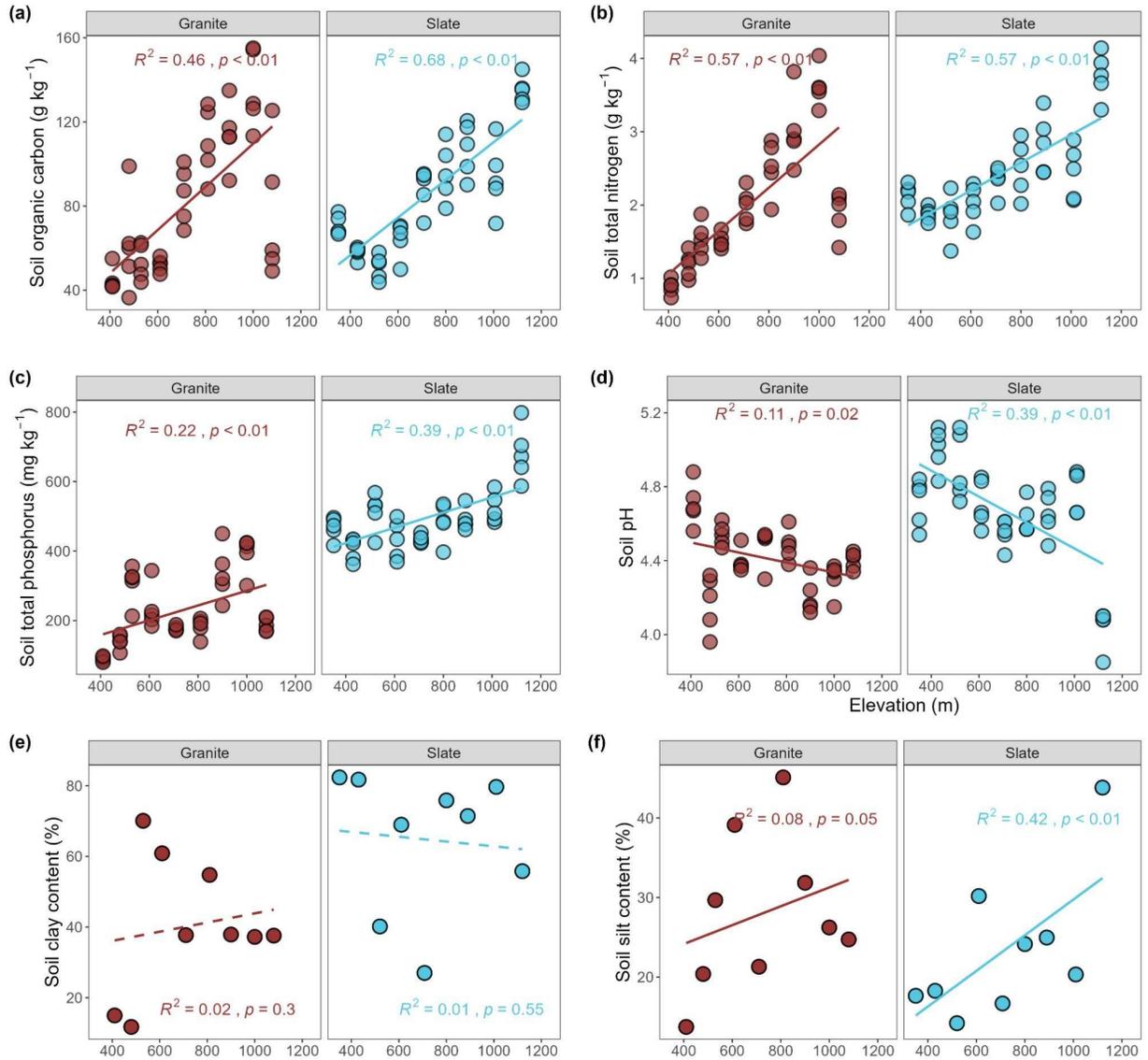
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774 **Figure S2. Elevational patterns of soil temperature, moisture, and vegetation**  
 775 **characteristics on granite and slate bedrock.** (a) Mean annual soil temperature; (b) Mean  
 776 annual soil moisture; (c) Plant diversity, measured by the Shannon index; (d) Above-ground  
 777 biomass. Solid and dashed lines indicate significant ( $p < 0.05$ ) and nonsignificant ( $p > 0.05$ )  
 778 linear regression relationships, respectively.



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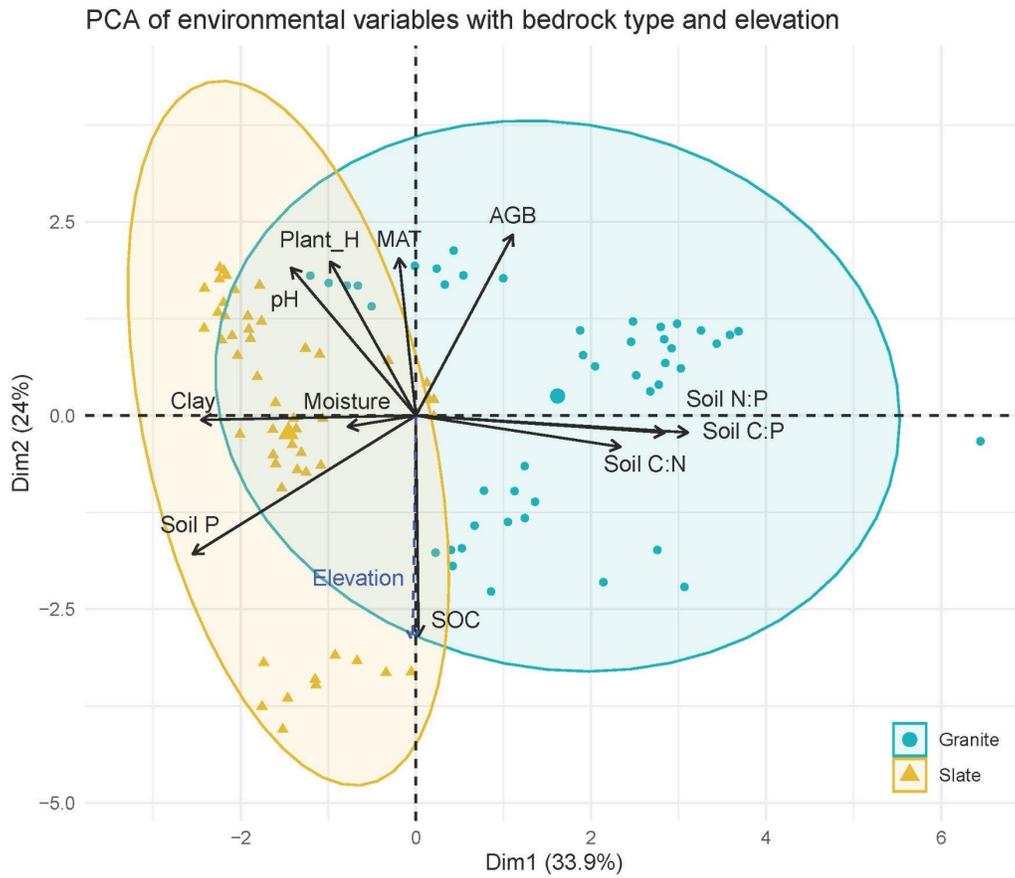
780 **Figure S3. Elevational patterns of soil physical-chemical properties on granite and slate**  
 781 **bedrock.** (A) soil organic carbon concentration; (B) soil total phosphorus concentration; (C) soil  
 782 clay content; (D) soil silt content; (E) soil pH; (F) soil moisture. Solid and dashed lines indicate  
 783 significant ( $p < 0.05$ ) and non-significant ( $p > 0.05$ ) linear regression relationships, respectively.



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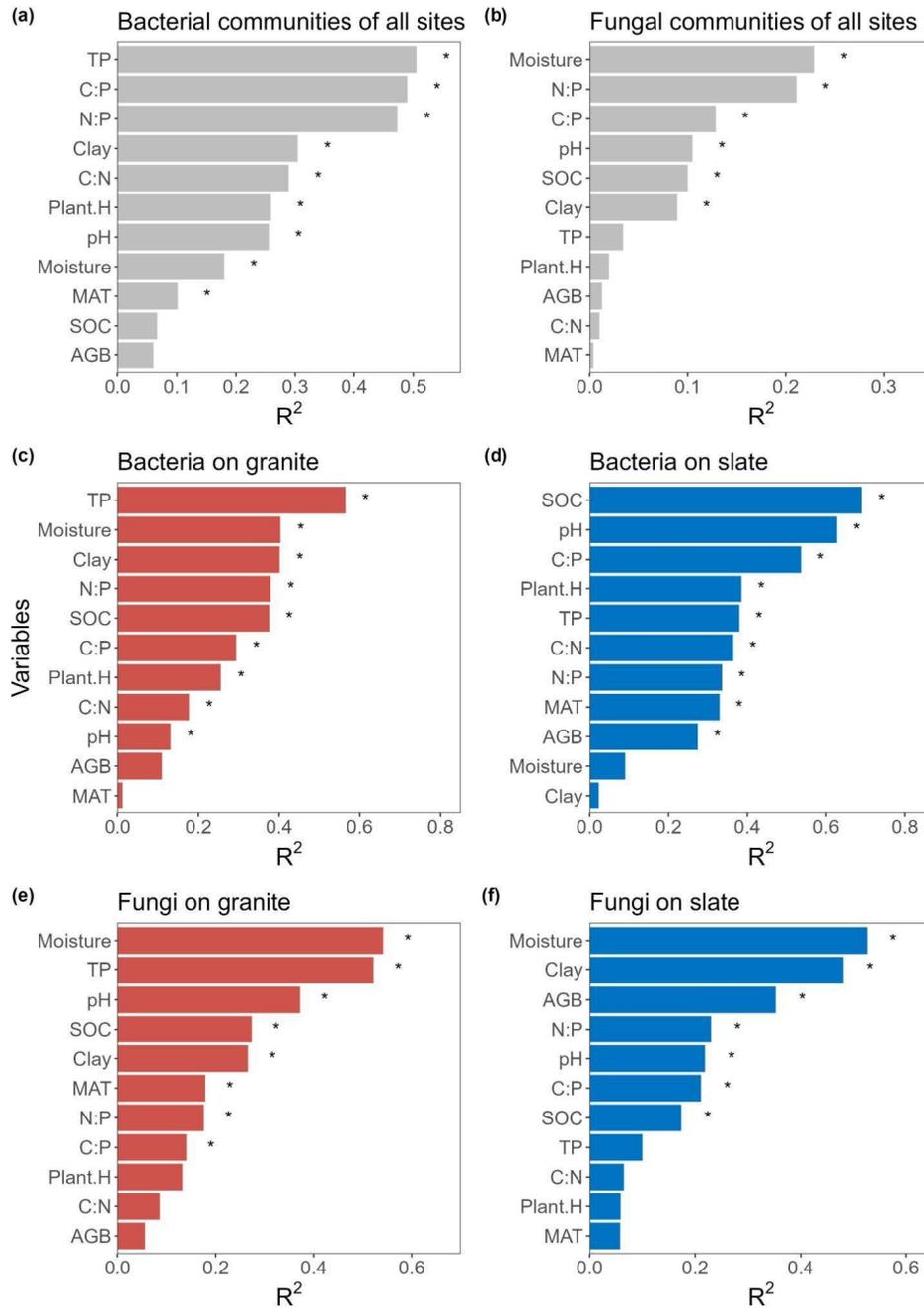
786 **Figure S4. Principal Component Analysis (PCA) of environmental variables vary along**  
787 **two elevational transects on granite and slate bedrocks.**



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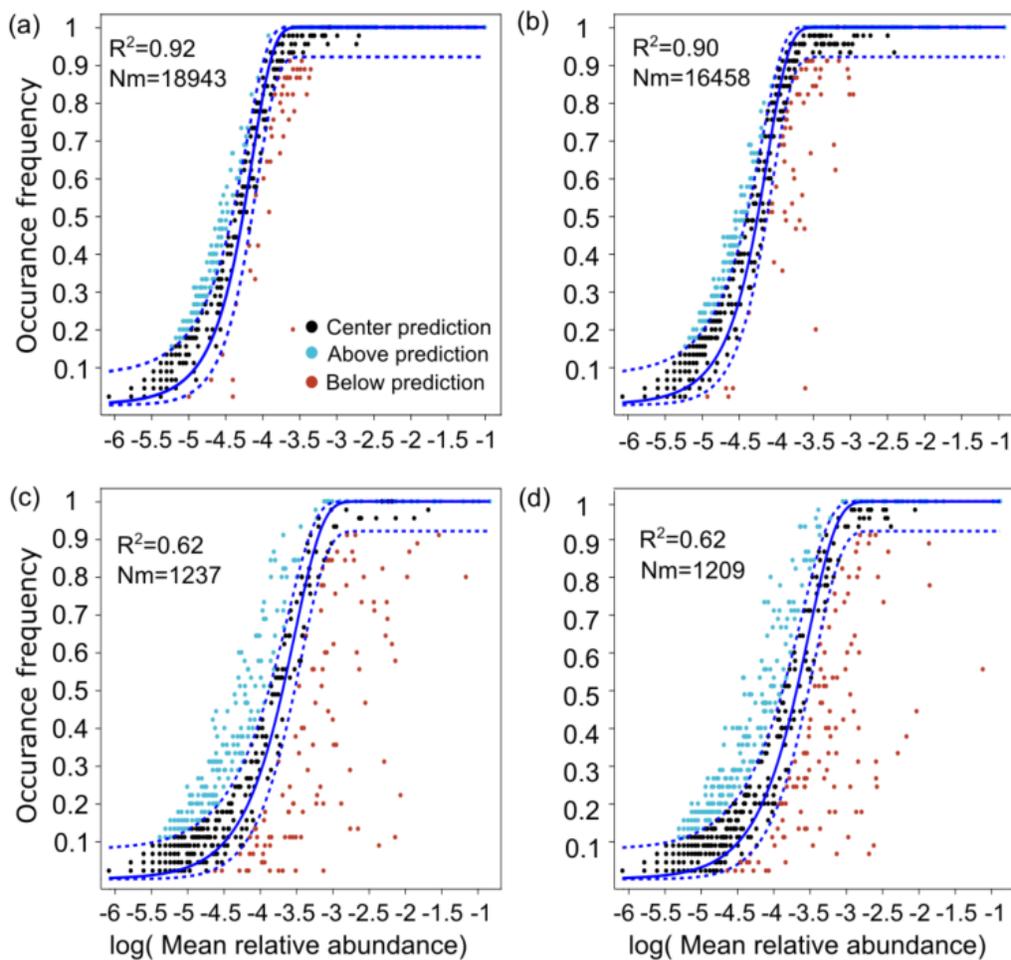
790 **Figure S5. Variables'  $R^2$  in the redundancy analysis (RDA) in Figure 3 in the main text.**  $R^2$   
 791 indicates the proportion of variation of soil microbial communities explained by the variables,  
 792 respectively. Star next to a bar indicates that it is statistically significant ( $p < 0.05$ ).



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795 **Figure S6. Effects of bedrock on microbial community assembly processes.** (a–b) Fit of  
 796 the neutral community model to bacterial communities on granite and slate, respectively. (c–d)  
 797 Fit of the neutral community model to fungal communities on granite and slate, respectively.  
 798 Black dots indicate the best fit to the model ( $\pm$  95% confidence intervals);  $R^2$  values indicate  
 799 level of neutral community model prediction accuracy and  $Nm$  indicates estimated migration  
 800 volume of samples. “ $R^2$ ” represented overall goodness of fit of the NCM, with higher values  
 801 indicating that the improved fit was the result of stochastic processes; “ $N$ ” represented microbial  
 802 metacommunity size (number of OTUs); “ $m$ ” represented migration rate of microbes, with  
 803 smaller values indicating less diffusion limitation.



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